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1     **Reliability of recharge rates estimated from groundwater**

2                     **age with a simplified analytical approach**

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24 **Abstract**

25 Groundwater age is often used to estimate groundwater recharge through a simplified  
26 analytical approach. This estimated recharge is thought to be representative of the  
27 mean recharge between the point of entry and the sampling point. However, given the  
28 complexity in actual recharge, whether the mean recharge is reasonable is still  
29 unclear. This study examined the validity of the method to estimate long-term average  
30 groundwater recharge and the possibility of obtaining reasonable spatial recharge  
31 pattern. We first validated our model in producing reasonable age distributions using a  
32 constant flux boundary condition. We then generated different flow fields and age  
33 patterns by using various spatially-varying flux boundary conditions with different  
34 magnitudes and wavelengths. Groundwater recharge was estimated and analyzed  
35 afterwards using the method at the spatial scale. We illustrated the main findings with  
36 a field example in the end. Our results suggest that we can estimate long-term average  
37 groundwater recharge with 10% error in many parts of an aquifer. The size of these  
38 areas decreases with the increase in both the amplitude and the wavelength. The  
39 chance of obtaining a reasonable groundwater recharge is higher if an age sample is  
40 collected from the middle of an aquifer and at downstream areas. Our study also  
41 indicates that the method can also be used to estimate local groundwater recharge if  
42 age samples are collected close to the water table. However, care must be taken to

43 determine groundwater age regardless of conditions.

44

45 Key words: simplified analytical approach; groundwater age; spatial groundwater  
46 recharge

## 47 **1. Introduction**

48 Groundwater recharge is defined as water reaching water table to replenish an aquifer.

49 It occurs in a number of forms including diffuse recharge, focused recharge and

50 artificial recharge and is an important component of a groundwater water balance (de

51 Vries & Simmers, 2002). Recharge rates are critical for determining sustainable

52 groundwater extraction rates, properly understanding groundwater flow and storage

53 and reducing uncertainty in groundwater numerical models. A variety of methods

54 have been developed to estimate groundwater recharge, such as soil water balance,

55 water table fluctuation and age tracer methods (Allison, Gee, & Tyler, 1994; Healy,

56 2010; Scanlon, Healy, & Cook, 2002). The age tracer method is often used in recharge

57 estimation because groundwater age usually integrates flow information across large

58 areas and over long periods (Cook & Böhlke, 2000).

59

60 The use of age tracers to estimate groundwater recharge needs to rely on a simplified

61 analytical method, introduced by Vogel (1967). The equation is given by

$$62 \quad R_v = \frac{H\theta}{t} \ln \left( \frac{H}{H-h} \right) \quad (1)$$

63 where  $R_v$  is the Vogel-based recharge [ $L T^{-1}$ ],  $t$  is the time since recharge at a point in

64 the aquifer [T],  $\theta$  is the porosity [-],  $H$  and  $h$  are the aquifer thickness [L] and the

65 vertical distance between the point of sampling and the land surface [L], respectively.

66 This method is based on several assumptions including horizontal flow, uniform  
67 recharge, homogeneous aquifer material and negligible dispersion effect. The relevant  
68 conceptualization is shown in Figure 1. Because this method requires very few  
69 parameters compared to some other analytical equations (Chesnaux, Molson, &  
70 Chapuis, 2005; Chesnaux, Santoni, Garel, & Huneau, 2018), it has been widely used  
71 in many different conditions despite assumptions made (Hinkle, Böhlke, Duff,  
72 Morgan, & Weick, 2007; Harrington, Cook, & Herczeg, 2002; Hagedorn, El-Kadi,  
73 Mair, Whittier, & Ha, 2011; Kozuskanich, Simmons, & Cook, 2014; McMahon,  
74 Plummer, Böhlke, Shapiro, & Hinkle, 2011). For example, Hagedorn, El-Kadi, Mair,  
75 Whittier, & Ha (2011) used this method to estimate recharge in a fractured aquifer on  
76 the Jeju Island by assuming that groundwater flow in fractures and porous media are  
77 equivalent. Hinkle, Böhlke, Duff, Morgan, & Weick (2007) estimated large-scale  
78 recharge with this method in order to examine distribution of nitrate and ammonium  
79 in groundwater.

80

81 Compared to the assumption of this analytical method, actual groundwater recharge is  
82 known to vary in space because of spatial distributions of controlling factors including  
83 climate, soil and vegetation (Cook & Böhlke, 2000; Cook, Walk, & Jolly, 1989; Kim  
84 & Jackson, 2012; Keese, Scanlon, & Reedy, 2005; Scanlon et al., 2006; Xie, Crosbie,  
85 Simmons, Cook, & Zhang, 2019; Xie, Crosbie, Yang, Wu, & Wang, 2018). It is  
86 evident that different spatial patterns in recharge may result in different flow fields  
87 which lead to different groundwater age patterns. Hence, groundwater recharge

88 estimated from the Vogel method would differ from location to location. Whether and  
89 under what conditions this simplified analytical method still produces representative  
90 long-term average estimates is still unclear.

91

92 Spatially varying groundwater recharge is one of several important factors in  
93 determining groundwater flow fields (Sanford, 2002). Estimating spatial groundwater  
94 recharge has been difficult. Therefore, many studies usually use a limited number of  
95 estimates to represent large areas (e.g., Xie, Cook, Shanafield, Simmons, & Zheng,  
96 2016; Yang et al., 2019). Groundwater tracers underneath water table result mostly  
97 from local recharge and so groundwater age must vary spatially (McMahon, Bohlke,  
98 & Carney, 2007). It is likely that the analytical method could also estimate spatial  
99 recharge and infer water table behavior by examining groundwater age at shallow  
100 parts of aquifers (McMahon, Bohlke, & Carney, 2007; Wood, Cook, & Harrington,  
101 2015). However, as the Vogel method was developed based on several assumptions,  
102 whether it can yield reasonable results warrants careful assessment.

103

104 This study attempted to examine whether the Vogel method is capable of estimating  
105 long-term average groundwater recharge and the possibility of estimating spatially-  
106 varying groundwater recharge with such a method. We first established a groundwater  
107 flow and age transport model with a constant recharge rate along the top boundary  
108 and examined it by comparing groundwater age distribution to that calculated with the  
109 analytical method. Afterwards, we changed constant groundwater recharge to different

spatially-varying recharge that varies in different sinusoidal patterns. The simulated groundwater age in the model was used to estimate recharge with the analytical method. The estimated recharge was then compared to the actual recharge to shed light on the feasibility of the Vogel method for estimating recharge under spatially changing conditions. A field example was employed in the end to illustrate how our theoretical analysis could assist in the interpretation of real-world data.

## 2. Methods

### 2.1 Governing equations

Several numerical experiments were performed to investigate groundwater age distribution and examine the validity of the simple method to estimate groundwater recharge. To simulate groundwater flow and age transport simultaneously, the numerical code HydroGeoSphere (Aquanty Inc., 2018) was selected. HydroGeoSphere uses the 3-D Richards equation to simulate variably saturated groundwater flow. As our system is fully saturated, the equation can be simplified into the follow form:

$$S_s \frac{\partial \psi}{\partial t} = -\nabla \cdot \mathbf{q} + Q \quad (2)$$

Where  $S_s$  is the specific storage [ $L^{-1}$ ],  $Q$  is the source and sink [ $T^{-1}$ ],  $\mathbf{q}$  is the Darcy flux tensor [ $L T^{-1}$ ],  $\psi$  is the pressure head [ $L$ ].  $\mathbf{q}$  is given by

$$\mathbf{q} = -\mathbf{K} \cdot \nabla (\psi + z) \quad (3)$$

where  $\mathbf{K}$  is the hydraulic conductivity tensor [ $L T^{-1}$ ] and  $z$  is the elevation head [ $L$ ].

