

1 **Impacts of anthropogenic change on biodiversity affect disease spillover risk**

2
3 **Caroline K. Glidden*¹, Nicole Nova*¹, Morgan P. Kain^{1,3}, Katherine M. Lagerstrom¹, Eloise B.**
4 **Skinner^{1,2}, Lisa Mandle^{1,3,4}, Susanne H. Sokolow^{1,4,5}, Raina K. Plowright⁶, Rodolfo Dirzo^{1,4},**
5 **Giulio A. De Leo^{1,4,5}, Erin A. Mordecai¹**

6
7 ***Contributed equally**

8
9 **Affiliations:**

10 ¹Department of Biology, Stanford University, Stanford, CA, USA

11 ²Centre for Planetary Health and Food Security Griffith University, Gold Coast, QLD 4222, Australia

12 ³Natural Capital Project, Stanford University, Stanford, CA, USA

13 ⁴Woods Institute for the Environment, Stanford University, Stanford, CA, USA

14 ⁵Hopkins Marine Station, Stanford University, Pacific Grove, CA, USA

15 ⁶Department of Microbiology and Immunology, Montana State University, Bozeman, MT, USA

16
17 **Abstract**

18 The integration of biodiversity conservation and public health has gained significant traction, leading to
19 new efforts to identify win–win solutions for sustainable development and health. At the forefront of these
20 efforts is pin-pointing ways that biodiversity conservation can reduce risk of zoonotic spillover, especially
21 given the consequences of pandemics and epidemics of wild animal origin. However, there is currently an
22 incomplete understanding of the mechanisms by which biodiversity change influences the spillover process,
23 limiting the application of integrated strategies aimed at achieving positive outcomes for both conservation
24 and disease management. One limitation has been a narrow focus on the relationship between infectious
25 disease and species richness only, thus missing other relevant dimensions of biodiversity. Here, we review
26 the literature, considering a broad scope of biodiversity definitions, to identify cases where zoonotic
27 pathogen spillover is mechanistically linked to changes in biodiversity. Extending biodiversity to include
28 other dimensions of it, such as functional diversity, landscape diversity, spatiotemporal diversity, and
29 interaction diversity, allows us to identify potential relationships between biodiversity change and zoonotic
30 spillover. By reframing the discussion of biodiversity and disease using mechanistic evidence while
31 encompassing multiple dimensions of biodiversity, we work toward general principles that can guide future
32 research and more effectively integrate the related goals of biodiversity conservation and spillover
33 prevention.

34
35
36
37
38
39
40
41
42
43
44

45 Introduction

46

47 The COVID-19 pandemic has brought the threat of zoonoses to the forefront, creating widespread demand
48 for managing ecological sources of disease spillover and emergence. Prior to this pandemic, increasing
49 recognition of the ties between healthy ecosystems and human health were beginning to garner support of
50 global conservation initiatives (e.g., the Bonn challenge) and spurred the United Nations' (UN) adoption of
51 sustainable development goals (the 2030 Agenda). Zoonotic spillover prevention is a biosecurity imperative
52 with a patent connection to the human–wildlife interface; thus, efforts are underway to identify win–win
53 solutions for biodiversity conservation and zoonotic disease management (see [1]). However, given the
54 incomplete understanding of the mechanisms linking biodiversity to infectious disease, a clear vision of
55 efficacy and pathways for win–win solutions are needed. Increased attention to, and resources for, zoonotic
56 disease prevention make it an opportune time to study the mechanisms connecting changes in biodiversity
57 with zoonotic disease spillover, and to identify (potentially synergistic) solutions for biodiversity
58 conservation and global health.

59

60 To date, the discussion around biodiversity and disease has been boxed into a contentious debate about the
61 existence and generality of a biodiversity–disease relationship: in particular, the extent to which
62 maintaining biodiversity protects against disease via a dilution effect, and the alternative possibility that
63 biodiversity can increase infectious disease transmission via an amplification effect (e.g., [2-9]). With a
64 few notable exceptions¹⁰⁻¹⁴, this debate has largely focused on correlations between host species richness
65 and reservoir host pathogen prevalence. However, the results of this research show that effects vary in
66 magnitude and direction, rendering this approach inactionable for public health interventions and context-
67 dependent. Here, we focus less on the narrow question of species richness and more on the broader
68 mechanistic relationships among a variety of components of biodiversity and the zoonotic spillover process,
69 followed by a review of general principles with applied relevance; finally, we present a number of
70 opportunities where ongoing conservation initiatives could consider these mechanisms further in order to
71 reduce disease spillover risks (Table 1, Figure 1).

72

73 Although biodiversity is often equated with species richness, biodiversity encompasses all forms of
74 variability among living organisms and the ecological complexes of which they are a part¹⁵. For example,
75 biodiversity includes the functional diversity of a community, habitat diversity of an ecosystem, and many
76 more interdependent characterizations¹⁵ (Box 1). At the same time, zoonotic spillover encompasses
77 multiple, interconnected processes in which pathogens circulating in one host population cause infections
78 in another^{16,17} (Figure 1). Zoonotic spillover is affected by many upstream ecological processes before a
79 pathogen actually spills into a human host, including reservoir host density, distribution, susceptibility, and
80 pathogen prevalence, infectiousness, survival, dissemination, and host-human contact. Once in the recipient
81 (human) host, a series of biological and epidemiological factors determine if onward transmission is
82 possible^{16,17,18,19} (Figure 1). To harmonize spillover prevention and biodiversity conservation, we need a
83 clear mechanistic understanding of how increases and decreases in multi-faceted aspects of biodiversity,
84 from cascading effects at population level to communities and ecosystems, influence all components of the
85 spillover process (Figure 1).

86

87 This review of the literature focuses on how infectious disease systems change with shifts in biodiversity,
88 highlighting case studies that suggest causal mechanisms (Table 1, Figure 1). We group case studies based

89 on the leading International Union for Conservation of Nature (IUCN)-classified threats to biodiversity.
90 While examples that mechanistically link environmental change to zoonotic spillover via at least one metric
91 of biodiversity change are scarce, our review identifies emerging generalities across disease systems and
92 anthropogenic disturbances. We then review ongoing sustainability initiatives that could incorporate
93 spillover prevention, emphasizing how reframing the discussion about biodiversity and disease may
94 facilitate win–win outcomes.

95

96 **Anthropogenic disturbance, biodiversity change, and disease spillover**

97

98 *Agricultural expansion & intensification*

99

100 As of 2019, agricultural expansion and intensification were the leading causes of biodiversity loss¹⁵.
101 Agricultural development fragments and clears previously extensive ecosystems, creating edge habitats that
102 increase human encroachment with wildlife²⁰, homogenizing landscapes to reduce availability of natural
103 resources for wildlife, and releasing pesticides, fertilizers, and antimicrobial compounds into the
104 environment. All of these factors contribute to population declines or even local extinctions of several
105 species²¹⁻²³ and may influence the dynamics of infectious diseases with an important environmental
106 component in their transmission cycle²⁴.

107

108 Clearing intact ecosystems for agriculture, development of plantations, and other land modifications
109 (including urbanization), drive the loss of large- and medium-bodied animals (i.e., defaunation) while
110 supporting the persistence or growth of populations of small-bodied animals²⁵⁻²⁸. Recent research has made
111 it clear that loss of functional diversity due to non-random patterns of defaunation has significant effects
112 on zoonotic spillover risk. Increase in disease risk spillover due to change in functional diversity of animal
113 communities may occur either through expansion or invasion of opportunistic zoonotic hosts that thrive in
114 human modified landscapes or through the cascading effect of human induced extirpation of predators and
115 competitors of zoonotic species, as described here after.

116

117 Small-bodied mammals are common pathogen reservoirs, with the rodent and bat orders containing the
118 highest number of known zoonotic hosts²⁹⁻³². Certain taxa of small-bodied animals are likely to predominate
119 in human-modified landscapes due to traits that make them adaptable to living in proximity to humans^{33,34}.
120 These traits, namely diet and habitat generalism with fast-paced life history, high population density and
121 promiscuity with human settlements, are positively correlated with zoonotic reservoir status^{30,35,36}. On a
122 global scale, the richness and abundance of zoonotic hosts (especially birds, bats, and rodents) positively
123 correlates with degree of human land management^{35,37}. Local studies in Kenya, Tanzania, and Madagascar
124 found that this change in functional diversity increases zoonotic disease risk: rodent communities in
125 croplands had a higher proportion of competent zoonotic reservoir hosts and higher prevalence of zoonotic
126 pathogens than in unmanaged areas^{14,38,39}.

127

128 Shifts in functional diversity of ecological communities may be driven also by the loss of interactions
129 among large, medium, and small bodied animals. In savanna ecosystems in Central Kenya, extirpation of
130 large herbivores resulted in changes in the plant community and competitive release of small herbivores,
131 leading to the increase in abundance of competent rodent hosts (*Saccostomus mearnsi*) and prevalence of
132 *Bartonella* and vectors^{40,41}. Predators of reservoir hosts and vectors might also exert a crucial role in

133 modulating the risk of disease spillover for humans¹⁰. In Senegal, the construction of the Diama dam in
134 1986 to prevent saltwater intrusion and support agriculture intensification blocked the migration of native
135 predators (the giant river prawn, *Macrobrachium vollenhoveni*) that consume snail vectors and free-living
136 *Schistosoma* spp., resulting in increased transmission of vector-borne parasites to humans⁴²— these findings
137 have been linked to construction of large dams and subsequent increases in schistosomiasis transmission
138 throughout Africa⁴³. In terrestrial, zoonotic disease systems, the presence of leopards may decrease risk of
139 rabies transmission to humans by preying on stray dogs in Mumbai, India⁴⁴. Further, predator loss might
140 trigger significantly more complex trophic cascades. The loss of wolves in the Northeastern USA was
141 followed up by an increase in coyotes which in turn led to a dramatic reduction of small-mammal predators
142 that control the abundance of rodents competent hosts for Lyme disease¹¹.

143
144 In general, land conversion can homogenize habitats, alter species distributions⁴⁵⁻⁴⁷ and change contact
145 patterns between wildlife and humans. The resulting decrease in diversity and availability of productive
146 and undisturbed habitat can shift the distribution of reservoir species to aggregate at high densities near
147 humans, increasing interspecific contact rates and contacts between humans, previously unencountered
148 mammals, and vectors, thereby increasing potential for transmission. For example, *Plasmodium knowlesi*
149 malaria is expanding in Malaysia and across Southeast Asia, partially due to forest loss and agricultural
150 land conversion⁴⁸⁻⁵³. Loss of diverse and undisturbed habitat has driven the primary *P. knowlesi* reservoirs,
151 long-tailed macaques (*Macaca fascicularis*) and pig-tailed macaques (*Macaca nemestrina*), to occupy small
152 forest fragments within or next to agricultural areas where they overlap with anthropophilic mosquito
153 vectors and people^{54,55,53}. This shift in distribution not only increases the density of reservoirs, potentially
154 increasing transmission among reservoir hosts, but also increases potential for macaque–vector–human
155 transmission⁵³. High profile zoonotic pathogens, such as Ebola virus, similarly spillover in forest
156 fragments^{56,57}, calling into question the effect of landscape configuration and diversity on zoonotic spillover
157 risk.

158
159 Lastly, loss of diverse habitat and skew in functional diversity towards favoring reservoir hosts may
160 concurrently increase the risk of antimicrobial resistant (AMR) zoonotic spillover. Runoff from antibiotic-
161 fed livestock forms wastewater lagoons where diverse bacteria mix and face strong selective pressures to
162 develop and share, via horizontal gene transfer (HGT), genes conferring resistances to those antibiotics^{58,59}.
163 This also occurs in aquacultural waters⁶⁰, wastewater from antibiotic-treated crops⁶¹, and effluent from
164 wastewater treatment plants (WWTP)⁶². Wildlife that contact polluted waters or soils can pick up these
165 AMR bacteria and transport them to both neighboring and distant croplands or livestock operations where
166 they can spill over to people⁶³⁻⁶⁷. Global rates of AMR are on the rise, driven by the misuse of antibiotics
167 in clinical settings as well as the areas described above, with an estimated 700,000 deaths worldwide caused
168 by AMR bacterial infections⁶⁸. Our current understanding of the propensity for diverse wild animal species
169 to harbor AMR bacteria is limited, in part due to little sampling efforts to-date⁶⁹. However, initial research
170 shows that animal populations highly exposed or adaptable to human modified habitats have higher
171 prevalence of AMR bacteria than animals with little to no contact with humans⁷⁰, suggesting they may be
172 more competent reservoir hosts of these potentially infectious agents. Smith et al. [70] found that prevalence
173 of AMR bacteria in agricultural areas decreased as the amount of native habitat increased, possibly due to
174 reducing contact rates of birds with livestock runoff. As such, diverse habitats may reduce the likelihood
175 of birds becoming inoculated with and transmitting AMR bacteria. Diverse habitats may jointly decrease
176 AMR risk by protecting croplands from livestock wastewater runoff⁷¹. The effect of biodiversity change on

177 AMR spillover is severely understudied but, given the threat of AMR bacteria to global public health⁷²,
178 warrants significant attention^{69,70}.

179
180 *Climate change*

181
182 Species may respond to climate change through plasticity⁷³, rapid adaptive evolution⁷⁴, and altitudinal and
183 latitudinal range shifts to the edge of their geographic range⁷⁵⁻⁷⁷. Alternatively, species may undergo global
184 extinction or, more frequently, local population extinctions^{78,79}. Together these responses can drive
185 biodiversity change in complex, nonlinear, and interdependent ways. In the absence of emigration, species
186 ranges may shrink, which has been observed primarily in polar and montane species (reviewed [80]), or
187 species may face extirpation or extinction (e.g., [81, 82]). Furthermore, the velocity of rising temperatures
188 differs among regions of the world, affecting species and populations differently⁸³. Here, we focus on case
189 studies of range shifts in response to rapid anthropogenic climate change, as it is the most immediately
190 observable impact of climate change on wildlife hosts harboring zoonotic pathogens^{84,85}. Plastic, adaptive,
191 and local declines or extirpation responses are currently well researched⁸⁶⁻⁸⁸, with the amphibian decline
192 being perhaps the most emblematic case⁸⁹, but rarely in the context of pathogen spillover.

193
194 The abundance of specialist species, and thus functional diversity, may decline with range shifts, especially
195 at high latitudes, although taxonomic diversity of some systems may increase with range shifts⁹⁰⁻⁹². This is
196 largely attributed to generalists outnumbering specialists in systems impacted by global change, as
197 generalists are able to thrive in a variety of ecological conditions, including human modified landscapes,
198 while specialists need specific resources and/or habitats to survive. At the same time, correlative analyses
199 suggest that zoonotic reservoirs are more likely to be generalist species^{35,36,93}, as they are more likely to
200 live in closer proximity to people and contact a wider range of other host species. Further, climate-induced
201 forest habitat loss may lead to an increase in abundance of extreme generalists with zoonotic reservoir
202 potential, as in the case of the highly adaptable deer mice harboring Sin Nombre virus⁹⁴.

