

The iDiv Ecotron - a flexible research platform for multitrophic biodiversity research

Anja Schmidt^{a,b,c}, Jes Hines^{b,c}, Manfred Türke^{b,c}, François Buscot^{a,b}, Martin Schädler^{a,b},
Alexandra Weigelt^{c,b}, Alban Gebler^{b,c}, Stefan Klotz^a, Tao Liud^d, Sascha Reth^e, Stefan
Trogisch^{f,b}, Jacques Roy^g, Christian Wirth^{b,c}, Nico Eisenhauer^{b,c}

Institutes

^aHelmholtz Centre for Environmental Research - UFZ

^bGerman Centre for Integrative Biodiversity Research (iDiv) Halle-Jena-Leipzig

^cLeipzig University

^dKey Laboratory of Vegetation Restoration and Management of Degraded Ecosystems,
South China Botanical Garden, Chinese Academy of Sciences, Guangzhou 510650, China

^eUmwelt-Geräte-Technik GmbH - UGT

^fMartin Luther University Halle-Wittenberg

^gFrench National Centre for Scientific Research - CNRS

Address corresponding author:

Dr. Anja Schmidt
Department of Community Ecology
Helmholtz Centre for Environmental Research - UFZ
Theodor-Lieser-Str. 4
06120 Halle (Saale)
Germany

29 Abstract

30 Across the globe, ecological communities are confronted with multiple global environmental
31 change drivers, and they are responding in complex ways ranging from behavioural,
32 physiological, and morphological changes within populations to changes in community
33 composition and food web structure with consequences for ecosystem functioning. A better
34 understanding of global change-induced alterations of multitrophic biodiversity and the
35 ecosystem-level responses in terrestrial ecosystems requires holistic and integrative
36 experimental approaches to manipulate and study complex communities and processes
37 above and below the ground. We argue that mesocosm experiments fill a critical gap in this
38 context, especially when based on ecological theory and coupled with microcosm
39 experiments, field experiments, and observational studies of macroecological patterns. We
40 describe the design and specifications of a novel terrestrial mesocosm facility, the iDiv
41 Ecotron. It was developed to allow the setup and maintenance of complex communities and
42 the manipulation of several abiotic factors in a near-natural way, while simultaneously
43 measuring multiple ecosystem functions. To demonstrate the capabilities of the facility, we
44 provide a case study. This study shows that changes in aboveground multitrophic interactions
45 caused by decreased predator densities can have cascading effects on the composition of
46 belowground communities. The iDiv Ecotrons technical features, which allow for the assembly
47 of an endless spectrum of ecosystem components, create the opportunity for collaboration
48 among researchers with an equally broad spectrum of expertise. In the last part, we outline
49 some of such components that will be implemented in future ecological experiments to be
50 realized in the iDiv Ecotron.

51 **Key words:** *food webs, biodiversity and ecosystem functioning, mesocosms, biotic*
52 *interactions, lysimeters, climate chambers*

53

54 Introduction

55 Ecosystems are threatened by a multitude of environmental change drivers (Pereira et al.,
56 2012, Murphy and Romanuk, 2014, Newbold et al., 2015, Maxwell et al., 2016, Díaz et al.,
57 2019). Over the last few decades, there has been an explosion of studies examining changes
58 in ecological communities and environmental conditions (Hines et al., 2019, Liu et al., 2011,
59 Stork and Astrin, 2014). The desire to draw generalizable conclusions from these studies led
60 to a period of synthesis, during which information from individual studies was compiled
61 allowing for quantitative evaluation of the variation in ecological changes across systems
62 (Gurevitch et al., 1992, Hillebrand et al., 2020, Halpern et al., 2020). Such comprehensive and
63 quantitative synthesis studies enabled researchers to identify generalizable patterns in
64 biodiversity (Calatayud et al., 2020), trends in biodiversity change (Dornelas et al., 2014,
65 Blowes et al., 2019), and relationships between biodiversity and ecosystem functioning (e.g.
66 Cardinale et al., 2012, Gessner et al., 2010, Lefcheck et al., 2015, Soliveres et al., 2016).
67 These high-impact synthesis studies can also serve as a roadmap for designing future
68 experiments, as they help to identify important knowledge gaps which need to be filled in order
69 to better understand the functioning of ecosystems and predict the consequences of climate
70 change.