Groundwater age transport was simulated with the mean age approach (Goode, 1996). As a groundwater sample contains a large number of water particles, mean age is usually used in analyses. It is reasonable to assume that the mean age of mixed water particles is a volume-weighted average if density is constant, and so the mean age will vary in a similar fashion to the concentration of a conservative solute. Hence, mean age transport can be described by a variant of the advection-dispersion equation. The governing equation is given by

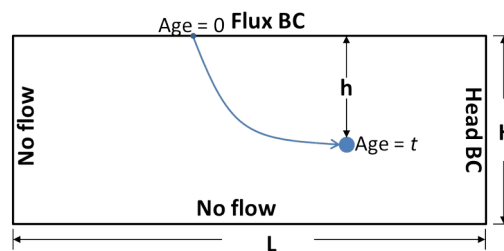
$$\frac{\partial(\phi A)}{\partial t} = -\nabla \cdot A \mathbf{q} + \nabla \cdot (\mathbf{D} \cdot \nabla A) + \phi + Q_A \quad (4)$$

where  $A$  is the mean groundwater age [T],  $\phi$  is the porosity [-],  $\mathbf{D}$  is the hydrodynamic dispersion tensor [L] and  $Q_A$  is the source/sink term of the age mass [-]. The first two terms on the right-hand side of the equation describe advective and dispersive transport of age mass, respectively. The third term is age increase at the rate of 1 year per year along flow paths. The last term represents the exchange of age mass with other domains. The reader is referred to Aquanty Inc. (2018) for a detailed description of numerical implementation.

## 2.2 Model setup

The numerical model employed is rectangular in shape (10,000 m in length and 100 m in thickness) as shown in Figure 1. Kozuskanich, Simmons, & Cook (2014) examined the effect of aquifer heterogeneity on the applicability of the Vogel method with the same model setup but smaller size. The model was used to perform groundwater flow and mean groundwater age transport simulations. For the groundwater flow

simulation, a flux boundary condition was applied to the top surface to represent actual groundwater recharge. A constant rate of 100 mm/y was used in this study. A head boundary condition of 100 m was specified to the right-hand side boundary. All the other sides were given no-flow boundary conditions. It should be noted that this model represents an unconfined aquifer with the sloping bottom despite the rectangular shape. For the mean age transport simulation, a mean age of zero was specified at the top boundary where the actual recharge occurred. The model was discretized with a grid size of 2 m vertically and 10 m laterally, resulting in a total of 50,000 elements. A porosity  $\theta$  of 0.3 and a hydraulic conductivity  $K$  of 0.8 m/d were specified to the model. Note that  $K$  does not affect modeling results given flux boundary conditions used. For the flow simulation, the model was simulated in steady state. For the age transport simulation, the model was run for 10,000 years. If evident change was observed between the last two time steps, the model was run again for 10,000 years.



**Figure 1** Schematic diagram of the model used in this study.  $L$  and  $H$  are 10,000 m and 100 m, respectively. A constant flux of 100 mm/y was used at the top boundary and a head of 100 m was specified to the right-hand side boundary. These values were chosen for demonstrative purposes only. Other values could also be used.



173

174 **2.3 Spatially-varying actual recharge**

175 In order to examine the validity of the Vogel method to estimate groundwater  
 176 recharge, we generated groundwater flow fields with different spatially-varying actual  
 177 recharge conditions along the top surface, realized through flux boundary conditions.

178 The spatially-varying actual recharge is given by

$$179 \quad R_m = A \left( \sin \frac{2\pi}{B} x \right) + C \quad (5)$$

180 where  $R_m$  is the modelled recharge [ $L T^{-1}$ ],  $A$ ,  $B$  and  $C$  represent the amplitude [ $L T^{-1}$ ],  
 181 the wavelength [ $L$ ] and the mean recharge [ $L T^{-1}$ ], respectively,  $x$  is the distance from  
 182 the left end of the top surface [ $L$ ].  $R_m$  is the theoretical recharge as opposed to the  
 183 estimated recharge with the Vogel method  $R_V$ . In order to be consistent with  
 184 conventional uses, we use millimeters per year for a recharge rate and meters for  
 185 length.

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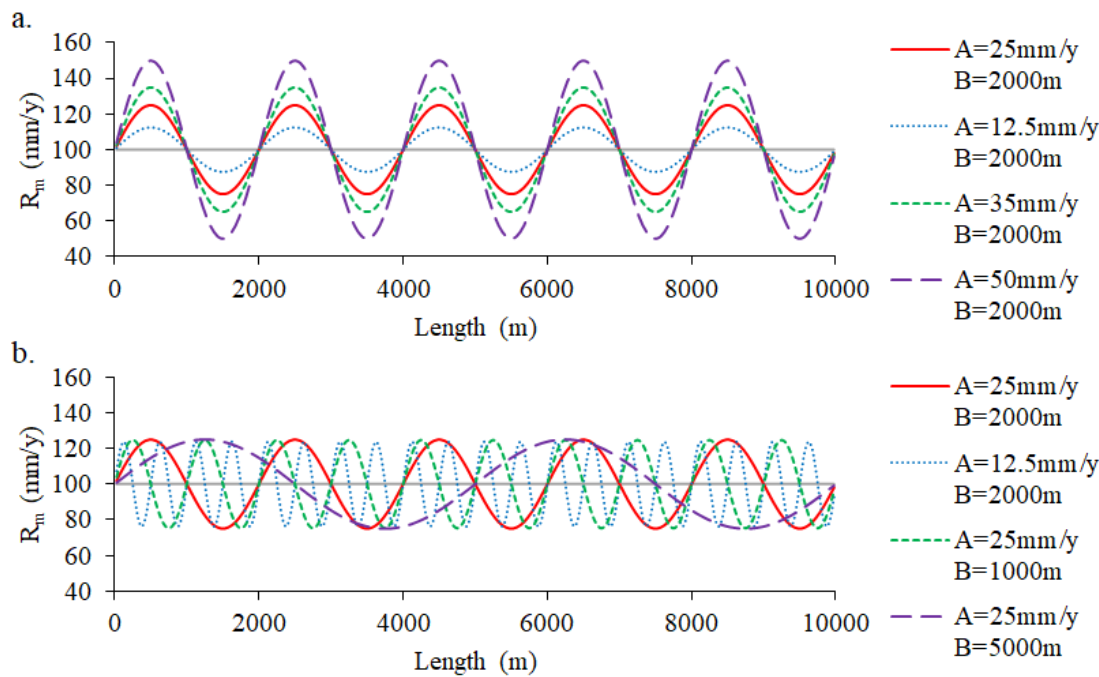
187 Two types of variations in the recharge were used, including changes in the amplitude  
 188 and changes in the wavelength. Four recharge amplitudes were considered including  
 189 12.5, 25, 35 and 50 mm/y and six recharge wavelengths were simulated including  
 190 500, 1000, 1250, 2000, 3500 and 5000 m.  $C$  was constant at 100 mm/y for all the  
 191 scenarios. Figure 2a and Figure 2b demonstrate how  $R_m$  varies with  $A$  and  $B$  spatially,  
 192 respectively. Apart from the mean recharge of 100 mm/y, we also examined several  
 193 different mean values, including 150 mm/y, 50 mm/y. The corresponding amplitudes  
 194 were varied between 10% and 50% of the mean values, and the wavelengths remained

195 unchanged. A total of 72 scenarios were simulated.

196

197 It should be noted that  $R_m$  is dependent on several factors including climate, soil,  
198 vegetation, topographic relief. Spatial  $R_m$  patterns are difficult to quantify precisely.

199 Despite highly idealized, the  $R_m$  patterns used in this study are expected to help  
200 improve theoretical understanding of using the Vogel method in complex conditions.



201

202 **Figure 2** Demonstrative sinusoidal variation in modelled groundwater recharge ( $R_m$ )  
203 with (a) the change in the amplitude A, and (b) the change in the wavelength B.