203
204 The Alaskan Arctic is currently exhibiting climate-induced shifts in host species, namely an increase in the
205 abundance of zoonotic hosts more likely to contact humans. Before contemporary climate change, the
206 ranges of two carnivores and rabies reservoir hosts, red and Arctic foxes, were separated⁹⁵; however, with
207 climate change the home range of the generalist red fox has expanded northward, encroaching on the
208 territory of the comparatively habitat specialist Arctic fox⁹⁶. Arctic fox numbers were already in decline
209 due to other effects of climate change, such as the loss of sea ice and tundra habitat as well as loss of
210 lemming prey, but red foxes are expediting this decline through intraguild predation and competition for
211 resources⁹⁷⁻⁹⁹. As Arctic fox populations are replaced by red fox populations, the red fox will become the
212 primary reservoir for rabies spillover. This shift in the reservoir community will likely increase epizootic
213 peaks of rabies as immigrant red foxes interact more with resident Arctic foxes, increasing both the
214 transmission rate and the overall density of susceptible individuals¹⁰⁰. Further, because the larger-bodied
215 red fox displays more aggressive behavior than the Arctic fox⁹⁵, and because it is more amenable to adapt
216 to human-dominated landscapes, contact rates between wild rabies reservoirs and dogs or humans might
217 increase, thus increasing rabies spillover risk.

218
219 Comparable to effects of direct land-conversion (e.g., agricultural expansion), climate change may reduce
220 habitat diversity and subsequently increase the likelihood of cross-species transmission through increased

221 habitat overlap and taxonomic diversity in confined areas¹⁰¹. For instance, the melting of sea ice alters,
222 disrupts, or even prevents migration patterns of animals such as wild caribou¹⁰², increasing the chance of
223 intermingling among caribou and other wild or domestic ungulates. Thus, people who rely on caribou and/or
224 other livestock might be at higher risk of brucellosis spillover under a warming climate in temperate
225 regions¹⁰³. In water-stressed parts of Africa, extreme droughts can similarly force many animals that
226 previously had little to no contact with one another (such as humans, wildlife, and livestock) to congregate
227 at common water sources^{104,105}. When water sources are more abundant, wildlife occupy heterogeneous
228 landscapes characterized by different types of surface water (e.g., rivers, seasonal pans, lakes)¹⁰⁶⁻¹⁰⁹. In
229 drought conditions, animals are forced to use the same watering hole where increased traffic and fecal loads
230 reduce water quality. In Chobe National Park, Botswana, these patterns and processes are associated with
231 increased loads of *E. coli*, the leading cause of diarrheal outbreaks¹⁰⁵. Following drought events, heavy
232 seasonal rainfall and flooding events mobilize pathogen-containing feces, subsequently leading to human
233 diarrheal outbreaks in neighboring communities¹¹⁰.

234

235 *Invasive species*

236

237 Invasive species (i.e., organisms introduced outside their natural range negatively impact native
238 biodiversity, ecosystem services, or human-wellbeing¹¹¹) present a significant threat to ecosystems.
239 Through processes such as predation, competition, or environmental modification, invasive species can
240 drastically decrease the biodiversity of an ecosystem; an estimated thirty species of invasive predators alone
241 are responsible for at least 58% of all bird, mammal, and reptile extinctions globally¹¹². Invasive species
242 can indirectly impact infectious disease by altering the structure and composition of the native community
243 in ways that either increase or decrease pathogen transmission.

244

245 Altering a native community to increase zoonotic spillover risk has been empirically demonstrated for the
246 Everglade virus, a mosquito-borne zoonotic virus. The introduction of the Burmese python (*Python*
247 *bivittatus*) to the Florida Everglades has led to large-scale declines in functional and taxonomic mammal
248 diversity due to precipitous loss of large and small-bodied mammals^{113,114}. With loss of mosquito food
249 sources due to python predation (on deer, racoons, and opossums), mosquito vectors of Everglades virus
250 fed dramatically more on the primary reservoir host of the virus, the hispid cotton rat (*Sigmodon hispidus*),
251 potentially increasing the risk of virus exposure to humans¹¹⁴.

252

253 In contrast, introduction of invasive species can reduce transmission of infectious disease from vectors to
254 people through predation on various vector life stages (larvivorous fish on malaria vectors¹¹⁵; crayfish on
255 schistosome intermediate hosts¹¹⁶). However, despite crayfish lowering the risk of schistosomiasis by
256 voraciously consuming snail intermediate hosts and free-living parasites, invasive crayfish compromised
257 other dimensions of human health by consuming rice and degrading canal banks with their burrows¹¹⁷.
258 Consequently, in scenarios where invasive species reduce disease risk there can still be a tension between
259 biodiversity impacts of invasive species and their specific ecological roles in infectious disease dynamics.

260

261 Invasive species may affect infectious disease dynamics by acting as vectors or reservoir hosts^{29, 39, 118-120},
262 sharing pathogens with native species¹²¹⁻¹²³ or providing resources for reservoirs and/or vectors^{124,120}. In
263 these cases, biodiversity conservation via invasive species control may simultaneously reduce zoonotic
264 spillover risk (see [120]). The same processes that drive species introductions, including global trade and

265 travel, may also drive disease emergence, suggesting that win–win solutions might be possible, though
266 potentially technically and politically challenging¹²⁵.

267

268

269 *Wildlife hunting, trade, and consumption*

270

271 One in five vertebrate species are impacted by trade¹²⁶, with some wildlife facing population declines and/or
272 species extinction due, mainly or in part, to the impacts of legal and illegal wildlife trade (e.g., tigers,
273 rhinos, elephants, sharks, and pangolins)^{127,128}. The illegal wildlife trade is estimated to be the world’s
274 second largest underground businesses (hypothesized to be a \$5–20 billion-dollar industry) after
275 narcotics¹²⁹. While it is still trumped by the \$300 billion-dollar legal wildlife trade industry, the majority of
276 legal wildlife trade (78%) is still made up of wild caught animals as opposed to those reared in captivity¹³⁰.

277

278 Epidemiological and genetic analyses have linked wildlife hunting, trade, and consumption to spillover and
279 spread of many high-profile zoonotic pathogens: rabies virus, Crimean-Congo hemorrhagic fever virus, the
280 plague-causing bacteria *Y. pestis*, monkeypox virus, coronaviruses, HIV, Marburg and Ebola viruses^{127, 128,}
281 ¹³⁰⁻¹³³. However, in order to stop or mitigate the spillover process, we need to have a better understanding
282 of the mechanisms linking the wildlife trade to the steps leading to spillover (Figure 1).

283

284 The wildlife trade highlights how anthropogenic pressures can increase spillover risk via a direct increase
285 in both taxonomic diversity and the number of interactions across taxa (i.e., interaction diversity) (defined
286 in Box 1) on very small spatial scales. Throughout the supply chain, the wildlife trade brings together high
287 densities of species that typically would not contact each other in natural habitats. These unique
288 assemblages and interactions can promote cross-species transmission, increasing the likelihood that a
289 pathogen may be transmitted to amplifying hosts (i.e., hosts in which a pathogen can rapidly replicate to
290 high concentrations) and/or humans (see refs in following sentence). Trade may also impact the spillover
291 process by promoting pathogen shedding from animals because of unsanitary conditions during, and stress
292 from, transportation and market¹³⁴⁻¹⁴¹. For example, the ancestor to SARS-CoV-1 is suspected to have been
293 transmitted from horseshoe bats (most likely *Rhinolophus sinicus*) to palm civets, two species that do not
294 interact in wild settings. However, palm civets served as amplifying hosts or as intermediate hosts within
295 wildlife markets, bringing the virus in closer proximity to humans¹⁴²⁻¹⁴⁴. Seroprevalence and virological
296 testing surveys of civets on farms versus those brought to markets in Guangdong, China suggest that palm
297 civets were exposed to the virus at the end of the supply chain¹⁴³⁻¹⁴⁵. A study performed in Vietnam showed
298 that coronavirus detection in field rats caught or reared for human consumption more than doubled when
299 testing field rats sold in markets, and further increased by 10-fold when testing field rats sold or served in
300 restaurants, compared with rats in the wild¹⁴⁰. Thus, the wildlife trade creates opportunities for increased
301 transmission among multiple wild animal species and puts humans in closer proximity to stressed and
302 infected wildlife, fueling the potential for spillover of pathogens.

303

304 The wildlife trade for human consumption can take on various forms, which in turn can interact to amplify
305 the effects of overharvesting of wild animals. For example, the wild meat trade in Ghana, which has driven
306 population declines of certain mammalian species in the last few decades, correlates with local declines in
307 fish supply, probably due to overfishing off the coast^{146, 147}. Conceivably, during periods when the demand
308 for wild meat is high, hunters and people involved with the butchering and preparation of the meat expose

309 themselves to a higher risk of disease spillover from bites, scratches, and otherwise coming into contact
310 with bodily fluids of animals serving as reservoirs for many pathogens. In the Congo basin and other regions
311 of pathogen emergence, wild meat serves as a protein source primarily in poor households, making the
312 banning of wild meat a controversial topic¹⁴⁸, though genetic and epidemiological evidence suggests that it
313 has contributed to the rise of emerging diseases and recent outbreaks via spillover from wildlife to humans
314 of pathogens like HIV, Ebola, Marburg, and monkeypox viruses^{135, 149, 150}. Indeed, phylogenetic analyses of
315 HIV suggest that approximately 10 spillover events occurred over the past century before HIV caused a
316 pandemic¹⁵¹. In Cameroon, simian foamy viruses regularly spill over and infect wild meat hunters, but no
317 human-to-human transmission has yet been established¹³¹.

318

319 Overexploitation of wild meat and other anthropogenic pressures have also been correlated with a decrease
320 in the proportion of large-bodied mammals and an increase in the proportion of small-bodied mammals
321 brought to market^{152, 153}. As a result, preliminary research suggests that overharvesting of wildlife may
322 influence the types of wild animals hunters and consumers are contacting; however, mechanistic links
323 between change in composition of wildlife markets and zoonotic disease risk have not yet been established.

324

325 *Urbanization*

326

327 With almost 70% of the world's population expected to live in urban areas by 2050¹⁵⁴, the transition from
328 rural to urban land-use is a complex and dynamic process likely to power the greatest landscape
329 transformations of the 21st century. Urban areas are characterized by high human density and an almost
330 completely built environment.

331

332 Land conversion can reduce temporal diversity of food sources, which can cause nomadic and migrating
333 species to forgo migration in favor of occupying the same habitat year-round. In some cases, formation of
334 resident populations may shift reservoir host dynamics to alter zoonotic spillover risk, particularly when
335 loss of seasonal, high-quality natural resources is paired with provisioning of non-seasonal, subpar food¹⁵⁵.
336 For example, loss of optimal winter resources, at least in part due to habitat loss, drives reservoir hosts
337 (*Pteropus* spp.) of Hendra virus from large nomadic groups that track seasonally abundant nectar sources
338 into small resident groups feeding on permanent, suboptimal food within and around cities^{156,16, 157}. Food
339 stress may promote viral shedding; simultaneously, the redistribution of reservoir hosts into smaller yet
340 more abundant colonies in human dominated systems increases the likelihood of the virus spilling into
341 amplifying hosts (horses) and humans¹⁵⁸. Reducing temporal diversity of resources or prohibiting access to
342 high quality seasonal food sources may also promote spillover with respect to agricultural conversion. For
343 example, agricultural conversion has limited the availability of high-quality winter resources for elk. Large
344 populations are now supported by lower-quality supplemental feeding, which reduces migration and
345 promotes high density aggregations, thereby increasing the spread of *Brucella abortus* among reservoir
346 hosts and potentially spillover to livestock¹⁵⁹⁻¹⁶².

347

348 Further, the rural to urban transition diversifies local economies from dependence on local agriculture to
349 trade of goods, services, and ideas with more distant places¹⁶³. Through trade with rural areas, urbanization
350 interacts with other threats to biodiversity, such as introduction of pathogens through the wildlife trade and
351 introduction of invasive species, to drive changes in zoonotic spillover¹⁶⁴. Drastic reduction of non-human
352 adapted animals in completely converted land (i.e., cities) may reduce the frequency of spillover of novel

353 zoonotic pathogens²⁰. At the same time, interactions between urbanization and other anthropogenic
354 disturbances creates circumstances for pathogen introduction, especially if pathogens can be sustained via
355 human–human transmission. For example, urban centers serve as hubs for long-distance shipping, with
356 urban wildlife markets often containing higher densities and diversity of wildlife. Thus, urban wildlife
357 markets create unique assemblages of species subsequently increasing the likelihood of novel cross-species
358 transmission¹⁶⁵. Then, in the rare case where the biology of the pathogen allows frequent human-to-human
359 transmission (e.g., high infectivity to humans, asymptomatic transmission, aerosol transmission¹⁸), the large
360 and dense human population found in cities can facilitate rapid pathogen spread, resulting in the largest
361 epidemics²⁰ or even pandemics. Spread of novel zoonotic pathogens may be mitigated by increased health
362 and subsequent reduced susceptibility in planned urban areas (i.e., built from a blueprint, with the city's
363 future in mind)¹⁶⁶. However, the opposite may be true in unplanned urban areas (i.e., built ad-hoc, without
364 centralized infrastructure and equitable distribution of resources) where human health might be
365 compromised by increased pollution, lack of affordable healthcare, and limited access to healthy food and
366 clean water^{167, 164}.

367
368

369 *Emerging generalities of the effects of anthropogenic disturbance on disease spillover through biodiversity*
370 *change*

371

372 While mechanistic research linking changes in biodiversity to zoonotic spillover risk is limited due to the
373 expense and logistical challenges of elucidating these relationships, by considering a broader range of
374 changes in biodiversity than just species richness and composition, we collect enough literature to propose
375 four general concepts that are potentially operationalizable in ongoing and developing biodiversity
376 initiatives. These generalities may motivate further integration of biodiversity and zoonotic pathogen
377 spillover research, potentially opening more avenues of funding as well as incorporation of multi-
378 disciplinary methods for collecting and analyzing data. Echoing Halsey [8], we distinguish between
379 *generality*, that which is mostly considered true, and *universality*, that which is considered true in all
380 possible contexts. These generalities may be more or less applicable for different ecosystems and disease
381 threats.

382

383 (1) Loss of spatially and temporally diverse habitat alters the distributions of reservoirs leading to increased
384 overlap more with other vertebrate hosts, vectors, and humans (see *Agricultural expansion &*
385 *intensification*, *Climate change*, and *Urbanization* sections). This generality suggests an opportunity:
386 preserving and restoring large, contiguous, and heterogeneous habitats could minimize encroachment of
387 humans and wildlife in their respective habitats, thus decreasing intraspecific contact rates and ultimately
388 transmission among reservoirs. Intact and diverse contiguous landscapes may also promote landscape
389 immunity, defined as ecological conditions that maintain and strengthen the immune system of wild animals
390 to reduce pathogen susceptibility and shedding, particularly for potential reservoir species including bats
391 and rodents¹⁶⁸. Further, targeted habitat conservation and restoration could encourage previous migration
392 patterns by re-creating or maintaining temporal diversity of high-quality food sources, such as nectar
393 resources for bats^{16, 120}. However, in some cases, resource provisioning—through invasive species, crops,
394 and even waste disposal practices—may reduce migration even when phenologically diverse habitats are
395 available^{169, 170}. More research differentiating the impact of habitat restoration versus limiting human
396 provisions (e.g., through clearing of invasive plants or better waste disposal management) is needed.