71 We have limited empirical evidence for at least three key aspects of environmental changes
72 in ecosystems and communities that draw a roadmap for future research. First, there are
73 limited numbers of ecosystem response variables that have been consistently studied across
74 systems. For example, the most commonly reported response variables are primary
75 production and decomposition (Cardinale et al., 2006, Schmidt et al., 2015a, Schmidt et al.,
76 2015b). However, the few existing multitrophic biodiversity studies indicate that the
77 interactions of higher trophic levels may be particularly important for multiple ecosystem
78 functions (Naeem et al., 1994, Lefcheck et al., 2015, Hines et al., 2015b, Soliveres et al.,
79 2016), and that especially these species might be very vulnerable to environmental changes

(Hines et al., 2015a, Voigt et al., 2003). Second, studies tend to investigate limited types of mechanisms and processes underlying changes in biodiversity, ecosystem functioning, and the relationship between the two (Hillebrand et al., 2020). That is, while there is strong emphasis on the effects of global change drivers on changes in species richness (Tilman and Downing, 1994, Harpole et al., 2016, Seabloom et al., 2021, but see Dornales, 2014; Vellend et al., 2013), there is less known about the ecosystem consequences of changes in behaviour (Wilson et al., 2020, Cordero-Rivera, 2017) and community composition (Spaak et al., 2017, Hillebrand et al., 2018) of species that persist in communities. Third, although ecosystems are confronted with complex cocktails of global change drivers (Bowler et al., 2020), so far only a limited number of their types and combinations have been studied in realistic experiments (Rineau et al., 2019, Rillig et al., 2019, but see Schädler et al., 2019, Korell et al., 2020). Especially with regard to climate change, understanding interactions between different environmental variables such as temperature and precipitation, land-use or biodiversity on ecosystem functioning is essential to make predictions for future ecosystem developments and the potential consequences for society (Roy et al., 2017). To address our current knowledge gaps, we need experiments which can simultaneously manipulate and measure different global change drivers (Vanderkelen et al., 2020) and investigate their impacts on a wide range of functional groups and trophic levels of organisms (De Boeck et al., 2020, Komatsu et al., 2019, Korell et al., 2020). Combining such “meta-scale” studies with laboratory and field studies, especially large-scale climate change experiments (like Schädler et al., 2019), provides the opportunity to understand the complex patterns of biodiversity-ecosystem function relationships and their responses to environmental changes as well as the underlying processes that operate across organizational levels of life (cell-individual-population-community-ecosystem; Ferlian et al. (2018)).

Here, we introduce the iDiv Ecotron platform (iDiv stands for the German Centre for Integrative Biodiversity Research Halle-Jena-Leipzig in Germany). This platform is a highly flexible experimental infrastructure that was specifically designed to perform multitrophic biodiversity

experiments in terrestrial ecosystems (Eisenhauer and Türke, 2018). In the following sections, we describe the iDiv Ecotron specifications and functioning, we highlight a case study experiment as an application possibility, and we provide an outlook on the potential contributions of future ecotron experiments. The concept of the iDiv Ecotron was to create a facility which allows the setup and maintenance of complex communities and manipulation of several abiotic factors in a near-natural way, while simultaneously measuring multiple ecosystem functions. Environmental conditions, such as humidity, nutrient supply, light, and precipitation, can be fully controlled and monitored (for details see App.1), which allows the iDiv Ecotron to be used for the simulation of multiple abiotic scenarios together with scenarios of above-belowground community change. The iDiv Ecotron offers the possibility to study a wide range of ecosystem responses, including above-belowground interactions of plants, microbes, and invertebrates. The platform can accommodate stand-alone experiments, and also provides complementary information to small- and large-scale experiments (lab-ecotron-field). Therefore the iDiv Ecotron links investigations at multiple experimental and spatial scales, and serves as a key component for collaborations between researchers from different disciplines to conduct interdisciplinary studies on the drivers of, and relationship between, biodiversity and ecosystem functioning. Consequently, this platform is likely to provide novel insights into ecosystem responses to global change.