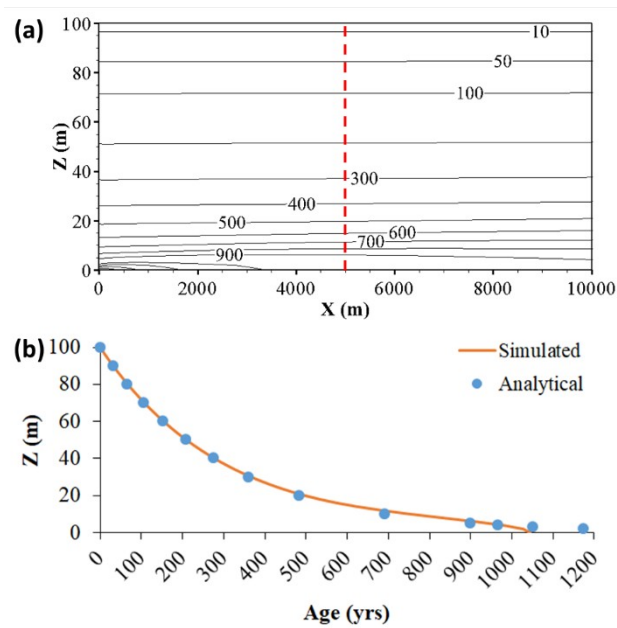
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205 **3. Results**

206 **3.1 Model validation**

207 Model validation was performed by examining the age distribution with constant  
208 recharge. The result shows that simulated groundwater age is constant horizontally for

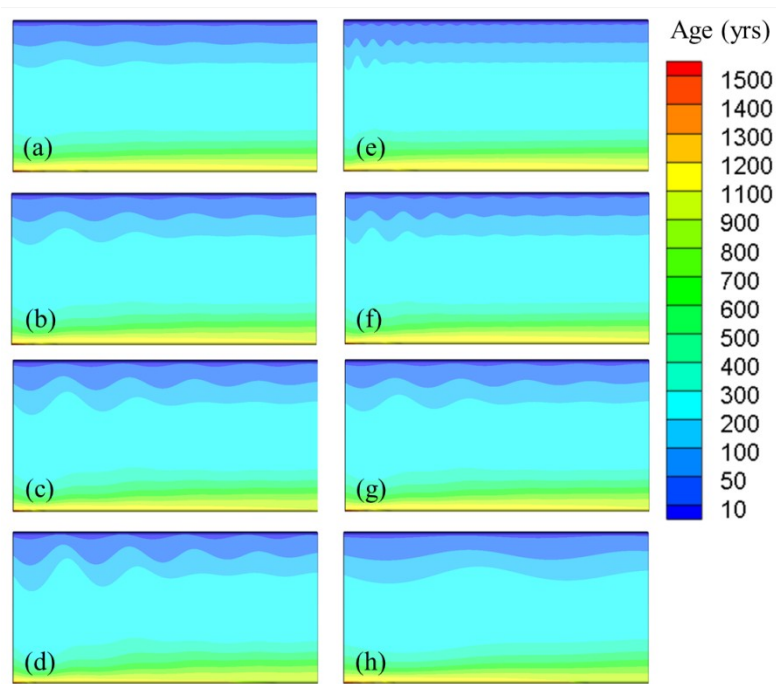
most part of the aquifer domain (approximately the range of  $Z = 20$  to  $Z = 100$  m) (Figure 3a). Close to the bottom ( $Z = 0$  to  $10$  m), the simulated age was slightly greater on the left than that on the right due to bottom boundary effect. Figure 3b shows an excellent match between the simulated and the analytical age profiles at  $X = 5000$  m. It can be seen that under ideal conditions the groundwater age increases with depth exponentially. The comparison suggests that our numerical model closely matches the analytical solution and therefore can be used to simulate more complex groundwater flow and age simulations.



**Figure 3** Numerical modeling results: (a) simulated groundwater age distribution with constant actual groundwater recharge at 100 mm/y for the entire top surface; (b) comparison of simulated to analytical groundwater age profiles at 5000 m as indicated with a red dashed line in (a). The analytical groundwater age was calculated by rearranging Equation (1) to obtain  $t$ .

### 3.2 Spatial distributions in simulated age and estimated recharge

We first simulated groundwater age with a specific  $R_m$  recharge case along the top surface ( $A = 25$  mm/y and  $B = 2000$  m, red solid curves in Figure 2). The spatial variation in the specific  $R_m$  causes groundwater age to vary with depth and also with distance further downgradient (Figure 4b). In general, large recharge forces groundwater to flow deeper than small recharge. As a result, areas with larger recharge tend to have younger groundwater at the same depth. The variation in the groundwater age weakens with depth because of the effect of the bottom boundary. This negative correlation between  $R_m$  and the groundwater age attenuates with distance downgradient due to the cumulative effect of the spatial recharge.



**Figure 4** Selected distribution of simulated groundwater age with  $R_m$  in sinusoidal patterns shown in Equation 2: (a) – (d): specified  $R_m$  parameters with same  $B$  at 2000

239 mm/y, but A at 12.5, 25, 35 and 50 mm/y, respectively; (e) – (h): specified  $R_m$   
240 parameters with same A at 25 mm/y, but B at 500, 1000, 2000 and 5000 mm/y,  
241 respectively. (g) is the same as (b).

242

243 The variability in the groundwater age distribution correlates well with the variability  
244 in  $R_m$ . As the amplitude of the sinusoidal recharge increases, the groundwater age  
245 fluctuates more strongly at the same location on the left side of the aquifer domain  
246 (Figure 4a to Figure 4d). As groundwater age is controlled by groundwater  
247 hydrodynamics, the age fluctuation weakens laterally (the amplitudes of the age  
248 isolines decrease from left to right). The change in the age distribution is more evident  
249 with the increase in B (Figure 4e to Figure 4g). Larger wavelength results in weaker  
250 variation in the age on the left of the domain but stronger variation on the right.