397 Overall, studying the mechanistic effect of spatiotemporal habitat diversity on each spillover process
398 (Figure 1) should lead to new insights that can guide evidence-based policy for both conserving natural
399 ecosystems and reducing spillover risk.

400

401 (2) Loss of large consumers and predators (change in functional diversity) can result in increased abundance
402 of animals with fast growth rates and relatively small ranges, such as rodent reservoirs and vectors (see
403 *Agricultural expansion & intensification*). Thus, habitat conservation, preservation, and restoration could
404 simultaneously reduce the density and improve the health of wildlife as well as supporting populations of
405 large predators and herbivores^{153, 171, 172}. In turn, predators and large consumers may be important in
406 ecotones between intact and anthropogenic landscapes, where they can regulate populations of small-bodied
407 reservoirs that thrive in human modified areas. More research is needed to understand the impacts large
408 herbivores and predators have on zoonotic disease regulation, especially within and around these ecotones.
409 If more evidence supports a beneficial effect of conserving predators and large herbivores for reducing
410 spillover risk, conservation of predators and large consumers may offer another promising win-win
411 situation.

412

413 (3) Human modification further affects functional diversity by changing habitats and shifting communities
414 toward dominance by species that are resilient to anthropogenic disturbance or thrive in human dominated
415 landscapes. As described previously, these animals are more likely to be zoonotic reservoirs (see
416 *Agricultural expansion & intensification, Climate change*). Change in functional diversity towards
417 synanthropic species has been observed across taxonomic groups of vertebrates (e.g., rodents and birds: see
418 *Agricultural expansion & intensification*; carnivores: see *Climate change*). Similar effects have been
419 observed for disease vectors in which generalists thrive in urban areas and have high capacity to transmit
420 pathogens to humans^{43, 173, 174}. Integrative approaches, such as direct management of invasive rodents and
421 vectors or indirect management through preserving intact habitat and mitigating impacts of climate change
422 to reduce range shifts of reservoirs and vectors, are likely necessary¹²⁰.

423

424 (4) Commercial wildlife trade, introduction of invasive species, and transportation of livestock and
425 companion animals are activities that increase interaction diversity, introducing more opportunities for
426 cross-species transmission among different species and increasing the chance of new pathogens emerging
427 that may have zoonotic spillover potential (see *Invasive species, Wildlife trade, and Urbanization*). Overall,
428 regulations and initiatives that reduce diversity of novel interspecific interactions should be adjusted to
429 incorporate spillover prevention.

430

431 Despite these generalities, cases of disease spillover driven by biodiversity change (in turn driven by
432 anthropogenic disturbances) are highly context-dependent. Further, biodiversity conservation is not a
433 panacea for zoonotic spillover prevention, and many systems are too complex or understudied to prescribe
434 clear links between biodiversity change and spillover risk. For example, highly diverse multi-host, multi-
435 vector systems such as West Nile Virus (WNV), Ross River virus^{175,176}, leishmaniasis¹⁷⁷, and Chagas
436 disease¹⁷⁸, require more studies to document ecological drivers of reservoir and vector abundances and
437 capacities to transmit disease. Further, reservoir host species that contribute most to transmission may be
438 variable along geographic and land-use gradients¹⁷⁹⁻¹⁸⁴. Thus, win-win solutions for alleviating zoonotic
439 disease burden and conserving biodiversity should be evaluated based on specific knowledge of the
440 ecological and social system contexts¹.

441

442 **Biodiversity and sustainability initiatives can combat disease spillover**

443

444 Our synthesis provides new research directions and avenues for management at the intersection of
445 biodiversity conservation and global health and begins to focus on a mechanistic understanding of the links
446 between environmental change, biodiversity, and infectious disease spillover. Solutions grounded in
447 ecological understanding may also require attention to the social context: promotion of long-term funding
448 for community conservation and sustainable livelihoods programs; legislation from global to local entities
449 for spillover prevention; and local awareness of, and investments in, monitoring the impacts of human
450 activities on biodiversity and spillover in understudied systems. Certain international initiatives are already
451 working towards sustainable solutions for promoting both public health (including preventing disease
452 burden) and conservation, such as the UN Sustainable Development Goals¹⁸⁵, Intergovernmental Science-
453 Policy Platform on Biodiversity and Ecosystem Services (IPBES) Nature's Contributions to People¹⁸⁶,
454 IUCN Global Standards for Nature-Based Solutions¹⁸⁷, Bridge Collaborative¹⁸⁸, Pan American and World
455 Health Organizations (PAHO/WHO) Climate Change and Health¹⁸⁹, Global Health Security Agenda¹⁹⁰, and
456 the collaboration among Food and Agriculture Organization (FAO), World Organisation for Animal Health
457 (OIE), and WHO (FAO-OIE-WHO Collaboration)¹⁹¹.

458

459 Here, we describe some ongoing biodiversity and sustainability initiatives that could incorporate spillover
460 prevention (e.g., to avoid unintended harms from biodiversity conservation or to broaden the benefits of
461 biodiversity conservation), emphasizing how reframing the discussion about biodiversity and disease to
462 focus on ecological understanding of mechanisms creates opportunities for synergistic solutions. Though
463 biodiversity may not causally affect spillover in all cases, sustainable development efforts can still jointly
464 benefit conservation and human health¹⁹², and at a minimum, avoid unintended harms.

465

466 *The Bonn Challenge*

467

468 The Bonn Challenge was launched by the Government of Germany and IUCN in 2011. Its goal was to
469 obtain pledges for 150 million hectares of degraded and deforested landscapes globally on which to begin
470 restoration by 2020 (which was successfully reached in 2017) and 350 million hectares by 2030¹⁹³. This
471 Challenge recognizes the benefits conserving and restoring degraded or deforested landscapes makes to
472 human health, wellbeing, and livelihood. However, the Challenge does not address any potential effects of
473 infectious disease burden or spillover directly. Where pledges lead to successful landscape-scale restoration
474 of wildlife habitat, especially for large-bodied predators and consumers, the effort could potentially help
475 reduce spillover risk driven by biodiversity change (e.g., increase in rodent abundance due to competitor
476 and predator release) related to agriculture and deforestation (see *Agricultural expansion & intensification*).

477

478 *Convention on Biological Diversity*

479

480 The United Nations' Convention on Biological Diversity (CBD) proposes a list of goals between 2020 and
481 2050 for nature-based solutions (NbS) for benefitting planetary health and human health^{194,195}, defined as:
482 "actions to protect, sustainably manage and restore natural or modified ecosystems that address societal
483 challenges effectively and adaptively, simultaneously providing human well-being and biodiversity
484 benefits." This clearly encompasses win-win solutions for human health and biodiversity conservation.

485 However, the CBD handbooks, including in 2020, do not mention actionable next steps for implementing
486 such proposed interventions or even what they are¹⁹⁵. Further, human health or well-being is not clearly
487 defined, and there is no mention of infectious diseases. However, these goals do address mitigation of
488 biodiversity loss and the anthropogenic pressures mentioned above. Thus, there should be potential for
489 efforts led by CBD to also target spillover prevention, which merits further investigation.

490

491 *Convention on International Trade in Endangered Species*

492

493 In 1973, 21 countries signed a global agreement called the Convention on International Trade in
494 Endangered Species (CITES) of Wild Fauna and Flora to regulate the international wildlife trade, and ban
495 trade of endangered species. Today, the CITES agreement is being implemented by 182 countries and the
496 European Union to regulate the trade of more than 35,000 species¹⁹⁶. CITES supports surveillance efforts
497 to track species under threat in the international wildlife trade, and works together with law enforcement
498 from wildlife organizations, national parks, customs, and the police force to control illegal wildlife trade
499 activity. However, the stated mission of CITES does not include the prevention of spillover. Only a few
500 CITES country members use strict veterinary import border controls for animals. Further, there are no
501 global regulations on pathogen screening related to the international wildlife trade. Conceivably, CITES
502 could adopt a pathogen screening regulation scheme to be implemented by all of its country members to
503 prevent the global spread of emerging diseases (see *Wildlife hunting, trade, and consumption*), especially
504 since emerging diseases may also hurt endangered wild populations¹³⁶. Another possible synergistic effort
505 that could prevent zoonotic spillover and preserve wildlife and biodiversity would be to reduce legal (and
506 prevent illegal) commercial overfishing. CITES recognizes the necessity to control bycatch of threatened
507 species, such as the critically endangered vaquita porpoise in Chinese and Mexican waters¹⁹⁶. Fisheries
508 collapse driven by overfishing has been associated with elevated bushmeat trade (see *Wildlife hunting,*
509 *trade, and consumption*) in other parts of the world. Further investigations are warranted to unravel
510 connections between various forms of wildlife exploitations and their downstream effects on spillover.

511

512 *Thirty By Thirty Resolution to Save Nature*

513

514 This initiative is part of a global campaign called the Campaign for Nature, spearheaded by the Wyss
515 Campaign for Nature, National Geographic Society, and over 100 conservation organizations¹⁹⁷. In 2020,
516 the Natural Resources Defense Council (NRDC) proposed a 30x30 “commitment to protect nature and life
517 on Earth” urging the US federal government to conserve at least 30% of US lands and at least 30% of ocean
518 regions by the year 2030^{198, 199}. Currently, the US protects only 12% of lands and 26% of its surrounding
519 ocean areas²⁰⁰. The Biden administration announced its support for this goal in 2021²⁰⁰; however, the
520 document written by the Agriculture, Commerce, and Interior Departments does not recognize the
521 additional health benefits of reduced spillover risk via the proposed conservation efforts²⁰¹. For example,
522 the document proposes conservation of wildlife habitat and corridors for safe passage of wildlife between
523 intact habitats. This would aid protection and maintenance of preserves and areas of intact habitat, as well
524 as restoration and conservation of natural predators and large consumers. Conceivably, such biodiversity
525 outcomes could help reduce spillover risk of zoonotic disease in the US where large-bodied mammals keep
526 reservoir populations in check (e.g., rodents, see *Agricultural expansion & intensification*) or where
527 corridors help migrations of large-bodied herbivores (e.g., caribou and brucellosis, see *Climate Change*).

528

529 *Payments for Ecosystem Services Program in Costa Rica*

530

531 The idea behind the Payments for Ecosystem Services Program (PES) effort is to let those who benefit from
532 ecosystem services (e.g., biodiversity conservation, watershed services, carbon sequestration, and
533 landscape beauty) compensate stewards of these services. For example, landowners keeping forests intact
534 should be compensated for the services their forests provide to users downstream (e.g., carbon
535 sequestration, clean air, clean rivers)²⁰². PES schemes have been applied globally²⁰³, with many successful
536 programs implemented on the local, regional, and countrywide level in South America²⁰³.

537

538 The government of Costa Rica was the first to implement a nation-wide PES program in 1997, which is
539 funded through taxation (e.g., fossil fuel taxes), tariffs, and contracts with the industry (e.g., private hydro-
540 electric producers), and loans from the World Bank and Global Environmental Facility Grant²⁰⁴. In addition
541 to landowner compensation for their land's ecosystem services in Costa Rica, Indigenous territories were
542 included as priority areas for conservation²⁰⁵.

543

544 PES programs do not explicitly include spillover prevention. However, spillover prevention could be
545 embedded in forest conservation and restoration aimed to improve biodiversity conservation and other
546 recognized ecosystem services or be introduced as its own ecosystem service. PES schemes that conserve
547 contiguous and diverse forests could potentially benefit spillover prevention in multiple ways. For example,
548 density of small-bodied mammal reservoir hosts may decrease (see *Agricultural expansion &*
549 *intensification*), and intact forests serve as carbon sinks thereby mitigating downstream effects of climate
550 change (see *Climate change*). For programs centered around preserving and restoring forest, providing
551 service suppliers with alternative income may further reduce contact with zoonotic disease reservoirs and
552 vectors. However, service providers are not necessarily adequately compensated in Costa Rica²⁰⁶, so
553 schemes should be improved to offer improved, sustainable livelihoods. Conversely, PES schemes that
554 promote sustainable forest management and agroforestry could theoretically increase spillover risk for some
555 zoonotic diseases and should be planned with care. Increased disease risk may result from provisioning
556 animals with high abundance and preferable resources, creating habitat for vectors, and from humans more
557 frequently entering forested areas²⁰⁷. In these cases, zoonotic spillover may be better managed through other
558 ecological interventions (e.g., barriers that limit human exposure to pathogens shed from reservoirs)²⁰⁸ or
559 vector management.

560

561 *Project Finance for Permanence*

562

563 Project Finance for Permanence (PFP) is another model that includes restoring and conserving contiguous
564 intact ecosystems²⁰⁹. These projects are funded by a diverse set of donors, including local and global
565 governments, private foundations, and private sector companies, and brokered and managed by NGOs (e.g.,
566 WWF) and government agencies. Examples of successful programs include the Amazon Region Protected
567 Areas (ARPA), the Great Bear Rainforest Project, and Forever Costa Rica^{210, 210}. Each project aims to
568 improve the abundance and management of intact ecosystems, although at different capacities. The ARPA
569 intends to create, consolidate, and maintain a 60 million hectare network of protected areas in the Brazilian
570 Amazon (3x larger than all US National Parks combined). The Great Bear Rainforest Project supports 21
571 million acres (~8.5 million hectares) while promoting sustainable development among the area's First
572 Nation People. The Forever Costa Rica Project has worked to maintain funding of and improve

573 management of existing protected areas. Although not a specific program objective, ARPA has likely
574 reduced cases of malaria transmission in the Inner Amazon by slowing the rate of deforestation²¹¹. This
575 example highlights the potential joint benefits of the PFP model for conservation and public health.
576 Spillover prevention is not yet incorporated in PFP programs, although they could be extended to zoonotic
577 spillover prevention via similar mechanisms to PES programs.

578
579

580 **Conclusions and future directions**

581

582 We identified mechanistic evidence in the literature that in certain systems anthropogenically-driven
583 biodiversity change increases zoonotic spillover risk. Several common themes emerged. First, the loss of
584 intact habitat increases overlap between reservoirs and other vertebrate hosts, vectors, and humans. Second,
585 loss of large-bodied consumers and predators (defaunation) can result in increased abundance of rodent
586 reservoirs. Third, human-modified landscapes change the functional diversity of species assemblages,
587 increasing the proportion of species that are able to adapt to, or even thrive in, these new environments.
588 Because these species tend to be better hosts for zoonotic pathogens, these landscape modifications increase
589 human exposure to zoonotic pathogens. Fourth, other forms of anthropogenic disturbance generated by
590 agriculture and trade of domestic animals and wildlife lead to the introduction of invasive species and
591 increase interaction diversity, facilitating opportunities for cross-species transmission and thus the potential
592 for emergence of novel pathogens with zoonotic spillover potential.