Setup & Design of the iDiv Ecotron

Based on some first facilities that were built in England (Imperial College ecotron in Silwood Park; Lawton, 1996) and the USA (Desert Institute Ecocells in Reno, Nevada, Griffin et al., 1996) in the 1990's, highly sophisticated experimental infrastructures, so-called 'ecotrons', started to get established world-wide at the beginning of the 21st century, reflecting the urgent need for such infrastructures accompanied by the rapid evolution in digital technology and electronics (e.g. Ecotron in Montpellier, France, Milcu et al., 2014; ExpoSCREEN in Munich, Germany; Ecotron in Hasselt, Belgium; Biotron in Lincoln, New Zealand). A review with detailed descriptions and comparisons of a variety of ecotrons worldwide can be found in Roy et al. (2020). The idea behind an ecotron is to combine the precision, specificity, and complete control of single independent and response variables of laboratory experiments and the realism and large-scale community- and environment-related aspects of field studies. Roy et al. (2020) define an ecotron as an "...experimental facility comprising a set of replicated enclosures designed to host ecosystems samples, enabling realistic simulation of above- and belowground environmental conditions, while simultaneously and automatically measuring ecosystem processes. Therefore, ecotrons provide continuous information on ecosystem functioning (fluxes of energy and matter)".

The iDiv Ecotron is located in a climate-controlled and blacked out hall on an area of 485 m² at the research station of the Helmholtz Centre for Environmental Research - UFZ in Bad Lauchstädt (Saxony-Anhalt, 51° 22' 60N, 11° 50' 60E, 118 m a.s.l.), Germany. The indoor research facility houses 24 identical experimental units (hereafter EcoUnits, see Fig. 1), each of which can contain one to four ecosystems, separated above- or belowground, or both. In this way, up to 96 sub-units with various biotic and abiotic variables to be manipulated and measured independently can be set up. These features are new in the world of ecotrons: the iDiv Ecotron, as an indoor facility, is completely independent of external weather conditions (unlike, for example, the Hasselt Ecotron in Belgium) and at the same time provides a large

number of individual independent chambers (unlike other indoor facilities, such as ExpoSCREEN in Munich, Germany, or the Montpellier Ecotron mesocosms in France; see Roy et al. 2020). The iDiv Ecotron concept was developed in cooperation with numerous scientists and technicians from iDiv, including strong participation by the UFZ, national and international collaborators, and the companies 'EMC - Gesellschaft zur Erfassung und Bewertung von Umweltdaten mbH', and 'Umwelt-Geräte-Technik GmbH (UGT), Müncheberg'.

EcoUnits are experimental chambers with the outer dimensions of 1.55 m × 1.55 m × 3.20 m (L × W × H), comprising a lower part, which can be filled with soil (belowground part), an upper part (aboveground part), and a technical section on the top. The frame of the chamber is constructed of aluminum construction profiles providing stability and flexibility.

The belowground part contains a container with internal dimensions of 1.24 m × 1.24 m × 0.80 m (L × W × H) made of welded PE-HD and a steel bottom. It can be filled with up to 1.23 m³ of soil, or alternatively equipped with four steel cylinders (lysimeters) measuring 0.50 m × 0.80 m (D × H), each of which can hold 0.16 m³ of soil. The container as well as the lysimeters feature pluggable openings in three different depths (9.5 cm, 21.5 cm, and 43.5 cm), where sensors for soil temperature, soil moisture, and water potential can be inserted. Additional larger openings in the same depths as those for the sensors offer the opportunity to install minirhizotrons (acrylic glass tubes) for horizontal monitoring of root development using a portable root scanner (see Möller et al., 2019).