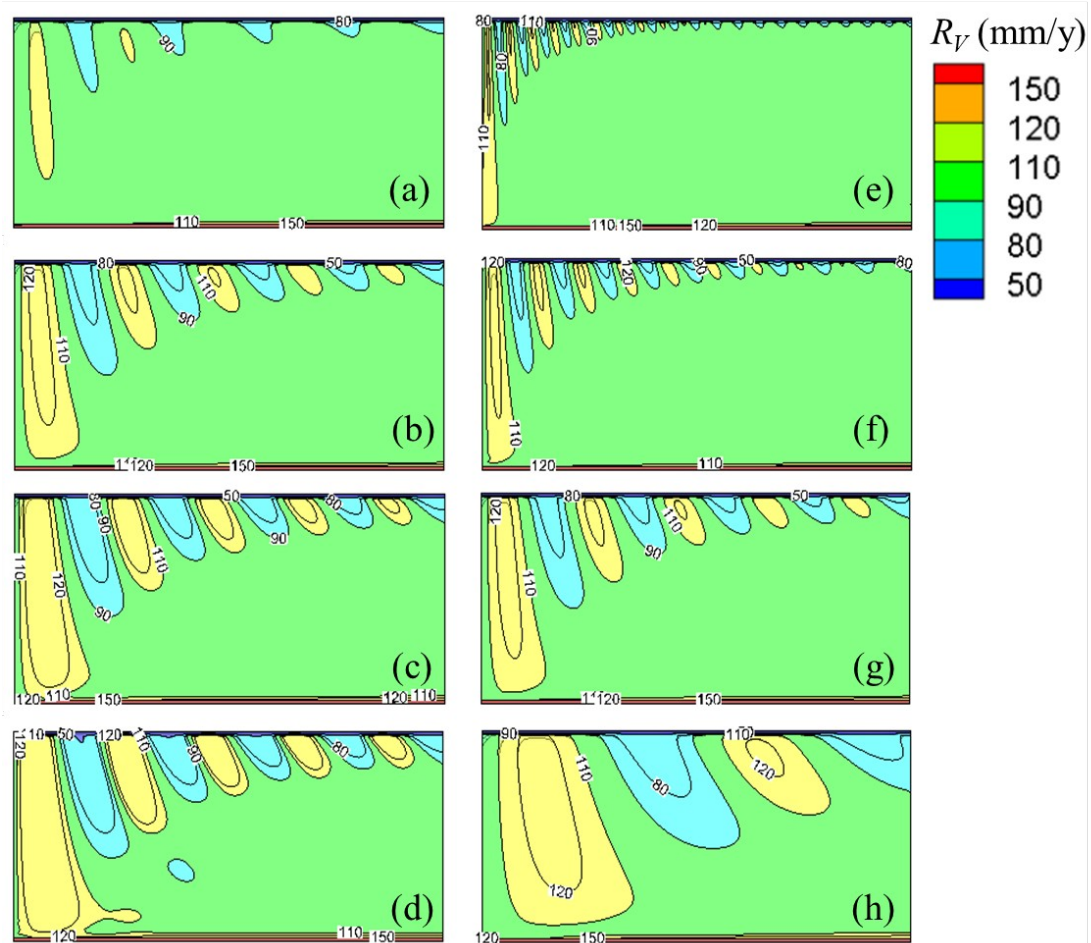
251

252 We then calculated  $R_v$  with the analytical method shown in Equation (1). At the lateral  
253 direction, the estimated recharge is strongly variable close to the land surface and  
254 becomes relatively stable deeper in the aquifer (Figure 5). Most of the estimated  
255 recharge falls in the range of 90–110 mm/y. The recharge close to the left boundary is  
256 more likely to deviate from the mean recharge (100 mm/y).

257

258 The variation in the groundwater recharge amplitude and wavelength strongly affects  
259 the groundwater recharge estimation. It can be seen that  $R_v$  at a certain depth (e.g.,  $Z$   
260 = 25 m, close to the bottom) becomes more variable as the amplitude increases. The

wavelength of  $R_V$  also changes in response to the change in the groundwater  
wavelength. The probability of overestimating or underestimating recharge rates  
increases with both the amplitude and the wavelength increase.



**Figure 5** Selected  $R_V$  distribution using simulated groundwater age (shown in Figure 4) and Equation 1: (a) – (d): specified  $R_m$  parameters with same B at 2000 mm/y, but A at 12.5, 25, 35 and 50 mm/y, respectively; (e) – (h): specified  $R_m$  parameters with same A at 25 mm/y, but B at 500, 1000, 2000 and 5000 mm/y, respectively. (g) is the same as (b). It should be emphasized that the  $R_V$  distribution indicates potential groundwater recharge for the top surface.

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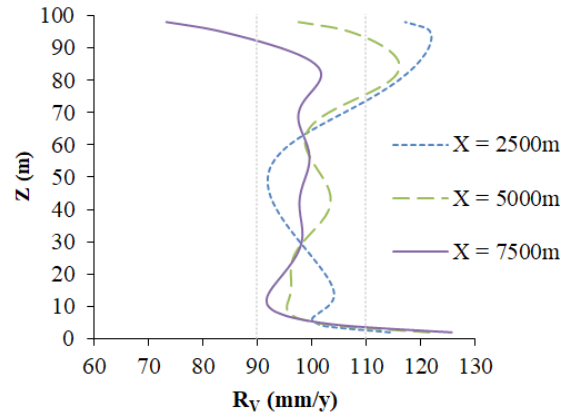
273 **3.3 Estimating long-term average recharge**

274 As shown in Figure 6,  $R_V$  varies with depth. At the shallow part of the aquifer,  $R_V$   
275 could strongly deviate from the mean  $R_m$ . But this  $R_V$  would be more representative of  
276 local recharge because of limited mixing of local water with regional groundwater  
277 passing by. Depending on the location of a sampling well,  $R_V$  near the water table  
278 could be lower or greater than the mean  $R_V$ .

279

280 In addition, Figure 6 suggests that  $R_V$  would generally be greater than the mean  $R_m$   
281 closer to the bottom boundary regardless of locations. This is because the bottom  
282 boundary restricted groundwater from flowing deeper down and so horizontal flow  
283 component becomes greater closer to the bottom boundary. This boundary effect  
284 results in younger groundwater age and therefore larger groundwater recharge.  
285 Despite different degrees of variability in  $R_V$  at different sampling well locations, All  
286 the mean  $R_V$  values estimated from these three locations fall in the accepted mean  
287 recharge range (103.9, 103.2 and 97.1 mm/y for  $X = 2500, 5000$  and  $7500$  m,  
288 respectively).

289



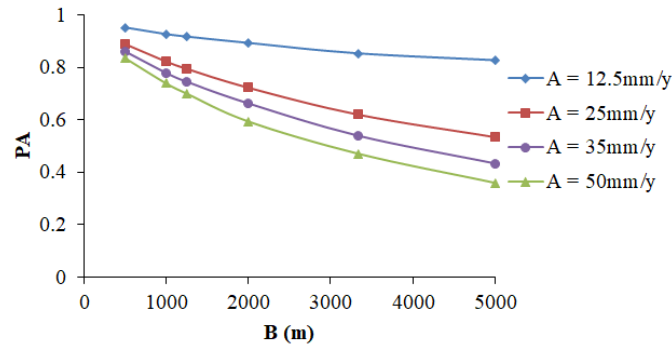
**Figure 6** Comparison of estimated recharge for the case with  $A = 25$  mm/y and  $B = 2000$  m at individual sampling well locations. Numbers shown on the legend indicate the distance from the model left boundary. Grey lines show the upper and lower bounds of the acceptable recharge range.

Recharge estimates are known to be uncertain. If 10% error is assumed in the mean recharge in our case, then  $R_V$  will be reasonable if it falls in the range of 90–110 mm/y (see grey bounds in Figure 6). We can obtain reasonable recharge estimates from all the sites if sampling is conducted within a reasonable sampling depth far from both the water table and the bottom boundary (e.g.,  $Z = 25$ –75 m in our case).  $R_V$  still fluctuates vertically but all the values are located between the uncertainty bounds of the actual recharge. Of course,  $R_V$  variability becomes weaker further downgradient if the sampling activity occurs at downstream locations (Compare the profile of  $X = 7500$  m to the other two profiles in Figure 6).

We can calculate the percentage of area (PA) that may yield acceptable recharge



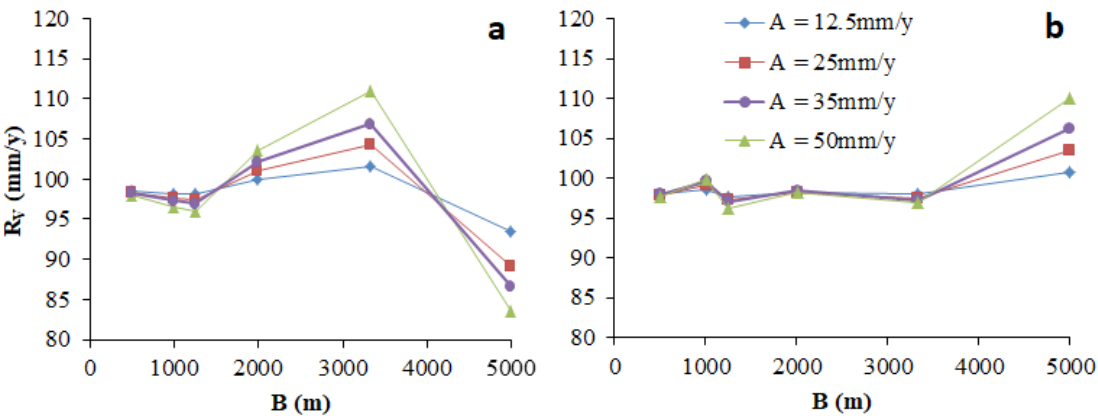
between 90 and 110 mm/y (green area in Figure 5). As seen in Figure 7, as high as 95% of the aquifer domain can provide the recharge in the acceptable range. PA generally decreases with the increase in A and also B. When A is small (12.5 mm/y), the estimated recharge does not change significantly with the increase in B (the lowest PA is 89% for A at 5000 m). However, dramatic change can be observed for the other A values. Only 36% of the aquifer domain can result in the recharge within the reasonable range for the worst scenario.