593

594 Certain disease systems are either understudied or too complex to elucidate the effects of biodiversity
595 change on spillover risk. In addition, some components of the spillover process (Figure 1) are better-studied
596 than others in the context of the impacts of biodiversity change. Based on our review, the effects of
597 biodiversity change on wildlife host susceptibility, pathogen shedding, and pathogen prevalence in the
598 reservoir for example (three important steps of spillover) are understudied compared to human pathogen
599 exposure (Figure 1). This may arise because wildlife host susceptibility, pathogen shedding and prevalence
600 are difficult to observe²¹². Another possibility could be lack of appreciation for the upstream mechanisms
601 in the human-environmental system by which ecological changes can affect animal health and well-being
602 which can, in turn, affect spillover. Wildlife hosts are not just containers and mixing vessels for zoonotic
603 diseases; when exposed to stress from anthropogenic activities. Hosts may experience suppressed immune
604 systems, rendering them more susceptible to opportunistic infections, more pathogen shedding, and altered
605 behavior that increases their exposure to pathogens^{213, 214}. Thus, studies of biodiversity conservation and
606 human health are needed to interrogate mechanisms by which environmental change (stressors) affecting
607 wild animal health may lead to changes in the process of disease spillover to people and domestic animals.
608 There is a need for spatially and temporally replicated field studies incorporating biodiversity change,
609 pathogen dynamics, and wildlife host immunology^{168, 214} in addition to human health outcomes.

610

611 Anthropogenic drivers of biodiversity change interact in complex ways, including synergies, direct and
612 indirect effects and complex feedbacks. The combined impacts among many different anthropogenic
613 disturbances may exacerbate the effects of biodiversity change on spillover risk. For example, defaunation
614 is caused by many anthropogenic drivers (i.e., wildlife trade, climate change, agricultural intensification,
615 and invasive species), which may additively, multiplicatively, or synergistically drive defaunation^{26, 215},
616 with downstream effects on ecosystem function and zoonotic spillover risk²¹⁶. For instance, tropical land

617 use change leading to deforestation and fragmentation negatively impacts the medium and large mammals
618 of rainforests in Southeast Mexico, while promoting the proliferation of small mammals (rodents, many of
619 them zoonotic reservoirs). Such fragmentation in turn facilitates additional hunting and poaching, which
620 exacerbates defaunation and, in turn, increases exposure of humans to these disease hosts²¹⁶. Very likely
621 this synergy land use change-overexploitation will be further complicated by the changes in climate regime
622 this topographically complex tropical area will experience. Moreover, feedback exists between spillover
623 outcomes and anthropogenic activity, ultimately altering the outcomes for biodiversity conservation efforts.
624 For example, spillback from humans to wildlife or other animals (reverse zoonosis (e.g.,²¹⁷)) or fear of
625 secondary spillovers may alter disease dynamics and potentially suppress anthropogenic activities that drive
626 biodiversity loss²¹¹. Thus, research involving the combined effects of multiple drivers of biodiversity
627 change and its relationship with spillover is an agenda that warrants further efforts.

628
629 The world today is undergoing rapid anthropogenic change with detrimental effects on biodiversity and the
630 health of organisms, including humans. Efforts are underway to combat the impact of anthropogenic
631 disturbances on biodiversity. However, since biodiversity change may affect zoonotic disease spillover
632 through multiple mechanisms, we urge biodiversity conservation efforts to also incorporate actions to
633 prevent spillover into existing programs. The issue of spillover is not only an issue for public health, but
634 also for conservation of threatened wildlife. Here, we argue that reframing discussions of biodiversity and
635 disease around a more inclusive definition of biodiversity, and considering the context of each of the
636 complex social-ecological systems in which the spillover process occurs (Figure 1, Box 1) are essential to
637 highlight mechanistic links between biodiversity and zoonotic spillover. This approach sheds light on how
638 to develop win-win interventions that prevent zoonotic spillover while protecting biodiversity.

639

640

641 **Acknowledgements**

642

643 We thank Gretchen Daily, Elizabeth Hadly, and members of the Mordecai Lab (Alexander Becker, Devin
644 Kirk, Marissa Childs, Lisa Couper, Johanna Farner, Mallory Harris, Isabel Dewel, Gowri Vadmal) for
645 thoughtful feedback on early drafts of the manuscript. CKG, EAM and LM were supported by the
646 National Science Foundation (NSF; DEB-2011147, with the Fogarty International Center). NN was
647 supported by the Philanthropic Educational Organization (PEO) Scholar Award from the International
648 Chapter of the PEO Sisterhood, and the Stanford Data Science Scholars program. MPK and LM were
649 supported by the Natural Capital Project. SHS and GADL were partially supported by the NSF (DEB-
650 2011179) and Belmont Forum of Climate Environment and Health and NSF initiative (ICER-2024383).
651 RKP was funded by the DARPA PREEMPT program (Cooperative Agreement: D18AC00031), the NSF
652 (DEB-1716698), and the USDA National Institute of Food and Agriculture (Hatch project 1015891).
653 EAM was also supported by the NSF (DEB-1518681), the National Institute of General Medical Sciences
654 (R35GM133439), the Terman Award, the Stanford King Center for Global Development, the Stanford
655 Woods Institute for the Environment, and the Stanford Center for Innovation in Global Health.

656

657 **References cited:**

- 658 1. Hopkins, S.R., Sokolow, S.H., Buck, J.C., De Leo, G.A., Jones, I.J., Kwong, L.H., LeBoa, C., Lund,
659 A.J., MacDonald, A.J., Nova, N., et al. (2020). How to identify win–win interventions that benefit
660 human health and conservation. *Nat. Sustain.* 4, 298–304.
- 661 2. Randolph, S.E., and Dobson, A.D.M. (2012). Pangloss revisited: a critique of the dilution effect and the
662 biodiversity–buffers–disease paradigm. *Parasitology* 139, 847–863.
- 663 3. Lafferty, K.D., and Wood, C.L. (2013). It’s a myth that protection against disease is a strong and
664 general service of biodiversity conservation: Response to Ostfeld and Keesing. *Trends Ecol. Evol.* 28,
665 503–504.
- 666 4. Ostfeld, R.S., and Keesing, F. (2013). Straw men don’t get Lyme disease: response to Wood and
667 Lafferty. *Trends Ecol. Evol.* 28, 502–503.
- 668 5. Salkeld, D.J., Padgett, K.A., and Jones, J.H. (2013). A meta-analysis suggesting that the relationship
669 between biodiversity and risk of zoonotic pathogen transmission is idiosyncratic. *Ecol. Lett.* 16, 679–
670 686.
- 671 6. Wood, C.L., Lafferty, K.D., DeLeo, G., Young, H.S., Hudson, P.J., and Kuris, A.M. (2014). Does
672 biodiversity protect humans against infectious disease? *Ecology* 95, 817–832.
- 673 7. Civitello, D.J., Cohen, J., Fatima, H., Halstead, N.T., Liriano, J., McMahon, T.A., Ortega, C.N., Sauer,
674 E.L., Sehgal, T., Young, S., et al. (2015). Biodiversity inhibits parasites: Broad evidence for the
675 dilution effect. *Proc. Natl. Acad. Sci.* 112, 8667–8671.
- 676 8. Halsey, S. (2019). Defuse the dilution effect debate. *Nat. Ecol. Evol.* 3, 145–146.
- 677 9. Rohr, J.R., Civitello, D.J., Halliday, F.W., Hudson, P.J., Lafferty, K.D., Wood, C.L., and Mordecai,
678 E.A. (2020). Towards common ground in the biodiversity–disease debate. *Nat. Ecol. Evol.* 4, 24–33.
- 679 10. Ostfeld, R.S., and Holt, R.D. (2004). Are predators good for your health? Evaluating evidence for top-
680 down regulation of zoonotic disease reservoirs. *Front. Ecol. Environ.* 2, 13–20.
- 681 11. Levi, T., Kilpatrick, A.M., Mangel, M., and Wilmers, C.C. (2012). Deer, predators, and the
682 emergence of Lyme disease. *Proc. Natl. Acad. Sci.* 109, 10942–10947.
- 683 12. Johnson, P.T.J., de Roode, J.C., and Fenton, A. (2015). Why infectious disease research needs
684 community ecology. *Science* 349, 1259504.
- 685 13. Levi, T., Massey, A.L., Holt, R.D., Keesing, F., Ostfeld, R.S., and Peres, C.A. (2016). Does
686 biodiversity protect humans against infectious disease? *Comment. Ecology* 97, 536–542.
- 687 14. Young, H.S., McCauley, D.J., Dirzo, R., Nunn, C.L., Campana, M.G., Agwanda, B., Otarola-Castillo,
688 E.R., Castillo, E.R., Pringle, R.M., Veblen, K.E., et al. (2017). Interacting effects of land use and
689 climate on rodent-borne pathogens in central Kenya. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.* 372.
- 690 15. Diaz, S., Settele, E., Brondizio, E., Ngo, H., Gueze, J., Agard, A., Balvanera, K., Brauman, S.,
691 Buthchart, K., Chan, L., et al. (2019). IPBES (2019): Summary for policymakers of the global

- 692 assessment report on biodiversity and ecosystem services of the Intergovernmental Science-Policy
693 Platform on Biodiversity and Ecosystem Services. (IPBES secretariat).
- 694 16. Plowright, R.K., Eby, P., Hudson, P.J., Smith, I.L., Westcott, D., Bryden, W.L., Middleton, D., Reid,
695 P.A., McFarlane, R.A., Martin, G., et al. (2015). Ecological dynamics of emerging bat virus spillover.
696 *Proc. R. Soc. B Biol. Sci.* 282.
- 697 17. Plowright, R.K., Parrish, C.R., McCallum, H., Hudson, P.J., Ko, A.I., Graham, A.L., and Lloyd-
698 Smith, J.O. (2017). Pathways to zoonotic spillover. *Nat. Rev. Microbiol.* 15, 502–510.
- 699 18. Wasik, B.R., de Wit, E., Munster, V., Lloyd-Smith, J.O., Martinez-Sobrido, L., and Parrish, C.R.
700 (2019). Onward transmission of viruses: how do viruses emerge to cause epidemics after spillover?
701 *Philos. Trans. R. Soc. Lond. B. Biol. Sci.* 374, 20190017.
- 702 19. Lloyd-Smith, J.O., George, D., Pepin, K.M., Pitzer, V.E., Pulliam, J.R.C., Dobson, A.P., Hudson, P.J.,
703 and Grenfell, B.T. (2009). Epidemic dynamics at the human-animal interface. *Science* 326, 1362–
704 1367.
- 705 20. Faust, C.L., McCallum, H.I., Bloomfield, L.S.P., Gottdenker, N.L., Gillespie, T.R., Torney, C.J.,
706 Dobson, A.P., and Plowright, R.K. (2018). Pathogen spillover during land conversion. *Ecol. Lett.* 21,
707 471–483.
- 708 21. Krebs, J.R., Wilson, J.D., Bradbury, R.B., and Siriwardena, G.M. (1999). The second Silent Spring?
709 *Nature* 400, 611–612.
- 710 22. Laurance, W.F., Sayer, J., and Cassman, K.G. (2014). Agricultural expansion and its impacts on
711 tropical nature. *Trends Ecol. Evol.* 29, 107–116.
- 712 23. Estrada, A., Garber, P.A., Rylands, A.B., Roos, C., Fernandez-Duque, E., Fiore, A.D., Nekaris, K.A.-
713 I., Nijman, V., Heymann, E.W., Lambert, J.E., et al. (2017). Impending extinction crisis of the world’s
714 primates: Why primates matter. *Sci. Adv.* 3, e1600946.
- 715 24. Rohr, J.R., Barrett, C.B., Civitello, D.J., Craft, M.E., Delius, B., DeLeo, G.A., Hudson, P.J., Jouanard,
716 N., Nguyen, K.H., Ostfeld, R.S., et al. (2019). Emerging human infectious diseases and the links to
717 global food production. *Nat. Sustain.* 2, 445–456.
- 718 25. Estes, J.A., Terborgh, J., Brashares, J.S., Power, M.E., Berger, J., Bond, W.J., Carpenter, S.R.,
719 Essington, T.E., Holt, R.D., Jackson, J.B.C., et al. (2011). Trophic Downgrading of Planet Earth.
720 *Science* 333, 301–306.
- 721 26. Dirzo, R., Young, H.S., Galetti, M., Ceballos, G., Isaac, N.J.B., and Collen, B. (2014). Defaunation in
722 the Anthropocene. *Science* 345, 401–406.
- 723 27. Graham, S.I., Kinnaird, M.F., O’Brien, T.G., Vågen, T.-G., Winowiecki, L.A., Young, T.P., and
724 Young, H.S. (2019). Effects of land-use change on community diversity and composition are highly
725 variable among functional groups. *Ecol. Appl.* 29, e01973.
- 726 28. Gutiérrez-Granados, G., and Dirzo, R. (2021). Logging drives contrasting animal body-size effects on
727 tropical forest mammal communities. *For. Ecol. Manag.* 481, 118700.

- 728 29. de Faria, M.T., Calderwood, M.S., Athanazio, D.A., McBride, A.J.A., Hartskeerl, R.A., Pereira,
729 M.M., Ko, A.I., and Reis, M.G. (2008). Carriage of *Leptospira interrogans* among domestic rats from
730 an urban setting highly endemic for leptospirosis in Brazil. *Acta Trop.* 108, 1–5.
- 731 30. Han, B.A., Schmidt, J.P., Bowden, S.E., and Drake, J.M. (2015). Rodent reservoirs of future zoonotic
732 diseases. *Proc. Natl. Acad. Sci.* 112, 7039–7044.
- 733 31. Luis, A.D., O’Shea, T.J., Hayman, D.T.S., Wood, J.L.N., Cunningham, A.A., Gilbert, A.T., Mills,
734 J.N., and Webb, C.T. (2015). Network analysis of host-virus communities in bats and rodents reveals
735 determinants of cross-species transmission. *Ecol. Lett.* 18, 1153–1162.
- 736 32. Han, B.A., Kramer, A.M., and Drake, J.M. (2016). Global Patterns of Zoonotic Disease in Mammals.
737 *Trends Parasitol.* 32, 565–577.
- 738 33. Purvis, A., Gittleman, J.L., Cowlishaw, G., and Mace, G.M. (2000). Predicting extinction risk in
739 declining species. *Proc. R. Soc. Lond. B Biol. Sci.* 267, 1947–1952.
- 740 34. Cusack, J. (2011). Characterising small mammal responses to tropical forest loss and degradation in
741 Northern Borneo using capture-mark-recapture methods. *Imp. Coll. Lond.*
- 742 35. Gibb, R., Redding, D.W., Chin, K.Q., Donnelly, C.A., Blackburn, T.M., Newbold, T., and Jones, K.E.
743 (2020). Zoonotic host diversity increases in human-dominated ecosystems. *Nature* 584, 398–402.
- 744 36. Johnson, C.K., Hitchens, P.L., Pandit, P.S., Rushmore, J., Evans, T.S., Young, C.C.W., and Doyle,
745 M.M. (2020). Global shifts in mammalian population trends reveal key predictors of virus spillover
746 risk. *Proc. Biol. Sci.* 287, 20192736.
- 747 37. Mendoza, H., Rubio, A.V., García-Peña, G.E., Suzán, G., and Simonetti, J.A. (2019). Does land-use
748 change increase the abundance of zoonotic reservoirs? Rodents say yes. *Eur. J. Wildl. Res.* 66, 6.
- 749 38. McCauley, D.J., Salkeld, D.J., Young, H.S., Makundi, R., Dirzo, R., Eckerlin, R.P., Lambin, E.F.,
750 Gaffikin, L., Barry, M., and Helgen, K.M. (2015). Effects of land use on plague (*Yersinia pestis*)
751 activity in rodents in Tanzania. *Am. J. Trop. Med. Hyg.* 92, 776–783.
- 752 39. Herrera, J.P., Wickenkamp, N.R., Turpin, M., Baudino, F., Tortosa, P., Goodman, S.M., Soarimalala,
753 V., Ranaivoson, T.N., and Nunn, C.L. (2020). Effects of land use, habitat characteristics, and small
754 mammal community composition on *Leptospira* prevalence in northeast Madagascar. *PLoS Negl.*
755 *Trop. Dis.* 14, e0008946.
- 756 40. Young, H.S., Dirzo, R., Helgen, K.M., McCauley, D.J., Billeter, S.A., Kosoy, M.Y., Osikowicz, L.M.,
757 Salkeld, D.J., Young, T.P., and Dittmar, K. (2014). Declines in large wildlife increase landscape-level
758 prevalence of rodent-borne disease in Africa. *Proc. Natl. Acad. Sci.* 111, 7036–7041.
- 759 41. Titcomb, G., Allan, B.F., Ainsworth, T., Henson, L., Hedlund, T., Pringle, R.M., Palmer, T.M.,
760 Njoroge, L., Campana, M.G., Fleischer, R.C., et al. (2017). Interacting effects of wildlife loss and
761 climate on ticks and tick-borne disease. *Proc. R. Soc. B Biol. Sci.* 284, 20170475.
- 762 42. Sokolow, S.H., Huttinger, E., Jouanard, N., Hsieh, M.H., Lafferty, K.D., Kuris, A.M., Riveau, G.,
763 Senghor, S., Thiam, C., N’Diaye, A., et al. (2015). Reduced transmission of human schistosomiasis
764 after restoration of a native river prawn that preys on the snail intermediate host. *Proc. Natl. Acad. Sci.*
765 112, 9650–9655.