Besides manually filling the lysimeters with soil, they can be used to excavate intact soil monoliths, including aboveground vegetation, directly from the field. This enables precise investigations of almost undisturbed soil systems, preserving their structure and stratification as well as their faunal and microbial soil communities. Both the lysimeters and the containers provide a living space of sufficient size to establish and study belowground organisms and processes. To achieve a near-natural soil temperature gradient with temperature decreasing from the surface to deeper soil depths, the bottom of the soil container was fitted with a coil

that circulates a cooling medium. This system can be regulated individually for each EcoUnit and automated with the data from the above- and belowground temperature sensors.

To allow pore water sampling and near natural drainage of water from the soil system, four suction systems are installed at the bottom of the soil container or one in each lysimeter. Each suction system consists of a suction cup ring with 8 suction cups, a pump, a control module, and two glass bottles. By applying negative pressure (max. -60 kPa), the suction systems continuously extract and collect pore water. When one bottle is filled, the control unit of each suction system automatically switches to the alternate bottle and empties the first one. To quantify the volume of water sampled, the system counts the number of bottle changes. This enables a continuous supply of soil water for chemical analyses and an automated recording of the total amount of collected water. Simultaneously, the negative pressure applied at the bottom of the lysimeter lowers the water potential from there up and reduces “unnatural” high plant transpiration. When the soil column is cut over the course of the monolith extraction, the water potential at the cut level becomes zero - it is brought to atmospheric pressure, which eases and therefore increases the extraction of water by plants. Here, the suction system can be used to apply the pressure that corresponds to the natural in situ water potential at that depth. This allows for these ecosystems to further approximate natural conditions (Groh et al. 2016). Optionally, single suction cups can also be installed in three different depths (9.5 cm, 21.5 cm, and 43.5 cm) by using the pluggable openings.

The aboveground part, with internal dimensions of 1.46 m × 1.46 m × 1.50 m (L × W × H), provides sufficient space for communities of large herbs or tree saplings (see Fig. 2) including their complex multitrophic interaction networks. In each quarter, a video camera can be installed (for details on the camera system see App.1), e.g. for monitoring vegetation development over time (Ulrich et al., 2020) or insect behavior, such as movement patterns, flower visitation of pollinators, and habitat use. By using infrared lights, the cameras can also operate in darkness.

204 The aboveground part is further equipped with an irrigation system consisting of a flow meter
205 and four electromagnetic valves with fixed nozzles. By sequentially processing the opening
206 times of the valves, each quarter of an EcoUnit can be automatically provided with individual
207 volumes of water at programmable times. All irrigation systems are supplied with deionized
208 water from a central reverse osmosis system. To compensate for the flow resistance caused
209 by different lengths of supply hoses to each EcoUnit, the water pressure at the water treatment
210 plant is increased to approx. 4 bar (400 kPa) and then reduced to a constant level of about 2
211 bar (200 kPa).

212 Ambient air temperature is maintained centrally in the Ecotron hall, but the air flow rate of each
213 sub-unit can be regulated individually. Climatic conditions are recorded by combined humidity
214 and temperature sensors installed in each quarter of an EcoUnit, usually placed at a height of
215 40 cm above soil surface. Conditions are continuously compared with those of the hall and,
216 as needed, automatically adjusted by increasing or decreasing the fan speed of the ventilation
217 system. All four quarters of the EcoUnit can be regulated individually.

218 Further, the top part of the EcoUnits is equipped with a diffuser holding 4 LED lamps adjustable
219 in color and intensity. The light system provides three individually dimmable color channels
220 (400 nm - 405 nm, 460 nm - 475 nm, 625 nm - 720 nm) as well as a dimmable white channel
221 (5000 K + 3000 K), and a binary (ON/OFF) infrared channel (840 nm - 850 nm). For the overall
222 luminance as well as for each color channel, the intensity can be set from 0% to 100%
223 individually, determining the general light color. This can be done either manually or
224 automated in an hourly resolution with an automatically linear transition between the settings.
225 In this way, the relative proportion of different wavelengths within the light spectrum can be
226 modified (e.g. a higher proportion of red light at dawn and dusk). The maximum photosynthetic
227 active radiation (PAR) 5 cm above the standard soil surface can reach about $400 \mu\text{mol s}^{-1} \text{m}^{-2}$
228 on average (detailed information on the heterogeneity of illumination can be found in App.2).
229 Two electrical cabinets provide the power supply for the lamps and a local control unit for all
230 sensors and actuators.