**Figure 7** Variation in the percentage of area (PA) with wavelength (B) for different amplitudes (A). Each PA was calculated by deriving the area with mean recharge between 90 and 110 mm/y first and then dividing this area by the domain area. A higher PA indicates a greater chance of obtaining true mean recharge.

Several depths can be sampled to obtain multiple recharge rates to compute mean  $R_V$ . The mean  $R_V$  in Figure 6 would be 101.0, 102.1 and 98.0 mm/y if we used samples from three depths at  $Z = 25, 50$  and  $75$  m. We computed the mean  $R_V$  for all the other cases. At the location of  $X = 5000$  m (Figure 8a), most mean  $R_V$  values are located between 90 and 110 mm/y, except when B is 5000 m. All the mean  $R_V$  values are

within the acceptable range at the location of  $X = 7500$  m (Figure 8b). In comparison to Figure 7, there is no clear trend in the mean  $R_V$  with the change in A or B. This is because groundwater age patterns are highly dependent on the forcing at the top boundary. A fixed sampling site could be located at either low or high values of the boundary forcing. Overall, it is more likely to obtain reasonable mean  $R_V$  if we use multiple depths and place a sampling well at a downstream site.

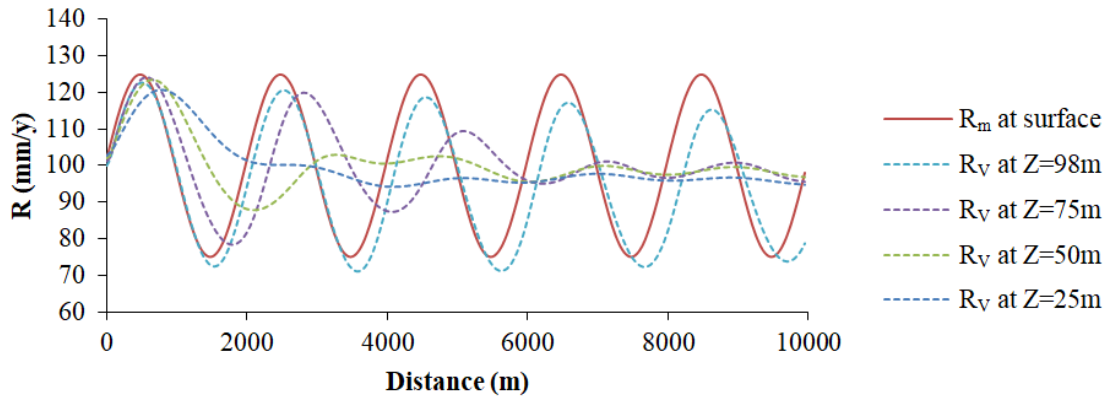


**Figure 8** Mean recharge at the depth range of 25–75 m at (a)  $X = 5000$  m; (b)  $X = 7500$  m

### 3.4 Estimating spatially-varying recharge

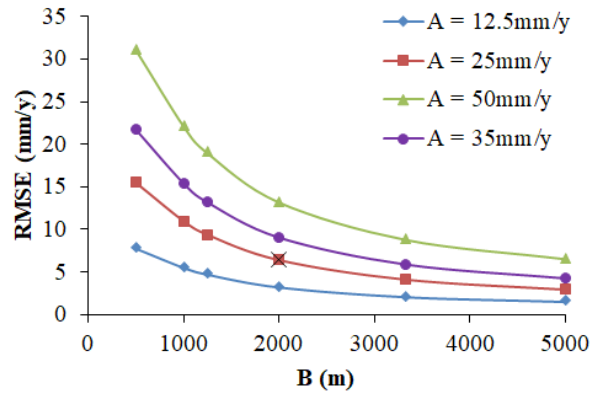
Figure 9 compares the recharge estimated from the Vogel method for different depths. As  $R_m$  varies in a sinusoidal manner,  $R_V$  also varies periodically with distance but its amplitude diminishes with the increase in the number of recharge cycles. It can be seen that the Vogel method slightly underestimate the recharge pattern when using the groundwater age at 98 m from the bottom (2 m beneath the water table) but overall the performance is reasonable in obtaining spatial  $R_V$ . When the groundwater age at 25

to 75 m is used, the Vogel method works better for estimating spatial  $R_V$  on the left-hand and mean  $R_V$  on the right-hand side of the model.



**Figure 9** Comparison of estimated recharge ( $R_V$ ) to modeled recharge ( $R_m$ ) for the case with  $A = 25$  mm/y and  $B = 2000$  m at different depths. Numbers shown on the legend indicate the distance from the model bottom. Root mean squared error (RMSE) between  $R_V$  and  $R_m$  is 18.2, 18.1, 17.9 and 8.1 mm/y for 25 m, 50 m, 75 m and 98 m from the bottom, respectively.

The spatial variation in  $R_V$  may be estimated reasonably for small magnitudes and large wavelengths, particularly for large wavelengths ( $B = 5000$  m in Figure 10). This condition generally occurs in fluvial plains with less variation in terrain topography. The ability of the method to estimate spatially-varying  $R_V$  becomes worse as the variable  $A$  increases and  $B$  decreases. This is mostly because larger variation in  $R_m$  causes stronger hydrodynamic mixing. Hence, it is difficult to estimate spatial  $R_V$  in mountainous regions with the Vogel method.



**Figure 10** Root mean squared error (RMSE) between spatial  $R_m$  and spatial  $R_V$  using groundwater age at  $Z = 98$  m (2 m below the water table) under different variability conditions. The black cross represents RMSE between the curves of  $R_m$  and  $R_V$  at  $Z = 98$  m in Figure 9.

It needs to be noted that we also conducted other scenarios with different mean recharge values (150 mm/y and 50 mm/y). The modeling shows similar trends in the changes of age,  $R_V$  distributions and RMSE between spatial  $R_m$  and spatial  $R_V$  presented above due to similar variation in  $R_m$  (not shown).

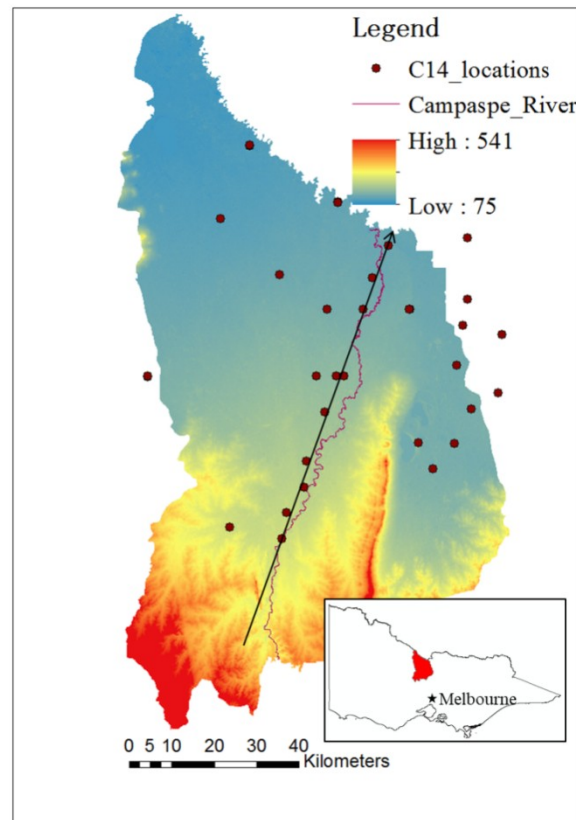
## 4. Field example

### 4.1 Field site and data description

Lower Campaspe catchment in southeast Australia was employed to demonstrate the variation in recharge rates (Figure 11). The entire catchment covers an area of 7,949 km<sup>2</sup> and forms part of the Murray-Darling Basin. The southern edge of the catchment is the Great Dividing Range (not shown in the map) and the lower part belongs to the

377 floodplains of the Campaspe and Murray rivers. Soils in this area are generally  
 378 Sodosols and Vodosols with varying thicknesses according to the Australian Soil  
 379 Classification system. The aquifers are composed of a shallow unconfined aquifer  
 380 with interbedded sand and clay and a deep semi-confined aquifer consisting of coarse  
 381 sand. The water table is usually 10–15 m below the land surface. Long-term annual  
 382 mean rainfall is around 400 mm/y in the north, whereas long-term annual mean  
 383 potential evapotranspiration is 1700 mm/y.