- 766 43. Sokolow, S.H., Jones, I.J., Jocque, M., La, D., Cords, O., Knight, A., Lund, A., Wood, C.L., Lafferty,
767 K.D., Hoover, C.M., et al. (2017). Nearly 400 million people are at higher risk of schistosomiasis
768 because dams block the migration of snail-eating river prawns. *Philos. Trans. R. Soc. B Biol. Sci.* *372*,
769 20160127.
- 770 44. Braczkowski, A.R., O'Bryan, C.J., Stringer, M.J., Watson, J.E., Possingham, H.P., and Beyer, H.L.
771 (2018). Leopards provide public health benefits in Mumbai, India. *Front. Ecol. Environ.* *16*, 176–182.
- 772 45. Benton, T.G., Vickery, J.A., and Wilson, J.D. (2003). Farmland biodiversity: is habitat heterogeneity
773 the key? *Trends Ecol. Evol.* *18*, 182–188.
- 774 46. Chace, J.F., and Walsh, J.J. (2006). Urban effects on native avifauna: a review. *Landsc. Urban Plan.*
775 *74*, 46–69.
- 776 47. Durán, A.P., Green, J.M.H., West, C.D., Visconti, P., Burgess, N.D., Virah-Sawmy, M., and
777 Balmford, A. (2020). A practical approach to measuring the biodiversity impacts of land conversion.
778 *Methods Ecol. Evol.* *11*, 910–921.
- 779 48. Chang, M.S., Hii, J., Buttner, P., and Mansoor, F. (1997). Changes in abundance and behaviour of
780 vector mosquitoes induced by land use during the development of an oil palm plantation in Sarawak.
781 *Trans. R. Soc. Trop. Med. Hyg.* *91*, 382–386.
- 782 49. Cox-Singh, J., Davis, T.M.E., Lee, K.-S., Shamsul, S.S.G., Matusop, A., Ratnam, S., Rahman, H.A.,
783 Conway, D.J., and Singh, B. (2008). *Plasmodium knowlesi* malaria in humans is widely distributed
784 and potentially life-threatening. *Clin. Infect. Dis. Off. Publ. Infect. Dis. Soc. Am.* *46*, 165–171.
- 785 50. Brock, P.M., Fornace, K.M., Parmiter, M., Cox, J., Drakeley, C.J., Ferguson, H.M., and Kao, R.R.
786 (2016). *Plasmodium knowlesi* transmission: integrating quantitative approaches from epidemiology
787 and ecology to understand malaria as a zoonosis. *Parasitology* *143*, 389–400.
- 788 51. Fornace, K.M., Abidin, T.R., Alexander, N., Brock, P., Grigg, M.J., Murphy, A., William, T., Menon,
789 J., Drakeley, C.J., and Cox, J. (2016). Association between Landscape Factors and Spatial Patterns of
790 *Plasmodium knowlesi* Infections in Sabah, Malaysia. *Emerg. Infect. Dis.* *22*, 201–208.
- 791 52. Manin, B.O., Ferguson, H.M., Vythilingam, I., Fornace, K., William, T., Torr, S.J., Drakeley, C., and
792 Chua, T.H. (2016). Investigating the Contribution of Peri-domestic Transmission to Risk of Zoonotic
793 Malaria Infection in Humans. *PLoS Negl. Trop. Dis.* *10*, e0005064.
- 794 53. Davidson, G., Chua, T.H., Cook, A., Speldewinde, P., and Weinstein, P. (2019). Defining the
795 ecological and evolutionary drivers of *Plasmodium knowlesi* transmission within a multi-scale
796 framework. *Malar. J.* *18*, 66.
- 797 54. Brant, H.L., Ewers, R.M., Vythilingam, I., Drakeley, C., Benedick, S., and Mumford, J.D. (2016).
798 Vertical stratification of adult mosquitoes (Diptera: Culicidae) within a tropical rainforest in Sabah,
799 Malaysia. *Malar. J.* *15*, 370.
- 800 55. Vythilingam, I., Wong, M.L., and Wan-Yussof, W.S. (2018). Current status of *Plasmodium knowlesi*
801 vectors: a public health concern? *Parasitology* *145*, 32–40.
- 802 56. Rulli, M.C., Santini, M., Hayman, D.T.S., and D'Odorico, P. (2017). The nexus between forest
803 fragmentation in Africa and Ebola virus disease outbreaks. *Sci. Rep.* *7*, 41613.

- 804 57. Olivero, J., Fa, J.E., Farfán, M.Á., Márquez, A.L., Real, R., Juste, F.J., Leendertz, S.A., and Nasi, R.
805 (2020). Human activities link fruit bat presence to Ebola virus disease outbreaks. *Mammal Rev.* 50, 1–
806 10.
- 807 58. Tymensen, L., Zaheer, R., Cook, S.R., Amoako, K.K., Goji, N., Read, R., Booker, C.W., Hannon,
808 S.J., Neumann, N., and McAllister, T.A. (2018). Clonal expansion of environmentally-adapted
809 *Escherichia coli* contributes to propagation of antibiotic resistance genes in beef cattle feedlots. *Sci.*
810 *Total Environ.* 637–638, 657–664.
- 811 59. Zhang, M., Liu, Y.-S., Zhao, J.-L., Liu, W.-R., He, L.-Y., Zhang, J.-N., Chen, J., He, L.-K., Zhang,
812 Q.-Q., and Ying, G.-G. (2018). Occurrence, fate and mass loadings of antibiotics in two swine
813 wastewater treatment systems. *Sci. Total Environ.* 639, 1421–1431.
- 814 60. Lulijwa, R., Rupia, E.J., and Alfaro, A.C. (2020). Antibiotic use in aquaculture, policies and
815 regulation, health and environmental risks: a review of the top 15 major producers. *Rev. Aquac.* 12,
816 640–663.
- 817 61. Taylor, P., and Reeder, R. (2020). Antibiotic use on crops in low and middle-income countries based
818 on recommendations made by agricultural advisors. *CABI Agric. Biosci.* 1, 1.
- 819 62. Sabri, N.A., Schmitt, H., Van der Zaan, B., Gerritsen, H.W., Zuidema, T., Rijnaarts, H.H.M., and
820 Langenhoff, A.A.M. (2020). Prevalence of antibiotics and antibiotic resistance genes in a wastewater
821 effluent-receiving river in the Netherlands. *J. Environ. Chem. Eng.* 8, 102245.
- 822 63. Bonnedahl, J., and Järhult, J.D. (2014). Antibiotic resistance in wild birds. *Ups. J. Med. Sci.* 119,
823 113–116.
- 824 64. Rivadeneira, P., Hilson, C., Justice-Allen, A., and Jay-Russell, M. (2016). Pathogen Risks Related to
825 the Movement of Birds Frequenting Livestock and Fresh Produce Growing Areas in the Southwestern
826 U.S. *Proc. Vertebr. Pest Conf.* 27.
- 827 65. Borges, C.A., Cardozo, M.V., Beraldo, L.G., Oliveira, E.S., Maluta, R.P., Barboza, K.B., Werther, K.,
828 and Ávila, F.A. (2017). Wild birds and urban pigeons as reservoirs for diarrheagenic *Escherichia coli*
829 with zoonotic potential. *J. Microbiol. Seoul Korea* 55, 344–348.
- 830 66. Ishibashi, S., Sumiyama, D., Kanazawa, T., and Murata, K. (2019). Prevalence of antimicrobial-
831 resistant *Escherichia coli* in endangered Okinawa rail (*Gallirallus okinawae*) inhabiting areas around a
832 livestock farm. *Vet. Med. Sci.* 5, 563–568.
- 833 67. Navarro-Gonzalez, N., Castillo-Contreras, R., Casas-Díaz, E., Morellet, N., Concepción Porrero, M.,
834 Molina-Vacas, G., Torres, R.T., Fonseca, C., Mentaberre, G., Domínguez, L., et al. (2018). Carriage of
835 antibiotic-resistant bacteria in urban versus rural wild boars. *Eur. J. Wildl. Res.* 64, 60.
- 836 68. O'Neill, J. (2014). Antimicrobial Resistance: Tackling a crisis for the health and wealth of nations.
- 837 69. Lagerstrom, K.M., and Hadly, E.A. (2021). The under-investigated wild side of *Escherichia coli*:
838 genetic diversity, pathogenicity and antimicrobial resistance in wild animals. *Proc. R. Soc. B Biol. Sci.*
839 288, 20210399.

- 840 70. Smith, O.M., Edworthy, A.B., Taylor, J.M., Jones, M.S., Tormanen, A.P., Kennedy, C.M., Fu, Z.,
841 Latimer, C.E., Cornell, K.A., Michelotti, L.A., et al. (2020). Agricultural intensification heightens food
842 safety risks posed by wild birds. *J. Appl. Ecol.* 57, 2246–2257.
- 843 71. Karp, D.S., Gennet, S., Kilonzo, C., Partyka, M., Chaumont, N., Atwill, E.R., and Kremen, C. (2015).
844 Comanaging fresh produce for nature conservation and food safety. *Proc. Natl. Acad. Sci.* 112, 11126–
845 11131.
- 846 72. Dhingra, S., Rahman, N.A.A., Peile, E., Rahman, M., Sartelli, M., Hassali, M.A., Islam, T., Islam, S.,
847 and Haque, M. (2020). Microbial Resistance Movements: An Overview of Global Public Health
848 Threats Posed by Antimicrobial Resistance, and How Best to Counter. *Front. Public Health* 8.
- 849 73. Bonamour, S., Chevin, L.-M., Charmantier, A., and Teplitsky, C. (2019). Phenotypic plasticity in
850 response to climate change: the importance of cue variation. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.*
851 374, 20180178.
- 852 74. Catullo, R.A., Llewelyn, J., Phillips, B.L., and Moritz, C.C. (2019). The Potential for Rapid Evolution
853 under Anthropogenic Climate Change. *Curr. Biol.* 29, R996–R1007.
- 854 75. Parmesan, C., and Yohe, G. (2003). A globally coherent fingerprint of climate change impacts across
855 natural systems. *Nature* 421, 37–42.
- 856 76. Chen, I.-C., Hill, J.K., Ohlemüller, R., Roy, D.B., and Thomas, C.D. (2011). Rapid Range Shifts of
857 Species Associated with High Levels of Climate Warming. *Science* 333, 1024–1026.
- 858 77. Nogués-Bravo, D., Rodríguez-Sánchez, F., Orsini, L., de Boer, E., Jansson, R., Morlon, H., Fordham,
859 D.A., and Jackson, S.T. (2018). Cracking the Code of Biodiversity Responses to Past Climate Change.
860 *Trends Ecol. Evol.* 33, 765–776.
- 861 78. Ceballos, G., Ehrlich, P.R., and Dirzo, R. (2017). Biological annihilation via the ongoing sixth mass
862 extinction signaled by vertebrate population losses and declines. *Proc. Natl. Acad. Sci.* 114, E6089–
863 E6096.
- 864 79. Román-Palacios, C., and Wiens, J.J. (2020). Recent responses to climate change reveal the drivers of
865 species extinction and survival. *Proc. Natl. Acad. Sci. U. S. A.* 117, 4211–4217.
- 866 80. Bellard, C., Bertelsmeier, C., Leadley, P., Thuiller, W., and Courchamp, F. (2012). Impacts of climate
867 change on the future of biodiversity. *Ecol. Lett.* 15, 365–377.
- 868 81. Hetem, R.S., Fuller, A., Maloney, S.K., and Mitchell, D. (2014). Responses of large mammals to
869 climate change. *Temperature* 1, 115–127.
- 870 82. Waller, N., Gynther, I., Freeman, A., Lavery, T.H., and Leung, L. (2017). The Bramble Cay melomys
871 *Melomys rubicola* (Rodentia: Muridae): a first mammalian extinction caused by human-induced
872 climate change? *Wildl. Res.* 44, 9–21.
- 873 83. Loarie, S.R., Duffy, P.B., Hamilton, H., Asner, G.P., Field, C.B., and Ackerly, D.D. (2009). The
874 velocity of climate change. *Nature* 462, 1052–1055.
- 875 84. Waits, A., Emelyanova, A., Oksanen, A., Abass, K., and Rautio, A. (2018). Human infectious
876 diseases and the changing climate in the Arctic. *Environ. Int.* 121, 703–713.