231 Control commands and settings of all manipulable environmental parameters are stored in a
232 central database and get transmitted to each EcoUnit via a network. In turn, the execution
233 confirmations as well as the timestamped sensor data of each EcoUnit are logged in the same
234 database. This asynchronous communication between EcoUnits and database server
235 provides a high operational reliability and independence of network's capacity bottlenecks. A
236 simple graphical user interface eases the handling of database entries.

237

Case Study - *Effects of aboveground predators on aboveground-belowground interactions and ecosystem functions*

Rationale

Aboveground-belowground interactions are known to determine the functioning of terrestrial ecosystems (Scheu, 2001, Wardle et al., 2004). Previous work has shown that aboveground invertebrate predators can induce trophic cascades that “trickle-down” to affect soil food webs and a broad range of ecosystem functions (Wardle et al., 2005). Here we present a case study conducted in the iDiv Ecotron to test how plant community composition may affect such trickle-down effects. Further, as plant-mediated effects of aboveground predators may additionally depend on the activity of soil ecosystem engineers, which structure the environment for (Brown, 1995, Eisenhauer, 2010) and the resource supply of soil food webs (Eisenhauer, 2010, Schwarzmuller et al., 2015), we investigated the effects of soil fauna on multitrophic diversity and ecosystem functions. The unique functionality of the iDiv Ecotron enabled us to study potential cascading effects of aboveground predators on herbivores, plants, and soil food webs, and how these effects are modulated by decomposer communities in the soil. Specifically we tested (1) if the target plant biomass would be lower in the presence of herbivores, an effect that would be alleviated by the presence and higher density of predators (e.g. Wardle et al., 2005). We further hypothesized (2) that the identity of the neighboring plant community will affect the biomass of the target plant with biomass being higher in a community with herb species compared to grass species due to elevated competition for soil resources in the presence of grasses (Eisenhauer and Scheu, 2008). Moreover, we expected (3) the presence of decomposers (earthworms and Collembola) to affect the tritrophic interactions aboveground, as decomposition and mineralization processes in soil can significantly alter the performance of the target plant (Scheu, 2003, van Groenigen et al., 2014) as well as the competition with the surrounding vegetation (Eisenhauer and Scheu, 2008, Sabais et al.,

2012). Finally, we hypothesized (4) that there will be trickle-down effects of aboveground predators on soil nematode density and species richness due to altered resource supply and that soil food web responses to these trickle-down effects will be modulated by earthworm presence as they significantly change the structure of the environment for and resource supply of other soil organisms (Brown, 1995, Eisenhauer, 2010).

Methods

In six EcoUnits in a lysimeter configuration a tritrophic system got established comprising a target plant (*Vicia faba* L.), its host-specific aphid (*Acyrtosiphon pisum* Harris), and a predator exclusively feeding on aphids (*Coccinella septempunctata* Linnaeus; details on initial densities can be found in (App. A.1). We further included a soil fauna treatment (with and without soil fauna) to test, if predator effects are modulated by the presence of macro- and meso-decomposers in the soil; and a “plant neighbor” treatment to test plant responses in different competitive environments and to increase variation for reproducibility purposes (Milcu et al., 2018). Concisely, we established an experimental setup with three treatment factors comprising *aboveground invertebrates*, *belowground invertebrates* and *surrounding vegetation* (see Fig. 3). Each treatment combination was replicated three times. While soil compartments were all fully isolated one from another (four per EcoUnit), the aboveground compartments allowed for an exchange of invertebrates between lysimeter pairs with an acrylic glass barrier of 15 cm height preventing the migration of soil invertebrates between lysimeters. In this way, there were two independent experimental units in each of the six EcoUnits resulting in twelve independent units and 24 sub-units in total (more details on the experimental setup and environmental conditions can be found in App. A.2).