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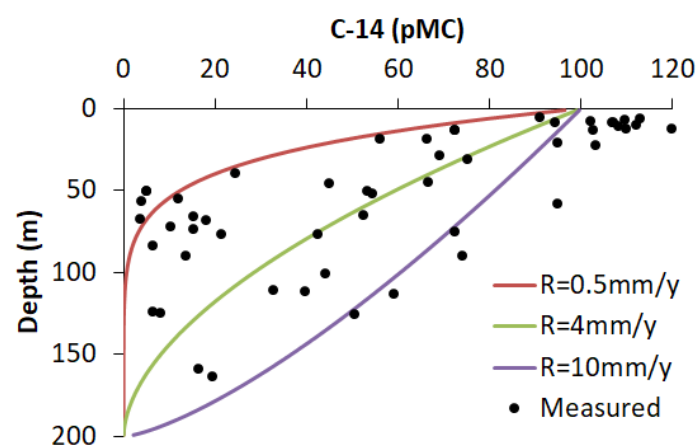


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386 **Figure 11** Location of the study area and groundwater  $^{14}\text{C}$  sampling sites. The black  
 387 arrow indicates the cross-section where C-14 data were examined.

388

Dozens of Carbon-14 ( $^{14}\text{C}$ ) samples were collected over the past years, including those published in previous studies (Cartwright, 2010) and those collected afterwards but not published. These data scattered in the lower Campaspe catchment, mostly along the river or close to the northern boundary. The screen depths of the bores where  $^{14}\text{C}$  samples were taken were lower than 20 m below the ground surface and the screen lengths were smaller than 10% of the screen depth.  $^{14}\text{C}$  was used to understand groundwater residence times and therefore help conceptualize the groundwater flow system which is critical for groundwater resource management. As shown in Figure 12, some bores have  $^{14}\text{C}$  activities greater than 100 pMC. This indicates that the groundwater around these bores was recharged after nuclear weapon tests in the 1950s to 1960s. However, this recharge process occurs usually in the shallow part of the aquifer, mostly likely in areas that have shallow water table.  $^{14}\text{C}$  values in most bores are usually lower than 100 pMC. There is a general negative correlation between  $^{14}\text{C}$  activity and groundwater depth.



**Figure 12** Relationship between  $^{14}\text{C}$  and groundwater depth. Black dots are measured  $^{14}\text{C}$ , whereas curves show ideal relationship for different recharge rates calculated

406 with Equation 1 and the  $^{14}\text{C}$  decay formula.

407

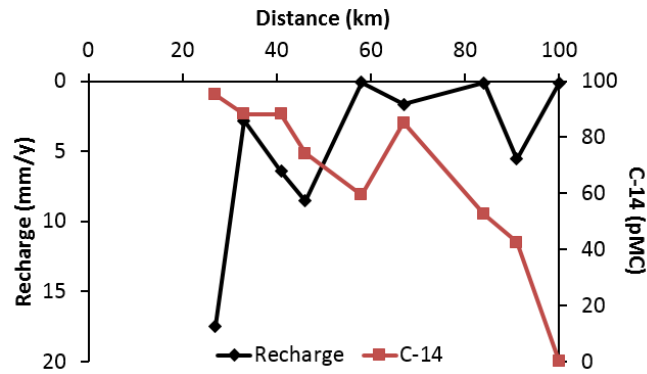
## 408 **4.2 Groundwater recharge estimates**

409 All the data could be used to estimate long-term average recharge over the entire  
410 catchment. We first assumed certain recharge rates and back calculated groundwater  
411 age with Equation (1). We then inferred  $^{14}\text{C}$  activities from the radioactive decay  
412 formula where the half life is 5730 years. The theoretical relationship between  $^{14}\text{C}$  and  
413 depth is shown in Figure 12 for different recharge scenarios. We can see that the  
414 average recharge may range from 0.5 mm/y to 10 mm/y. This recharge represents  
415 long-term average recharge over the entire area. Of course, some data are still outside  
416 of the  $^{14}\text{C}$ -depth curves for lower and upper recharge bounds. These data may indicate  
417 smaller or larger recharge rates in some parts of the catchment.

418

419 Individual  $^{14}\text{C}$  data can be used to estimate recharge spatially. It can be seen that  
420 recharge rates are generally larger in the south of the catchment but become lower  
421 towards the north (Figure 13). According to our theoretical analysis in Section 3, the  
422 data in the northern part are more representative of larger-scale recharge rates. Hence,  
423 if we have only limited data, it is best to make use of data located in the downgradient  
424 area. This finding is consistent with the recharge bounds identified from all the data  
425 above.

426



**Figure 13** Groundwater  $^{14}\text{C}$  data and recharge rates versus distance from the designated starting point along a transect shown in Figure 11.

Our theoretical analysis also indicates that combined use of tracers at different depths could be used to obtain recharge representative of larger areas. It is not very common to have  $^{14}\text{C}$  data at different depths in one bore. This is also the case in our field site. However, if we could combine tracer data close to the downstream (e.g., the three sites close to the 100 km), we would obtain the average recharge rate of 1.9 mm/y. This recharge estimate is still within the range identified in Figure 12. However, according to our theoretical analysis, this recharge estimate may be more representative of true recharge rates, although it is still very subjective.

## 5. Discussion

This study examined the validity of the popular Vogel method to quantify mean groundwater recharge and the possibility of using this method to estimate recharge reliably. Several numerical experiments were performed to simulate groundwater age distribution under different boundary conditions and one field example was provided.



445

## 446 **5.1 Estimation of long-term average groundwater recharge**

447 Our study showed that groundwater age determined from certain part of an aquifer  
448 can be used to infer long-term average groundwater recharge through the Vogel  
449 method (Vogel, 1967) under spatially-varying conditions. This ability of the Vogel  
450 method to estimate long-term average recharge is because age tracers are largely  
451 driven by advection and dispersion which is related to hydraulic forcing at the top  
452 surface (Sanford, 2010). Of course, as actual recharge is usually spatially varying  
453 (McMahon, Plummer, Böhlke, Shapiro, & Hinkle, 2011), estimated recharge will be  
454 closer to the actual long-term average value if samples are collected from a  
455 downgradient monitoring well and at a relatively large depth.

456

457 Our study indicates that long-term average groundwater recharge could be estimated  
458 reasonably if groundwater age samples are collected properly. However, it is often  
459 hardly possible to ascertain whether groundwater wells are placed in proper places, as  
460 groundwater age is strongly influenced by groundwater hydrodynamics and  
461 dispersion effects (McMahon, Plummer, Böhlke, Shapiro, & Hinkle, 2011). Therefore,  
462 any samples could be biased from the actual recharge like many existing studies (e.g.,  
463 Hinkle, Böhlke, Duff, Morgan, & Weick, 2007; Harrington, Cook, & Herczeg, 2002;  
464 Hagedorn, El-Kadi, Mair, Whittier, & Ha, 2011; McMahon, Plummer, Böhlke,  
465 Shapiro, & Hinkle, 2011). Hence, combining recharge estimates from multiple depths  
466 would yield more reliable long-term average recharge estimates (Harrington, Cook, &

Herczeg, 2002). This is particularly true if our sampling sites are located upgradient where estimated recharge may vary strongly with depth (see Figure 6). In reality, many groundwater wells are determined beforehand and probably with long screens for water supply purposes. These wells could also be utilized for estimating long-term average recharge if the entire screens are located within the unconfined aquifer, because groundwater from different depths will mix completely during sampling.