- 877 85. Carlson, C.J., Albery, G.F., Merow, C., Trisos, C.H., Zipfel, C.M., Eskew, E.A., Olival, K.J., Ross,
878 N., and Bansal, S. (2020). Climate change will drive novel cross-species viral transmission (*Ecology*).
- 879 86. Couper, L., Farner, J., Caldwell, J., Childs, M., Harris, M., Kirk, D., Nova, N., Shocket, M., Skinner,
880 E., Uricchio, L., et al. (2020). How will mosquitoes adapt to climate change? (Preprints).
- 881 87. Parmesan, C. (2006). Ecological and Evolutionary Responses to Recent Climate Change. *Annu. Rev.*
882 *Ecol. Evol. Syst.* 37, 637–669.
- 883 88. Réale, D., McAdam, A.G., Boutin, S., and Berteaux, D. (2003). Genetic and plastic responses of a
884 northern mammal to climate change. *Proc. R. Soc. B Biol. Sci.* 270, 591–596.
- 885 89. Alan Pounds, J., Bustamante, M.R., Coloma, L.A., Consuegra, J.A., Fogden, M.P.L., Foster, P.N., La
886 Marca, E., Masters, K.L., Merino-Viteri, A., Puschendorf, R., et al. (2006). Widespread amphibian
887 extinctions from epidemic disease driven by global warming. *Nature* 439, 161–167.
- 888 90. Davey, C.M., Chamberlain, D.E., Newson, S.E., Noble, D.G., and Johnston, A. (2012). Rise of the
889 generalists: evidence for climate driven homogenization in avian communities. *Glob. Ecol. Biogeogr.*
890 21, 568–578.
- 891 91. Hof, A.R., Jansson, R., and Nilsson, C. (2012). Future Climate Change Will Favour Non-Specialist
892 Mammals in the (Sub)Arctics. *PLOS ONE* 7, e52574.
- 893 92. Brustolin, M.C., Nagelkerken, I., Ferreira, C.M., Goldenberg, S.U., Ullah, H., and Fonseca, G. (2019).
894 Future ocean climate homogenizes communities across habitats through diversity loss and rise of
895 generalist species. *Glob. Change Biol.* 25, 3539–3548.
- 896 93. Walsh, M.G., and Hossain, S. (2020). Population structure and diet generalism define a preliminary
897 ecological profile of zoonotic virus hosts in the Western Ghats, India. *Epidemics* 33, 100416.
- 898 94. Lehmer, E.M., Korb, J., Bombaci, S., McLean, N., Ghachu, J., Hart, L., Kelly, A., Jara-Molinar, E.,
899 O'Brien, C., and Wright, K. (2012). The Interplay of Plant and Animal Disease in a Changing
900 Landscape: The Role of Sudden Aspen Decline in Moderating Sin Nombre Virus Prevalence in
901 Natural Deer Mouse Populations. *EcoHealth* 9, 205–216.
- 902 95. Hersteinsson, P., and MacDonald, D.W. (1992). Interspecific Competition and the Geographical
903 Distribution of Red and Arctic Foxes *Vulpes Vulpes* and *Alopex lagopus*. *Oikos* 64, 505–515.
- 904 96. Elmhagen, B., Tannerfeldt, M., Verucci, P., and Angerbjörn, A. (2000). The arctic fox (*Alopex*
905 *lagopus*): an opportunistic specialist. *J. Zool.* 251, 139–149.
- 906 97. Rodnikova, A., Ims, R.A., Sokolov, A., Skogstad, G., Sokolov, V., Shtro, V., and Fuglei, E. (2011).
907 Red fox takeover of arctic fox breeding den: an observation from Yamal Peninsula, Russia. *Polar Biol.*
908 34, 1609.
- 909 98. Berteaux, D., Thierry, A.-M., Alisauskas, R., Angerbjörn, A., Buchel, E., Doronina, L., Ehrich, D.,
910 Eide, N.E., Erlandsson, R., Flagstad, Ø., et al. (2017). Harmonizing circumpolar monitoring of Arctic
911 fox: benefits, opportunities, challenges and recommendations. *Polar Res.* 36.
- 912 99. Elmhagen, B., Berteaux, D., Burgess, R.M., Ehrich, D., Gallant, D., Henttonen, H., Ims, R.A.,
913 Killengreend, S.T., Niemimaa, J., Norén, K., et al. (2017). Homage to Hersteinsson and Macdonald:

- 914 climate warming and resource subsidies cause red fox range expansion and Arctic fox decline. *Polar*
915 *Res.* 36.
- 916 100. Simon, A., Tardy, O., Hurford, A., Lecomte, N., Bélanger, D., and Leighton, P. (2019). Dynamics
917 and persistence of rabies in the Arctic. *Polar Res.*
- 918 101. Plowright, R.K., Cross, P.C., Tabor, G.M., Almborg, E., and Hudson, P.J. (2021). Climate change
919 and infectious disease dynamics. In *New directions in conservation medicine: applied cases of*
920 *ecological health* (Oxford University Press), pp. 111–121.
- 921 102. Post, E., Bhatt, U.S., Bitz, C.M., Brodie, J.F., Fulton, T.L., Hebblewhite, M., Kerby, J., Kutz, S.J.,
922 Stirling, I., and Walker, D.A. (2013). Ecological Consequences of Sea-Ice Decline. *Science* 341, 519–
923 524.
- 924 103. Hueffer, K., Parkinson, A.J., Gerlach, R., and Berner, J. (2013). Zoonotic infections in Alaska:
925 disease prevalence, potential impact of climate change and recommended actions for earlier disease
926 detection, research, prevention and control. *Int. J. Circumpolar Health* 72.
- 927 104. Clifford, D., Kazwala, D.R., and Coppolillo, P. (2008). Evaluating and managing zoonotic disease
928 risk in rural Tanzania.
- 929 105. Fox, J.T., and Alexander, K.A. (2015). Spatiotemporal Variation and the Role of Wildlife in
930 Seasonal Water Quality Declines in the Chobe River, Botswana. *PLOS ONE* 10, e0139936.
- 931 106. Weir, J., and Davison, E. (1965). Daily Occurrence of African Game Animals at Water Water Holes
932 During Dry Weather. *Zool. Afr.* 1, 353–368.
- 933 107. de Beer, Y., and van Aarde, R.J. (2008). Do landscape heterogeneity and water distribution explain
934 aspects of elephant home range in southern Africa’s arid savannas? *J. Arid Environ.* 72, 2017–2025.
- 935 108. Edwards, S., Gange, A.C., and Wiesel, I. (2015). Spatiotemporal resource partitioning of water
936 sources by African carnivores on Namibian commercial farmlands. *J. Zool.* 297, 22–31.
- 937 109. Roug, A., Muse, E.A., Clifford, D.L., Larsen, R., Paul, G., Mathayo, D., Mpanduji, D., Mazet,
938 J.A.K., Kazwala, R., Kiwango, H., et al. (2020). Seasonal movements and habitat use of African
939 buffalo in Ruaha National Park, Tanzania. *BMC Ecol.* 20, 6.
- 940 110. Alexander, K.A., Heaney, A.K., and Shaman, J. (2018). Hydrometeorology and flood pulse
941 dynamics drive diarrheal disease outbreaks and increase vulnerability to climate change in surface-
942 water-dependent populations: A retrospective analysis. *PLOS Med.* 15, e1002688.
- 943 111. Dasgupta, P., Great Britain, and Treasury (2021). *Economics of Biodiversity: The Dasgupta Review.*
- 944 112. Doherty, T.S., Glen, A.S., Nimmo, D.G., Ritchie, E.G., and Dickman, C.R. (2016). Invasive
945 predators and global biodiversity loss. *Proc. Natl. Acad. Sci. U. S. A.* 113, 11261–11265.
- 946 113. Dorcas, M.E., Willson, J.D., Reed, R.N., Snow, R.W., Rochford, M.R., Miller, M.A., Meshaka,
947 W.E., Andreadis, P.T., Mazzotti, F.J., Romagosa, C.M., et al. (2012). Severe mammal declines
948 coincide with proliferation of invasive Burmese pythons in Everglades National Park. *Proc. Natl.*
949 *Acad. Sci. U. S. A.* 109, 2418–2422.

- 950 114. Hoyer, I.J., Blosser, E.M., Acevedo, C., Thompson, A.C., Reeves, L.E., and Burkett-Cadena, N.D.
951 (2017). Mammal decline, linked to invasive Burmese python, shifts host use of vector mosquito
952 towards reservoir hosts of a zoonotic disease. *Biol. Lett.* *13*.
- 953 115. Walshe, D.P., Garner, P., Adeel, A.A., Pyke, G.H., and Burkot, T.R. (2017). Larvivorous fish for
954 preventing malaria transmission. *Cochrane Database Syst. Rev.* *12*, CD008090.
- 955 116. Mkoji, G.M., Hofkin, B.V., Kuris, A.M., Stewart-Oaten, A., Mungai, B.N., Kihara, J.H., Mungai, F.,
956 Yundu, J., Mbui, J., Rashid, J.R., et al. (1999). Impact of the crayfish *Procambarus clarkii* on
957 *Schistosoma haematobium* transmission in Kenya. *Am. J. Trop. Med. Hyg.* *61*, 751–759.
- 958 117. Angeler, D.G., Sánchez-Carrillo, S., García, G., and Alvarez-Cobelas, M. (2001). The influence of
959 *Procambarus clarkii* (Cambaridae, Decapoda) on water quality and sediment characteristics in a
960 Spanish floodplain wetland. *Hydrobiologia* *464*, 89–98.
- 961 118. De Lima, H., De Guglielmo, Z., Rodríguez, A., Convit, J., and Rodriguez, N. (2002). Cotton Rats
962 (*Sigmodon hispidus*) and Black Rats (*Rattus rattus*) as Possible Reservoirs of *Leishmania* spp. in Lara
963 State, Venezuela. *Mem. Inst. Oswaldo Cruz* *97*, 169–174.
- 964 119. Bhatt, S., Gething, P.W., Brady, O.J., Messina, J.P., Farlow, A.W., Moyes, C.L., Drake, J.M.,
965 Brownstein, J.S., Hoen, A.G., Sankoh, O., et al. (2013). The global distribution and burden of dengue.
966 *Nature* *496*, 504–507.
- 967 120. Reaser, J.K., Witt, A., Tabor, G.M., Hudson, P.J., and Plowright, R.K. (2021). Ecological
968 countermeasures for preventing zoonotic disease outbreaks: when ecological restoration is a human
969 health imperative. *Restor. Ecol.* *29*, e13357.
- 970 121. Tompkins, D.M., White, A.R., and Boots, M. (2003). Ecological replacement of native red squirrels
971 by invasive greys driven by disease. *Ecol. Lett.* *6*, 189–196.
- 972 122. Colla, S.R., Otterstatter, M.C., Gegear, R.J., and Thomson, J.D. (2006). Plight of the bumble bee:
973 Pathogen spillover from commercial to wild populations. *Biol. Conserv.* *129*, 461–467.
- 974 123. Mordecai, E.A. (2013). Despite spillover, a shared pathogen promotes native plant persistence in a
975 cheatgrass-invaded grassland. *Ecology* *94*, 2744–2753.
- 976 124. Allan, B.F., Dutra, H.P., Goessling, L.S., Barnett, K., Chase, J.M., Marquis, R.J., Pang, G., Storch,
977 G.A., Thach, R.E., and Orrock, J.L. (2010). Invasive honeysuckle eradication reduces tick-borne
978 disease risk by altering host dynamics. *Proc. Natl. Acad. Sci. U. S. A.* *107*, 18523–18527.
- 979 125. Lodge, D.M., Williams, S., MacIsaac, H.J., Hayes, K.R., Leung, B., Reichard, S., Mack, R.N.,
980 Moyle, P.B., Smith, M., Andow, D.A., et al. (2006). Biological Invasions: Recommendations for U.s.
981 Policy and Management. *Ecol. Appl.* *16*, 2035–2054.
- 982 126. Scheffers, B.R., Oliveira, B.F., Lamb, I., and Edwards, D.P. (2019). Global wildlife trade across the
983 tree of life. *Science* *366*, 71–76.
- 984 127. Zhang, L., Hua, N., and Sun, S. (2008). Wildlife trade, consumption and conservation awareness in
985 southwest China. *Biodivers. Conserv.* *17*, 1493–1516.

- 986 128. Rosen, G.E., and Smith, K.F. (2010). Summarizing the Evidence on the International Trade in Illegal
987 Wildlife. *EcoHealth* 7, 24–32.
- 988 129. Wyler, L.S., and Sheikh, P.A. International Illegal Trade in Wildlife: Threats and U.S. Policy. 26.
- 989 130. Smith, K.M., Zambrana-Torrel, C., White, A., Asmussen, M., Machalaba, C., Kennedy, S., Lopez,
990 K., Wolf, T.M., Daszak, P., Travis, D.A., et al. (2017). Summarizing US Wildlife Trade with an Eye
991 Toward Assessing the Risk of Infectious Disease Introduction. *EcoHealth* 14, 29–39.
- 992 131. Wolfe, N.D., Switzer, W.M., Carr, J.K., Bhullar, V.B., Shanmugam, V., Tamoufe, U., Prosser, A.T.,
993 Torimiro, J.N., Wright, A., Mpoudi-Ngole, E., et al. (2004). Naturally acquired simian retrovirus
994 infections in central African hunters. *Lancet Lond. Engl.* 363, 932–937.
- 995 132. Gómez, A., and Aguirre, A.A. (2008). Infectious diseases and the illegal wildlife trade. *Ann. N. Y.*
996 *Acad. Sci.* 1149, 16–19.
- 997 133. Pavlin, B.I., Schloegel, L.M., and Daszak, P. (2009). Risk of importing zoonotic diseases through
998 wildlife trade, United States. *Emerg. Infect. Dis.* 15, 1721–1726.
- 999 134. Bell, D., Robertson, S., and Hunter, P.R. (2004). Animal origins of SARS coronavirus: possible links
1000 with the international trade in small carnivores. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.* 359, 1107–
1001 1114.
- 1002 135. Wolfe, N.D., Daszak, P., Kilpatrick, A.M., and Burke, D.S. (2005). Bushmeat hunting, deforestation,
1003 and prediction of zoonoses emergence. *Emerg. Infect. Dis.* 11, 1822–1827.
- 1004 136. Travis, D.A., Watson, R.P., and Tauer, A. (2011). The spread of pathogens through trade in wildlife.
1005 *Rev. Sci. Tech. Int. Off. Epizoot.* 30, 219–239.
- 1006 137. Greutorex, Z.F., Olson, S.H., Singhalath, S., Silithammavong, S., Khammavong, K., Fine, A.E.,
1007 Weisman, W., Douangneun, B., Theppangna, W., Keatts, L., et al. (2016). Wildlife Trade and Human
1008 Health in Lao PDR: An Assessment of the Zoonotic Disease Risk in Markets. *PloS One* 11, e0150666.
- 1009 138. Judson, S.D., Fischer, R., Judson, A., and Munster, V.J. (2016). Ecological Contexts of Index Cases
1010 and Spillover Events of Different Ebolaviruses. *PLoS Pathog.* 12, e1005780.
- 1011 139. Cantlay, J.C., Ingram, D.J., and Meredith, A.L. (2017). A Review of Zoonotic Infection Risks
1012 Associated with the Wild Meat Trade in Malaysia. *EcoHealth* 14, 361–388.
- 1013 140. Huong, N.Q., Nga, N.T.T., Long, N.V., Luu, B.D., Latinne, A., Pruvot, M., Phuong, N.T., Quang,
1014 L.T.V., Hung, V.V., Lan, N.T., et al. (2020). Coronavirus testing indicates transmission risk increases
1015 along wildlife supply chains for human consumption in Viet Nam, 2013-2014. *PloS One* 15,
1016 e0237129.
- 1017 141. Kasper, K., Schweikhard, J., Lehmann, M., Ebert, C.L., Erbe, P., Wayakone, S., Nguyen, T.Q., Le,
1018 M.D., and Ziegler, T. (2020). The extent of the illegal trade with terrestrial vertebrates in markets and
1019 households in Khammouane Province, Lao PDR. *Nat. Conserv.* 41, 25–45.
- 1020 142. Hu, B., Zeng, L.-P., Yang, X.-L., Ge, X.-Y., Zhang, W., Li, B., Xie, J.-Z., Shen, X.-R., Zhang, Y.-Z.,
1021 Wang, N., et al. (2017). Discovery of a rich gene pool of bat SARS-related coronaviruses provides
1022 new insights into the origin of SARS coronavirus. *PLOS Pathog.* 13, e1006698.