Details on treatment factors:

(1) *aboveground invertebrates*; the treatment was established to test if predator effects depend on their density (**4 levels**: all aboveground invertebrates absent [*Control*], only aboveground herbivores present [*Herbivores only*], aboveground herbivores present with aboveground predators in low density [*Coccinella low*], aboveground herbivores present with aboveground predators in high density [*Coccinella high*]).

(2) *belowground invertebrates*; to half of the lysimeters earthworms and Collembola were added to test if predator performance is modulated by the presence of macro- and meso-decomposers in the soil (**2 levels**: earthworms and Collembola present [*with soil fauna*] versus earthworms and Collembola absent [*no soil fauna*]); soil invertebrate species list and initial densities can be found in App. A.3).

(3) *surrounding vegetation*; the focal plants (*Vicia faba* L) were each surrounded by a herb or grass monoculture (**4 levels**: *Bellis perennis* L., *Centaurea jacea* L., *Festuca pratensis* Huds., *Holcus lanatus* L.; details on plants can be found in App. A.4).

Analyzed response variables were dry weight (g) of the focal plant (*Vicia faba*), nematode density, nematode species richness (all three recorded during the harvest at the end of the experiment), maximum numbers of aphids (peak number of individuals counted in one assessment during the experiment), days of aphid infestation (number of days beans were infested with aphids; details can be found in App. A.5). The experiment ran for 124 days, from 03 February 2017 to 06 June 2017.

Results

The target plant (for brevity ‘bean’ in the following) dry weight differed significantly depending on the neighboring plant species ($F_{3, 48} = 5.16$, $P < 0.01$; Fig. 4, App.8) and the aboveground

invertebrate treatments ($F_{3, 48} = 6.48$, $P < 0.001$; Fig. 4, App.8), whereas it did not differ among belowground invertebrate treatments as well as with any of the two- or three-way interactions of the three variables tested. Bean dry weight was lowest in patches with *B. perennis* and *H. lanatus*, whereas it was significantly higher in *C. jacea* patches (Fig. 4). Furthermore, bean dry weight was highest in the aboveground invertebrate 'Control' and the 'Coccinella high' treatments, whereas it was lowest in the 'Herbivores only' treatment.

The maximum number of aphids and number of days of aphid infestation differed significantly between the aboveground invertebrate treatments ($F_{1, 24} = 8.24$, $P = 0.01$; Fig. 4, App.8; and $F_{3, 48} = 63.19$, $P < 0.001$, respectively; App.8). Further, the maximum number of aphids showed significant differences in the interaction between plant neighbor species and belowground invertebrates ($F_{3, 24} = 5.82$, $P = < 0.01$; Fig. 4, App.8). In general, numbers of aphids were higher in the 'Coccinella low' treatment compared to the 'Coccinella high' treatment. Depending on the plant neighbor identity, maximum number of aphids slightly decreased (*B. perennis* and *F. pratensis*) or increased (*C. jacea* and *H. lanatus*) with the presence of belowground invertebrates, but effects were not statistically significant.

Nematode densities differed significantly only between plant neighbor species ($F_{3, 48} = 2.86$, $P = 0.05$; App.8). Highest numbers were found in patches where *C. jacea* was planted and lowest numbers in plots with *F. pratensis* (significant differences were found only between these two). For nematode species richness, only the interaction between plant neighbor species and the aboveground invertebrate treatment was significant ($F_{9, 48} = 2.21$, $P = 0.04$; Fig. 4, App.8). Although the post-hoc Tukey's HSD test showed no significant differences between factor levels, nematode species richness was lowest in the 'Herbivores only' treatment in the presence of *F. pratensis*, while it was highest in the 'Control' treatment in the presence of *C. jacea*.