## **5.2 Spatial groundwater recharge estimation**

Actual groundwater recharge is known to vary in space and time. Many efforts have been made to derive spatial groundwater recharge (e.g., Crosbie, McCallum, Walker, & Chiew, 2010; Harrington, Cook, & Herczeg, 2002; Keese, Scanlon, & Reedy, 2005; Nolan, Baehr, & Kauffman, 2003; Xie, Crosbie, Simmons, Cook, & Zhang, 2019). Most methods make use of unsaturated zones as they are easily accessible and cost is relatively low (Keese, Scanlon, & Reedy, 2005; Nolan, Baehr, & Kauffman, 2003; Xie, Crosbie, Simmons, Cook, & Zhang, 2019). While they reflect on the variation in spatial groundwater recharge, the results are in fact potential groundwater recharge. Whether potential recharge is equal to actual recharge is dependent on a number of factors including thickness of unsaturated zones, degree of soil heterogeneity, and moisture content of underlying soil layers. Our study proposes that the Vogel method can also be used to estimate spatially varying actual recharge directly without the need to consider the factors mentioned above. What is required is to reliably determine

groundwater age close to the water table. Although this method may be theoretically better than the soil water balance method, it is much more expensive to obtain a sufficient number of groundwater age samples for the same purpose. Therefore, combinations of field methods including groundwater age method, soil water balance, chloride mass balance and tritium profile method are encouraged (e.g., Healy, 2010; Scanlon, Healy, & Cook, 2002).

This study assumed that groundwater age at the water table is zero. This assumption is appropriate in humid regions where water table is shallow. Infiltrating water can reach the water table within a short time. However, this assumption is not always valid in arid and semiarid regions. In thick arid unsaturated zones it can take hundreds and even up to thousands of years for infiltrating water to reach the water table (Cook, Edmunds, & Gaye, 1992; Love et al., 2013; Wood, Cook, & Harrington, 2015). Therefore, under these conditions when estimating recharge in arid and semiarid regions, groundwater age above the water table must also be known in priori.

### **5.3 Limitations**

Our study suggests that both long-term average and spatially varying groundwater recharge could be estimated with the Vogel method under spatially varying conditions. Our study represents a further step towards better quantifying groundwater recharge with existing methods. We built our study on the simple homogeneous and isotropic aquifer conceptualization to examine the potential impact of complex boundary

conditions on groundwater recharge estimation. We are acutely aware that aquifer heterogeneity may significantly affect flow fields and therefore recharge estimation as shown by Kozuskanich, Simmons, & Cook (2014). Given that numerous existing studies assume homogeneous settings (e.g., Hinkle, Böhlke, Duff, Morgan, & Weick, 2007; Harrington, Cook, & Herczeg, 2002; Hagedorn, El-Kadi, Mair, Whittier, & Ha, 2011) and our main objective was not on assessing the impact of aquifer heterogeneity, we based our model on those existing studies with homogeneous conceptualization.

In addition, actual groundwater recharge is variant both in space and time. As the Vogel method assumes a steady state setting, this study did not consider temporal changes in recharge. Temporal variation in recharge are best examined through numerical modeling which has been frequently studied (e.g., Xie, Crosbie, Simmons, Cook, & Zhang, 2019). Dispersive mixing is another factor that may contribute to the complexity in estimating recharge with this method, but it is beyond the scope of our study. Greater dispersivity causes stronger mixing in water particles and decay isotopes and likely results in a younger age and a greater recharge rate. This impact of dispersion needs to be studied in a future study.

## 6. Conclusions

This study performed several numerical experiments in order to examine the validity

of the simplified analytical method to estimate spatial recharge. We first generated different flow fields and age patterns with different spatially-varying fluxes as model upper boundary conditions. We then estimated groundwater recharge using the analytical method at the spatial scale and examined the differences between analytical and modelled recharge rates. A field example was provided in the end to illustrate our results as much as possible. Several remarks can be made based on this study.

Long-term average groundwater recharge can be estimated reasonably well provided that the age sample is collected from the middle of an aquifer and at downstream areas. Multiple groundwater age measurements can be averaged to improve the precision of the long-term average groundwater recharge. This simple analytical method can also be used to estimate local groundwater recharge if age samples are collected below the water table.

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## 553 **Conflicts of Interest**

554 The authors declare no conflict of interest.

555

## 556 **Data Availability**

557 Data used in this study can be provided by the corresponding author upon request.

558

## 559 **References**

560 Aquanty Inc. (2018). HydroGeoSphere. A three-dimensional numerical model  
561 describing fully-integrated subsurface and surface flow and solute transport.

562 Waterloo, Ontario, Canada. <https://www.aquanty.com/hgs-download>

563 Allison, G. B., Gee, G. W., & Tyler, S. W. (1994). Vadose-zone techniques for  
564 estimating groundwater recharge in arid and semiarid regions. *Soil Science*

565 *Society of America Journal*, 58, 6-14. [https://doi.org/](https://doi.org/10.2136/sssaj1994.03615995005800010002x)

566 10.2136/sssaj1994.03615995005800010002x

567 Cartwright, I. (2010). Using groundwater geochemistry and environmental isotopes to  
568 assess the correction of  $^{14}\text{C}$  ages in a silicate-dominated aquifer system. *Journal*

569 *of Hydrology*, 382, 174–187. <https://doi.org/10.1016/j.jhydrol.2009.12.032>

570 Chesnaux, R., Molson, J. W., & Chapuis, R. P. (2005). An analytical solution for  
571 ground water transit time through unconfined aquifers. *Ground Water*, 43, 511-

572 517. <https://doi.org/10.1111/j.1745-6584.2005.0056.x>

573 Chesnaux, R., Santoni, S., Garel, E., & Huneau, F. (2018). An analytical method for

574 assessing recharge using groundwater travel time in Dupuit flow aquifers.

- 575 *Ground Water*, 56, 986-992. <https://doi.org/10.1111/gwat.12794>
- 576 Cook, P. G., & Böhlke, J. K. (2000). Determining timescales for groundwater flow  
577 and solute transport. In P. G. Cook & A. L. Herczeg (Eds.), *Environmental*  
578 *tracers in subsurface hydrology* (pp. 1-30). Boston: Kluwer Academic Publisher.
- 579 Cook, P. G., Edmunds, W. M., & Gaye, C. B. (1992). Estimating paleorecharge and  
580 paleoclimate from unsaturated zone profiles. *Water Resources Research*, 28,  
581 2721-2731. <https://doi.org/10.1029/92WR01298>
- 582 Cook, P. G., Walk, G. R., & Jolly, I. D. (1989). Spatial variability of groundwater  
583 recharge in a semiarid region. *Journal of Hydrology*, 111, 195-212.  
584 [https://doi.org/10.1016/0022-1694\(89\)90260-6](https://doi.org/10.1016/0022-1694(89)90260-6)
- 585 Crosbie, R. S., McCallum, J. L., Walker, G. R., & Chiew, F. H. S. (2010). Modelling  
586 climate-change impacts on groundwater recharge in the Murray-Darling Basin,  
587 Australia. *Hydrogeology Journal*, 18, 1639-1656. [https://doi.org/](https://doi.org/10.1007/s10040-010-0625-x)  
588 [10.1007/s10040-010-0625-x](https://doi.org/10.1007/s10040-010-0625-x)
- 589 de Vries, J. J., & Simmers, I. (2002). Groundwater recharge: An overview of  
590 processes and challenges. *Hydrogeology Journal*, 10, 5-17. [https://doi.org/](https://doi.org/10.1007/s10040-001-0171-7)  
591 [10.1007/s10040-001-0171-7](https://doi.org/10.1007/s10040-001-0171-7)
- 592 Goode, D. J. (1996). Direct simulation of groundwater age. *Water Resources*  
593 *Research*, 32, 289-296. <https://doi.org/10.1029/95WR03401>
- 594 Healy, R. W. (2010). *Estimating Groundwater Recharge*. Cambridge University Press.
- 595 Hinkle, S. R., Böhlke, J. K., Duff, J. H., Morgan, D. S., & Weick, R. J. (2007).  
596 Aquifer-scale controls on the distribution of nitrate and ammonium in ground