- 1023 143. Ye, Z.-W., Yuan, S., Yuen, K.-S., Fung, S.-Y., Chan, C.-P., and Jin, D.-Y. (2020). Zoonotic origins
1024 of human coronaviruses. *Int. J. Biol. Sci.* *16*, 1686–1697.
- 1025 144. Nova, N. (2021). Cross-Species Transmission of Emerging Coronaviruses in Humans and Domestic
1026 Mammals (Preprints).
- 1027 145. Kan, B., Wang, M., Jing, H., Xu, H., Jiang, X., Yan, M., Liang, W., Zheng, H., Wan, K., Liu, Q., et
1028 al. (2005). Molecular evolution analysis and geographic investigation of severe acute respiratory
1029 syndrome coronavirus-like virus in palm civets at an animal market and on farms. *J. Virol.* *79*, 11892–
1030 11900.
- 1031 146. Ghana running out of sea food due to overfishing | Africanews
1032 <https://www.africanews.com/2019/07/30/overfishing/>.
- 1033 147. Brashares, J.S., Arcese, P., Sam, M.K., Coppolillo, P.B., Sinclair, A.R.E., and Balmford, A. (2004).
1034 Bushmeat Hunting, Wildlife Declines, and Fish Supply in West Africa. *Science* *306*, 1180–1183.
- 1035 148. Eskew, E.A., and Carlson, C.J. (2020). Overselling wildlife trade bans will not bolster conservation
1036 or pandemic preparedness. *Lancet Planet. Health* *4*, e215–e216.
- 1037 149. Leroy, E.M., Rouquet, P., Formenty, P., Souquière, S., Kilbourne, A., Froment, J.-M., Bermejo, M.,
1038 Smit, S., Karesh, W., Swanepoel, R., et al. (2004). Multiple Ebola Virus Transmission Events and
1039 Rapid Decline of Central African Wildlife. *Science* *303*, 387–390.
- 1040 150. Kurpiers, L.A., Schulte-Herbrüggen, B., Ejotre, I., and Reeder, D.M. (2016). Bushmeat and
1041 Emerging Infectious Diseases: Lessons from Africa. In *Problematic Wildlife: A Cross-Disciplinary*
1042 *Approach*, F. M. Angelici, ed. (Springer International Publishing), pp. 507–551.
- 1043 151. Hahn, B.H., Shaw, G.M., De Cock, K.M., and Sharp, P.M. (2000). AIDS as a zoonosis: scientific
1044 and public health implications. *Science* *287*, 607–614.
- 1045 152. Fa, J.E., Olivero, J., Farfán, M.Á., Márquez, A.L., Duarte, J., Nackoney, J., Hall, A., Dupain, J.,
1046 Seymour, S., Johnson, P.J., et al. (2015). Correlates of bushmeat in markets and depletion of wildlife.
1047 *Conserv. Biol.* *29*, 805–815.
- 1048 153. Ripple, W.J., Newsome, T.M., Wolf, C., Dirzo, R., Everatt, K.T., Galetti, M., Hayward, M.W.,
1049 Kerley, G.I.H., Levi, T., Lindsey, P.A., et al. (2015). Collapse of the world’s largest herbivores. *Sci.*
1050 *Adv.* *1*, e1400103.
- 1051 154. World Urbanization Prospects - Population Division - United Nations
1052 <https://population.un.org/wup/Publications/>.
- 1053 155. Becker, D.J., Streicker, D.G., and Altizer, S. (2018). Using host species traits to understand the
1054 consequences of resource provisioning for host–parasite interactions. *J. Anim. Ecol.* *87*, 511–525.
- 1055 156. Plowright, R.K., Foley, P., Field, H.E., Dobson, A.P., Foley, J.E., Eby, P., and Daszak, P. (2011).
1056 Urban habituation, ecological connectivity and epidemic dampening: the emergence of Hendra virus
1057 from flying foxes (*Pteropus* spp.). *Proc. Biol. Sci.* *278*, 3703–3712.

- 1058 157. Kessler, M.K., Becker, D.J., Peel, A.J., Justice, N.V., Lunn, T., Crowley, D.E., Jones, D.N., Eby, P.,
1059 Sánchez, C.A., and Plowright, R.K. (2018). Changing resource landscapes and spillover of
1060 henipaviruses. *Ann. N. Y. Acad. Sci.* *1429*, 78–99.
- 1061 158. Plowright, R.K., Peel, A.J., Streicker, D.G., Gilbert, A.T., McCallum, H., Wood, J., Baker, M.L.,
1062 and Restif, O. (2016). Transmission or Within-Host Dynamics Driving Pulses of Zoonotic Viruses in
1063 Reservoir–Host Populations. *PLoS Negl. Trop. Dis.* *10*, e0004796.
- 1064 159. Smith, B.L. (2001). Winter Feeding of Elk in Western North America. *J. Wildl. Manag.* *65*, 173–
1065 190.
- 1066 160. Jones, J.D., Kauffman, M.J., Monteith, K.L., Scurlock, B.M., Albeke, S.E., and Cross, P.C. (2014).
1067 Supplemental feeding alters migration of a temperate ungulate. *Ecol. Appl.* *24*, 1769–1779.
- 1068 161. Brennan, A., Cross, P.C., Portacci, K., Scurlock, B.M., and Edwards, W.H. (2017). Shifting
1069 brucellosis risk in livestock coincides with spreading seroprevalence in elk. *PLOS ONE* *12*, e0178780.
- 1070 162. Couch, C.E., Wise, B.L., Scurlock, B.M., Rogerson, J.D., Fuda, R.K., Cole, E.K., Szcodronski, K.E.,
1071 Sepulveda, A.J., Hutchins, P.R., and Cross, P.C. (2021). Effects of supplemental feeding on the fecal
1072 bacterial communities of Rocky Mountain elk in the Greater Yellowstone Ecosystem. *PLOS ONE* *16*,
1073 e0249521.
- 1074 163. Berdegúe, J.A., Rosada, T., and Bebbington, A.J. (2014). The Rural Transformation. In *International*
1075 *Development: Ideas, Experience, and Prospects* (Oxford University Press), pp. 463–478.
- 1076 164. Santiago-Alarcon, D., and MacGregor-Fors, I. (2020). Cities and pandemics: urban areas are ground
1077 zero for the transmission of emerging human infectious diseases. *J. Urban Ecol.* *6*.
- 1078 165. Naguib, M.M., Li, R., Ling, J., Grace, D., Nguyen-Viet, H., and Lindahl, J.F. (2021). Live and Wet
1079 Markets: Food Access versus the Risk of Disease Emergence. *Trends Microbiol.* *0*.
- 1080 166. Wood, C.L., McInturff, A., Young, H.S., Kim, D., and Lafferty, K.D. (2017). Human infectious
1081 disease burdens decrease with urbanization but not with biodiversity. *Philos. Trans. R. Soc. Lond. B.*
1082 *Biol. Sci.* *372*.
- 1083 167. Moore, M., Gould, P., and Keary, B.S. (2003). Global urbanization and impact on health. *Int. J. Hyg.*
1084 *Environ. Health* *206*, 269–278.
- 1085 168. Plowright, R.K., Reaser, J.K., Locke, H., Woodley, S.J., Patz, J.A., Becker, D.J., Oppler, G.,
1086 Hudson, P.J., and Tabor, G.M. (2021). Land use-induced spillover: a call to action to safeguard
1087 environmental, animal, and human health. *Lancet Planet. Health* *5*, e237–e245.
- 1088 169. Tortosa, F.S., Caballero, J.M., and Reyes-López, J. (2002). Effect of Rubbish Dumps on Breeding
1089 Success in the White Stork in Southern Spain. *Waterbirds Int. J. Waterbird Biol.* *25*, 39–43.
- 1090 170. Satterfield, D.A., Maerz, J.C., and Altizer, S. (2015). Loss of migratory behaviour increases infection
1091 risk for a butterfly host. *Proc. Biol. Sci.* *282*, 20141734.
- 1092 171. Stoner, C., Caro, T., Mduma, S., Mlingwa, C., Sabuni, G., and Borner, M. (2007). Assessment of
1093 Effectiveness of Protection Strategies in Tanzania Based on a Decade of Survey Data for Large
1094 Herbivores. *Conserv. Biol.* *21*, 635–646.

- 1095 172. McIntosh, A.R., McHugh, P.A., Plank, M.J., Jellyman, P.G., Warburton, H.J., and Greig, H.S.
1096 (2018). Capacity to support predators scales with habitat size. *Sci. Adv.* *4*, eaap7523.
- 1097 173. Vittor, A.Y., Pan, W., Gilman, R.H., Tielsch, J., Glass, G., Shields, T., Sánchez-Lozano, W., Pinedo,
1098 V.V., Salas-Cobos, E., Flores, S., et al. (2009). Linking deforestation to malaria in the Amazon:
1099 characterization of the breeding habitat of the principal malaria vector, *Anopheles darlingi*. *Am. J.*
1100 *Trop. Med. Hyg.* *81*, 5–12.
- 1101 174. Xia, S., Dweck, H., Lutomiah, J., Sang, R., McBride, C., Rose, N., Ayala, D., and Powell, J. (2021).
1102 Larval breeding sites of the mosquito *Aedes aegypti* in forest and domestic habitats in Africa and the
1103 potential association with oviposition evolution (Preprints).
- 1104 175. Stephenson, E.B., Peel, A.J., Reid, S.A., Jansen, C.C., and McCallum, H. (2018). The non-human
1105 reservoirs of Ross River virus: a systematic review of the evidence. *Parasit. Vectors* *11*, 188.
- 1106 176. Kain, M.P., Skinner, E., McCallum, H., and Mordecai, E.A. Physiology and ecology combine to
1107 determine host and vector importance for Ross River virus and other vector-borne diseases. *54*.
- 1108 177. Stephens, C.R., González-Salazar, C., Sánchez-Cordero, V., Becker, I., Rebollar-Tellez, E.,
1109 Rodríguez-Moreno, Á., Berzunza-Cruz, M., Balcells, C.D., Gutiérrez-Granados, G., Hidalgo-Mihart,
1110 M., et al. (2016). Can You Judge a Disease Host by the Company It Keeps? Predicting Disease Hosts
1111 and Their Relative Importance: A Case Study for Leishmaniasis. *PLoS Negl. Trop. Dis.* *10*, e0005004.
- 1112 178. Jansen, A.M., Xavier, S.C. das C., and Roque, A.L.R. (2018). *Trypanosoma cruzi* transmission in the
1113 wild and its most important reservoir hosts in Brazil. *Parasit. Vectors* *11*, 502.
- 1114 179. Kilpatrick, A.M., Daszak, P., Jones, M.J., Marra, P.P., and Kramer, L.D. (2006). Host heterogeneity
1115 dominates West Nile virus transmission. *Proc. Biol. Sci.* *273*, 2327–2333.
- 1116 180. Brown, H.E., Childs, J.E., Diuk-Wasser, M.A., and Fish, D. (2008). Ecological factors associated
1117 with West Nile virus transmission, northeastern United States. *Emerg. Infect. Dis.* *14*, 1539–1545.
- 1118 181. Bradley, C.A., Gibbs, S.E.J., and Altizer, S. (2008). Urban land use predicts West Nile virus
1119 exposure in songbirds. *Ecol. Appl. Publ. Ecol. Soc. Am.* *18*, 1083–1092.
- 1120 182. Bowden, S.E., Magori, K., and Drake, J.M. (2011). Regional differences in the association between
1121 land cover and West Nile virus disease incidence in humans in the United States. *Am. J. Trop. Med.*
1122 *Hyg.* *84*, 234–238.
- 1123 183. Nolan, M.S., Schuermann, J., and Murray, K.O. (2013). West Nile Virus Infection among Humans,
1124 Texas, USA, 2002–2011. *Emerg. Infect. Dis.* *19*, 137–139.
- 1125 184. Kain, M.P., and Bolker, B.M. (2019). Predicting West Nile virus transmission in North American
1126 bird communities using phylogenetic mixed effects models and eBird citizen science data. *Parasit.*
1127 *Vectors* *12*, 395.
- 1128 185. *THE 17 GOALS*. (n.d.). United Nations. Retrieved June 7, 2021, from <https://sdgs.un.org/goals>
- 1129 186. *Nature's contributions to people*. (2017, August 12). IPBES. [http://www.ipbes.net/glossary/natures-](http://www.ipbes.net/glossary/natures-contributions-people)
1130 [contributions-people](http://www.ipbes.net/glossary/natures-contributions-people)

- 1131 187. *IUCN Global Standard for NbS*. (2020, July 14). IUCN. [https://www.iucn.org/theme/nature-based-](https://www.iucn.org/theme/nature-based-solutions/resources/iucn-global-standard-nbs)
1132 solutions/resources/iucn-global-standard-nbs
- 1133 188. *Bridge Collaborative*. (n.d.). Bridge Collaborative. Retrieved June 9, 2021, from
1134 <https://bridgecollaborativeglobal.org/>
- 1135 189. *Climate Change and Health*. (n.d.). PAHO/WHO | Pan American Health Organization. Retrieved
1136 June 7, 2021, from <https://www.paho.org/en/topics/climate-change-and-health>
- 1137 190. *Global Health Security Agenda*. (n.d.). Global Health Security Agenda. Retrieved June 7, 2021, from
1138 <https://ghsagenda.org/>
- 1139 191. *The FAO-OIE-WHO Collaboration* (pp. 1–6). (2010). FAO, OIE, WHO.
- 1140 192. Hopkins, S.R. Lafferty, K.D., Wood, C.L. Olson, S.H. Buck, J.C., de Leo, G.A., Fiorella, K.J.,
1141 Fornberg, J.L., Garchitorena, A., Jones, I.J., et al. In review. Diversity and evidence gaps among
1142 potential win–win solutions for conservation and human infectious disease control. *Lancet Planet*
1143 *Health*.
- 1144 193. *About The Challenge*. (n.d.). Bon Challenge. Retrieved June 7, 2021, from
1145 <https://www.bonchallenge.org/about>
- 1146 194. *Nature-based Solutions*. (2016, September 27). Commission on Ecosystem Management.
1147 [https://www.iucn.org/commissions/commission-ecosystem-management/our-work/nature-based-](https://www.iucn.org/commissions/commission-ecosystem-management/our-work/nature-based-solutions)
1148 solutions
- 1149 195. *POST2020 Global Biodiversity Framework: Discussion Paper* (pp. 1–10). (2019). Convention on
1150 Biological Diversity. <https://www.cbd.int/conferences/post2020/post2020-prep-01/documents>
- 1151 196. *CITES*. (n.d.). WWF. Retrieved June 7, 2021, from <https://www.worldwildlife.org/pages/cites>
- 1152 197. *Campaign For nature*. (n.d.). Campaign for Nature. Retrieved June 9, 2021, from
1153 <https://www.campaignfornature.org>.
- 1154 198. *30x30: NRDC's Commitment to Protect Nature and Life on Earth* | NRDC. (n.d.). Retrieved June 7,
1155 2021, from <https://www.nrdc.org/30x30-nrdcs-commitment-protect-nature-and-life-earth>.
- 1156 199. February 07, and O'Shea, 2020 Alison Chase Helen Saving Nature Will Take Bold Action: "Thirty
1157 by Thirty." NRDC. [https://www.nrdc.org/experts/alison-chase/saving-nature-will-take-bold-action-](https://www.nrdc.org/experts/alison-chase/saving-nature-will-take-bold-action-thirty-thirty)
1158 thirty-thirty.
- 1159 200. Executive Order on Tackling the Climate Crisis at Home and Abroad (2021). White House.
1160 [https://www.whitehouse.gov/briefing-room/presidential-actions/2021/01/27/executive-order-on-](https://www.whitehouse.gov/briefing-room/presidential-actions/2021/01/27/executive-order-on-tackling-the-climate-crisis-at-home-and-abroad/)
1161 tackling-the-climate-crisis-at-home-and-abroad/.
- 1162 201. *Conserving and Restoring America the Beautiful* (pp. 1–24). (2021). US Department of the Interior,
1163 US Department of Agriculture, US Department of Commerce, Council on Environmental Quality.
- 1164 202. Grima, N., Singh, S.J., Smetschka, B., and Ringhofer, L. (2016). Payment for Ecosystem Services
1165 (PES) in Latin America: Analysing the performance of 40 case studies. *Ecosyst. Serv.* 17, 24–32.