336 Discussion

337 In contrast to our expectations, beans did not generally benefit from growing in herb
338 communities, while being suppressed by more dominant nitrophilous grasses (Eisenhauer and
339 Scheu, 2008). We observed opposing effects for the two grass species and for the two herb
340 species on bean biomass. Among the four neighboring plant species, *H. lanatus* produced by
341 far the highest amount of aboveground plant biomass (139.5 g) at the end of the experiment
342 compared to the other three species (*F. pratensis*: 92.1 g, *C. jacea*: 51.1 g, *B. perennis*: 5.3
343 g), and, as graminoid species typically produce a dense and large root system, we speculate
344 that also root biomass was highest (not assessed in this study). Thus, both enhanced
345 aboveground light competition and belowground competition for resources may have
346 contributed to an overall advantage in resource acquisition over the bean, causing low bean
347 biomass. Indeed, it has been often confirmed that grasses are stronger competitors compared
348 to herbaceous species (Tilman, 1982, Del-Val and Crawley, 2005). Moreover, another
349 potential explanation for the patterns found in our study may be that in patches of low biomass,
350 e.g., in *B. perennis* patches, the habitat structure for predators was comparably low leading to
351 a migration to more favorable habitat structures. This effect may have cascaded to lower
352 trophic levels increasing abundances of herbivores and decreasing plant performance
353 (Romero and Koricheva, 2011). The importance of such non-trophic interactions based on
354 habitat structure has been often highlighted (Kalinkat et al., 2013, Majdi et al., 2014).

355 Our results confirm the often found tritrophic relationships between predators, herbivores, and
356 primary producers, where predators, in our case ladybirds, exert a top-down control on aphid
357 abundances which, in turn, have a top-down effect on the bean (Romero and Koricheva,
358 2011). Surprisingly, the effects of plant neighbor species on aphid abundances were opposing
359 for communities without and with belowground invertebrates. These findings highlight the
360 significance of aboveground-belowground interactions and show that decomposers can
361 influence aboveground multitrophic interactions by altering the competition between plants

(Wardle et al., 2004). Moreover, we found that trickle-down effects of aboveground invertebrates on soil food webs (here represented by soil nematode species richness) depend on plant community composition. This finding suggests that the competitive environment of a focal plant can alter its effects on soil community composition, potentially through changes in the amount and quality of plant-derived resources entering the soil (Hooper et al., 2000).

Taken together, our study shows distinct interaction effects between aboveground and belowground invertebrate communities on multitrophic interactions and community composition in the sub-compartments. These changes are likely to alter how communities function, which may have subsequent feedback effects on nutrient cycling and community composition. The results of our study highlight the need for infrastructures that allow to manipulate food webs of high complexity, which can hardly be realized experimentally under field or simplified laboratory conditions, and at the same time, taking advantage of measuring and controlling a large fraction of other non-targeted parameters including environmental conditions.

377 Outlook

378 Over the last several decades ecologists have written thousands of papers about changes in
379 climate and biological communities. Yet, some important knowledge gaps remain. Here we
380 discuss the relevance of mesocosm research as an underappreciated scale of inquiry. We
381 further develop this line of reasoning by describing three opportunities where the iDiv Ecotron
382 is particularly well suited to address challenges limiting an integrative understanding of
383 biodiversity and ecosystem functioning.

384 Mesoecology is an important and often overlooked scale in environmental change research
385 (Stewart et al., 2013). While macroecological studies provide more realistic abiotic and biotic
386 context for investigating ecosystem processes, complex communities and environmental
387 conditions can only be controlled, and causality of patterns inferred, to a very limited extent,
388 and often with very few replicates (Lawton et al., 1993, Eisenhauer and Türke, 2018). On the
389 other hand, laboratory microcosm studies can fully control and alter external factors and allow
390 for high replication (Benton et al., 2007). However, laboratory studies are often limited to
391 investigating single mechanisms and processes under artificial and simplified environmental
392 conditions (Lawton et al., 1993). They are prone to experimental artefacts caused by the
393 simplification of complex interactions which may bias results and induce misleading
394 conclusions (Roy et al., 2020, Milcu et al., 2018, Carpenter, 1996, Carpenter, 1999, Schindler,
395 1998). The iDiv Ecotron provides an important middle ground, especially with the possibility of
396 extracting and implementing up to 96 intact soil monoliths which allows for precise
397 investigations of almost undisturbed soil systems, while preserving their structure and
398 stratification as well as their faunal and microbial soil communities. Mesocosm experiments
399 close the gap between small- and large-scale studies and they allow scientists working
400 together across levels of organization from cells to ecosystems to test basic and applied
401 ecological questions. However, attempts to do so will profit from including a few key aspects
402 of research that serve as future opportunities.