- 597 water near La Pine, Oregon, USA. *Journal of Hydrology*, 333, 486-503.  
 598 <https://doi.org/10.1016/j.jhydrol.2006.09.013>
- 599 Harrington, G. A., Cook, P. G., & Herczeg, A. L. (2002). Spatial and temporal  
 600 variability of ground water recharge in central Australia: A tracer approach.  
 601 *Ground Water*, 40, 518-527. <https://doi.org/10.1111/j.1745-6584.2002.tb02536.x>
- 602 Hagedorn, B., El-Kadi, A. I., Mair, A., Whittier, R. B., & Ha, K. (2011). Estimating  
 603 recharge in fractured aquifers of a temperate humid to semiarid volcanic island  
 604 (Jeju, Korea) from water table fluctuations, and Cl, CFC-12 and  $^3\text{H}$  chemistry.  
 605 *Journal of Hydrology*, 409, 650-662. [https://doi.org/](https://doi.org/10.1016/j.jhydrol.2011.08.060)  
 606 [10.1016/j.jhydrol.2011.08.060](https://doi.org/10.1016/j.jhydrol.2011.08.060)
- 607 Kim, J. H., & Jackson, R. B. (2012). A global analysis of groundwater recharge for  
 608 vegetation, climate, and soils. *Vadose Zone Journal*, 11. [https://doi.org/](https://doi.org/10.2136/vzj2011.0021RA)  
 609 [10.2136/vzj2011.0021RA](https://doi.org/10.2136/vzj2011.0021RA)
- 610 Kozuskanich, J., Simmons, C. T., & Cook, P. G. (2014). Estimating recharge rate from  
 611 groundwater age using a simplified analytical approach: Applicability and error  
 612 estimation in heterogeneous porous media. *Journal of Hydrology*, 511, 290-294.  
 613 <https://doi.org/10.1016/j.jhydrol.2014.01.058>
- 614 Keese, K. E., Scanlon, B. R., & Reedy, R. C. (2005). Assessing controls on diffuse  
 615 groundwater recharge using unsaturated flow modeling. *Water Resources*  
 616 *Research*, 41, W06010. <https://doi.org/10.1029/2004WR003841>
- 617 Love, A. J., Shand, P., Karlstrom, K., Crossey, L., Rousseau-Gueutin, P., Priestley,  
 618 S., ... Keppel, M. (2013). Geochemistry and travertine dating provide new



- insights into the hydrogeology of the Great Artesian Basin, South Australia. *Procedia Earth and Planetary Science*, 7, 521-524. <https://doi.org/10.1016/j.proeps.2013.03.065>
- McMahon, P. B., Bohlke, J. K., & Carney, C. P. (2007). Vertical gradients in water chemistry and age in the northern High Plains Aquifer, Nebraska, 2003. *US Geol. Surv. Sci. Invest. Rep.* (pp. 2006-5294).
- McMahon, P. B., Plummer, L. N., Böhlke, J. K., Shapiro, S. D., & Hinkle, S. R. (2011). A comparison of recharge rates in aquifers of the United States based on groundwater-age data. *Hydrogeology Journal*, 19, 779-800. <https://doi.org/10.1007/s10040-011-0722-5>
- Nolan, B. T., Baehr, A. L., & Kauffman, L. J. (2003). Spatial variability of groundwater recharge and its effect on shallow groundwater quality in southern New Jersey. *Vadose Zone Journal*, 2, 677-691.
- Priestley, S., Shand, P., Love, A., Crossey, L., & Karlstrom, K. (2013). Groundwater recharge, hydrodynamics and hydrochemistry of the western Great Artesian Basin. In A. J. Love, D. Wohling, S. Fulton, P. Rousseau-Gueutin, & S. De Ritter (Eds.), *Allocating water and maintaining springs in the Great Artesian Basin* (pp. 171-201). Canberra, Australia: Natl Water Community.
- Sanford, W. (2002). Recharge and groundwater models: An overview. *Hydrogeology Journal*, 10, 110-120. <https://doi.org/10.1007/s10040-001-0173-5>
- Sanford, W. (2010). Coastal flow. *Nature Geoscience*, 3, 671-672. <https://doi.org/10.1038/ngeo958>

- 641 Scanlon, B. R., Healy, R. W., & Cook, P. G. (2002). Choosing appropriate techniques  
 642 for quantifying groundwater recharge. *Hydrogeology Journal*, 10, 18-39.  
 643 <https://doi.org/10.1007/s10040-011-0722-5>  
 644 Scanlon, B. R., Keese, K. E., Flint, A. L., Flint, L. E., Gaye, C. B., Edmunds, W. M.,  
 645 & Simmers, I. (2006). Global synthesis of groundwater recharge in semiarid and  
 646 arid regions. *Hydrological Processes*, 20, 3335-3370.  
 647 <https://doi.org/10.1002/hyp.6335>
- 648 Vogel, J. C. (1967). Investigation of groundwater flow with radiocarbon. In *Isotopes*  
 649 *in Hydrology* (pp. 355-369). Vienna: IAEA.
- 650 Wood, C., Cook, P. G., & Harrington, G. A. (2015). Vertical carbon-14 profiles for  
 651 resolving spatial variability in recharge in arid environments. *Journal of*  
 652 *Hydrology*, 520, 134-142. <https://doi.org/10.1016/j.jhydrol.2014.11.044>
- 653 Xie, Y., Crosbie, R., Simmons, C. T., Cook, P. G., & Zhang, L. (2019). Uncertainty  
 654 assessment of spatial-scale groundwater recharge estimated from unsaturated  
 655 flow modelling. *Hydrogeology Journal*, 27, 379-393.  
 656 <https://doi.org/10.1007/s10040-018-1840-0>
- 657 Xie, Y., Cook, P. G., Shanafield, M., Simmons, C. T., & Zheng, C. (2016). Uncertainty  
 658 of natural tracer methods for quantifying river-aquifer interaction in a large river.  
 659 *Journal of Hydrology*, 535, 135-147. [https://doi.org/](https://doi.org/10.1016/j.jhydrol.2016.01.071)  
 660 [10.1016/j.jhydrol.2016.01.071](https://doi.org/10.1016/j.jhydrol.2016.01.071)
- 661 Xie, Y., Crosbie, R., Yang, J., Wu, J., & Wang, W. (2018). Usefulness of soil moisture  
 662 and actual evapotranspiration data for constraining potential groundwater

663 recharge in semiarid regions. *Water Resources Research*, 54, 4929-4945.  
664 <https://doi.org/10.1029/2018WR023257>  
665 Yang, Y., Wang, Z., Xie, Y., Ataie-Ashtiani, B., Simmons, C. T., Luo, Q., ... Wu, J.  
666 (2019). Impacts of groundwater depth on regional scale soil gleyization under  
667 changing climate in the Poyang Lake Basin, China. *Journal of Hydrology*, 568,  
668 501-516. <https://doi.org/10.1016/j.jhydrol.2018.11.006>