- 1166 203. Schomers, S., and Matzdorf, B. (2013). Payments for ecosystem services: A review and comparison
1167 of developing and industrialized countries. *Ecosyst. Serv.* *6*, 16–30.
- 1168 204. Sánchez-Azofeifa, G.A., Pfaff, A., Robalino, J.A., and Boomhower, J.P. (2007). Costa Rica's
1169 Payment for Environmental Services Program: Intention, Implementation, and Impact. *Conserv. Biol.*
1170 *21*, 1165–1173.
- 1171 205. *FONAFIFO | Sitio Web*. (n.d.). Retrieved June 7, 2021, from
1172 <https://www.fonafifo.go.cr/en/conozcanos/proyectos-finalizados/>
- 1173 206. Arriagada, R.A., Sills, E.O., Ferraro, P.J., and Pattanayak, S.K. (2015). Do Payments Pay Off?
1174 Evidence from Participation in Costa Rica's PES Program. *PLOS ONE* *10*, e0131544.
- 1175 207. Rosenstock, T.S., Dawson, I.K., Aynekulu, E., Chomba, S., Degrande, A., Fornace, K., Jamnadass,
1176 R., Kimaro, A., Kindt, R., Lamanna, C., et al. (2019). A Planetary Health Perspective on Agroforestry
1177 in Sub-Saharan Africa. *One Earth* *1*, 330–344.
- 1178 208. Sokolow, S.H., Nova, N., Pepin, K.M., Peel, A.J., Pulliam, J.R.C., Manlove, K., Cross, P.C., Becker,
1179 D.J., Plowright, R.K., McCallum, H., et al. (2019). Ecological interventions to prevent and manage
1180 zoonotic pathogen spillover. *Philos. Trans. R. Soc. B Biol. Sci.* *374*, 20180342.
- 1181 209. *Project Finance for Permanence: Key Outcomes and Lessons Learned* (pp. 1–15). (2015). WWF.
1182 [https://www.worldwildlife.org/publications/project-finance-for-permanence-key-outcomes-and-](https://www.worldwildlife.org/publications/project-finance-for-permanence-key-outcomes-and-lessons-learned)
1183 [lessons-learned](https://www.worldwildlife.org/publications/project-finance-for-permanence-key-outcomes-and-lessons-learned)
- 1184 210. *Assessments of three landscape-scale conservation deals: ARPA, Great Bear, and Forever Costa*
1185 *Rica* (pp. 1–22). (2011). Linden Trust for Conservation, Gordon and Betty Moore Foundation,
1186 Redstone Strategy Group, LLC. [https://www.redstonestrategy.com/wp-content/uploads/2016/07/2013-](https://www.redstonestrategy.com/wp-content/uploads/2016/07/2013-01-04-PFP-Paper.pdf)
1187 [01-04-PFP-Paper.pdf](https://www.redstonestrategy.com/wp-content/uploads/2016/07/2013-01-04-PFP-Paper.pdf)
- 1188 211. MacDonald, A.J., and Mordecai, E.A. (2019). Amazon deforestation drives malaria transmission,
1189 and malaria burden reduces forest clearing. *Proc. Natl. Acad. Sci.* *116*, 22212–22218.
- 1190 212. Plowright, R.K., Becker, D.J., McCallum, H., and Manlove, K.R. (2019). Sampling to elucidate the
1191 dynamics of infections in reservoir hosts. *Philos. Trans. R. Soc. B Biol. Sci.* *374*, 20180336.
- 1192 213. Baker, S.E., Cain, R., van Kesteren, F., Zommers, Z.A., D'Cruze, N., and Macdonald, D.W. (2013).
1193 Rough Trade: Animal Welfare in the Global Wildlife Trade. *BioScience* *63*, 928–938.
- 1194 214. Becker, D.J., Albery, G.F., Kessler, M.K., Lunn, T.J., Falvo, C.A., Czirják, G.Á., Martin, L.B., and
1195 Plowright, R.K. (2020). Macroimmunology: The drivers and consequences of spatial patterns in
1196 wildlife immune defence. *J. Anim. Ecol.* *89*, 972–995.
- 1197 215. Young, H.S., McCauley, D.J., Galetti, M., and Dirzo, R. (2016). Patterns, Causes, and Consequences
1198 of Anthropocene Defaunation. *Annu. Rev. Ecol. Evol. Syst.* *47*, 333–358.
- 1199 216. McFadden, T., and Dirzo, R. In review. Harnessing multilayer networks to predict metacommunity
1200 responses to global environmental change. *J. Ecol.*

1201 217. Olival, K.J., Cryan, P.M., Amman, B.R., Baric, R.S., Blehert, D.S., Brook, C.E., Calisher, C.H.,
1202 Castle, K.T., Coleman, J.T.H., Daszak, P., et al. (2020). Possibility for reverse zoonotic transmission
1203 of SARS-CoV-2 to free-ranging wildlife: A case study of bats. *PLOS Pathog.* *16*, e1008758.

1204 218. Naeem, S., Duffy, J.E., and Zavaleta, E. (2012). The Functions of Biological Diversity in an Age of
1205 Extinction. *Science* *336*, 1401–1406.

1206

Box 1. Examples of dimensions of biodiversity

Biodiversity is made up of a number of dimensions, with multiple axes affecting zoonotic spillover risk. Below are a handful of examples described by [218]:

- Genetic diversity: Aspects of genomic variability, including nucleotide, allelic, chromosomal, and genotypic.
- Taxonomic diversity: The number and relative abundance of taxa (e.g., species, genera, and onward). Disease-diversity relationships are typically described within the context of species richness.
- Phylogenetic diversity: Relationships among taxa based upon phylogenetic distance (i.e., amount of time the most recent common ancestor of both taxa).
- Spatial or temporal diversity: Rates of turnover of taxa through space and time.
- Interaction diversity: Characteristics of the network of linkages, such as abundance and variation in type and strength, defined by biotic interactions, including competition, facilitation, predation, and contact.
- Landscape diversity: Number, relative abundance, and distribution of different habitat types within a landscape (i.e., habitat diversity).
- Functional diversity: Variation in the degree of expression of multiple functional traits, i.e., the different types of processes in a community that are important to its structure and dynamic stability.

Table 1. Case studies of mechanisms connecting anthropogenic disturbance with biodiversity change and its downstream effects on infectious diseases. Figure 1 illustrates the overall framework for linking anthropogenic disturbance to biodiversity change to disease spillover via the spillover layers being affected in each case study.

Anthropogenic disturbance	Biodiversity change (type and direction)	Mechanisms of biodiversity change	Infectious disease case studies			
			Spillover layers affected	Disease impacts	No. in Figure 1	References
Agricultural expansion & intensification	Functional diversity (decreased)	Loss of large consumers increases rodent richness and abundance	Pathogen prevalence in wildlife host	Increased prevalence of Bartonella in rodents in Kenya	1	40
	Landscape diversity (decreased)	Resources become limited, pushing animals into human modified landscapes	Pathogen prevalence in wildlife host; human exposure to pathogen	Increased prevalence and spillover (zoonotic transmission) of <i>P. knowlesi</i> in Borneo	2	53
Climate change	Spatiotemporal diversity (increased)	Polar species replaced by migrating nonpolar species (via predation and resource competition)	Wildlife host density & distribution; pathogen survival & spread; human exposure to pathogen	Increased spillover risk of rabies in Alaska as a polar reservoir of rabies (Arctic fox) is being replaced by a more human-landscape adaptable reservoir species (red fox)	3	95, 100
	Taxonomic and interaction diversity (increased)	Drought and reduction in water resources leads to increased density and diversity of hosts around shared water resources	Wildlife host density & distribution	Increased spillover risk of <i>E. coli</i> in Botswana	4	105, 110
Invasive species	Taxonomic, functional, and interaction diversity (decreased)	Introduction of Burmese python reduces abundance of large- and medium-sized mammals	Human exposure to pathogen	Increased spillover risk of Everglade virus in Florida as mosquito disease vectors feed on rodent reservoirs more frequently	5	114
Wildlife trade	Taxonomic, phylogenetic, genetic, functional, interaction, and landscape diversity (decreased)	Removal of wild, mostly large-bodied animals (via hunting, trapping, transfer, killing) or overfishing directly reduces abundance and diversity of terrestrial and marine wildlife species	Wildlife host susceptibility to infection; pathogen shedding in wildlife host; pathogen survival and spread; human exposure to pathogen	Increased spillover risk of Ebola in the Congo Basin as demand for wild meat from small-bodied mammals such as bats (Ebola reservoirs) increases (hunters and preparers of the bushmeat are exposed to bat bites, scratches, or blood)	6	147, 149, 152, 153

Urbanization	Temporal diversity of habitats (decreased)	Resources become limited, pushing migrating animals to form resident populations in human modified landscapes	Pathogen prevalence and shedding in wildlife host; human exposure to pathogen	Increased prevalence, shedding, and spill of Hendra virus	7	16
Urbanization, Wildlife trade	Taxonomic and interaction diversity (increased)	Wildlife markets aggregate novel assemblages of hosts, increasing host richness that is unique to markets and the food supply chain	Wildlife host density & distribution, susceptibility to infection, and pathogen shedding (and human–human transmission)	Increased wildlife susceptibility to infection, reservoir density, pathogen shedding and spread of SARS viruses	8	140, 144, 145

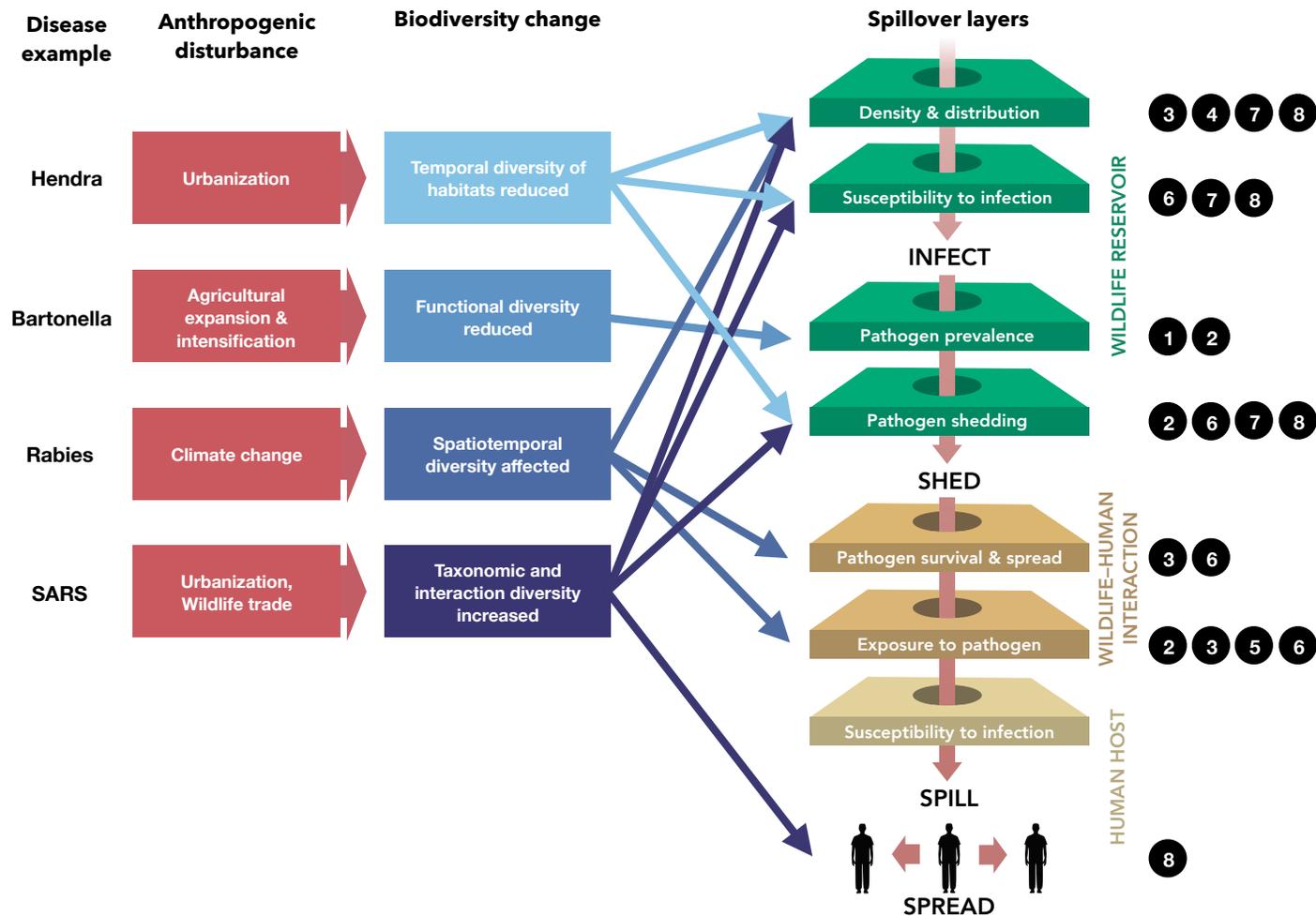


Figure 1. The anthropogenic disturbance, biodiversity change, and spillover cascade. To understand mechanisms connecting anthropogenic disturbance with spillover via biodiversity change, it is imperative to investigate how anthropogenic disturbance impacts biodiversity, and how those effects drive the perforation of the layers (intermediate processes) leading to spillover (shown using four case studies from Table 1 as examples). Zoonotic spillover arises from the alignment of multiple processes (depicted as layers). Apart from human susceptibility to infection, we found that each layer can be affected by biodiversity change, especially when considering biodiversity along multiple axes (Box 1). Connecting biodiversity change to explicit processes helps us to better understand how, when, and why biodiversity change impacts zoonotic disease risk. Numbers next to each layer correspond to case studies highlighted in Table 1.