403 *Opportunity 1: Multi-trophic diversity change*

404 Although many studies have evaluated responses of plant species to environmental variation,
405 ecologists have yet to demonstrate the collective importance of these responses for the full
406 complement of plants' interaction partners above and below the ground. This is particularly
407 important because not all taxa that interact with plants perceive environmental variation at the
408 same scale (Veen et al., 2019, Heinen et al., 2018). Therefore, although it has been shown
409 that diversity can beget diversity, and patterns in plant diversity can parallel patterns of soil
410 diversity and aboveground consumer diversity (Scherber et al., 2010, Eisenhauer et al., 2013),
411 these patterns may be mismatched (Cameron et al., 2019) and/or further decoupled by
412 environmental change drivers (Bardgett and Wardle, 2010, Thakur, 2020). Future iDiv Ecotron
413 experiments will evaluate differences in spatial and temporal response to drivers that may
414 explain mismatches in above- and belowground biodiversity (Eisenhauer and Türke, 2018).
415 The iDiv Ecotron allows for simultaneous manipulation of aboveground and belowground
416 biodiversity, with particular emphasis on belowground sub-systems through the use of intact
417 soil cores, the examination of roots via rhizotrons, and large enough spatial scale to examine
418 differences in patterns of aboveground and belowground diversity. Rigorously testing factors
419 that influence aboveground-belowground relationships is critical, because they form key
420 pathways by which environmental variation influences community assembly, biodiversity
421 effects on ecosystem functioning, and the impacts of environmental change on community
422 dynamics. To develop effective plans to conserve biodiversity, we need meso-scale empirical
423 studies that test the mechanisms underlying effects of environmental drivers on aboveground-
424 belowground biodiversity and ecosystem functioning.

425 *Opportunity 2: Beyond presence/absence—Behavioural and chemical mechanisms of plants*
426 *and animal interactions*

427 Traditionally, experimental examinations of food web interactions have been conducted by
428 stocking simplified communities into microcosms or field plots and quantifying the outcome of

the interactions by counting the presence and abundance of species after a designated time period. It is likely that phenotypic changes (e.g. changes in behaviour, chemistry, or morphology) serve as precursors to the numerical changes in community composition that are typically quantified, or that phenotypic changes can drive major changes in ecosystem functioning on their own (Matthews et al., 2011, Turcotte and Levine, 2016). Yet, phenotypic responses are more often evaluated in highly simplified communities with limited emphasis on interaction complexity. We see considerable potential for iDiv Ecotron studies to extend highly simplified laboratory experiments showing effects of environmental drivers on phenotypic responses (e.g. behavioural, morphological, and physiological change). Changes in local foraging and behaviour/activity patterns may be an important mechanism underlying changes in biodiversity-ecosystem function relationships (Jeltsch et al., 2013). The iDiv Ecotron can be fit with a landscape of sensors for detecting movement of animals tagged with RFID chips. Repulsed (or aggregated) animal activity patterns can point to the importance of non-trophic and trait mediated interactions (e.g. fear). Such behavioural changes are not limited to animals. For example, behaviour changes of plants emission of plant volatiles can be turned off and on depending on plants interaction partners. Plant volatiles play key roles in plant defence against aboveground and belowground herbivores, plant competition, and plant communication (Pierik et al., 2014). Yet, research of plant volatiles is often conducted on isolated plants or pairs of plants. These aspects of phenotypic changes (animal movement, plant volatiles) are difficult to assess in field conditions where signals may be detected by ecological communities but not my scientific instruments due to difficulties relocating animals in larger more complex landscapes, or buffering effects of wind. Future iDiv Ecotron experiments will examine the role of aboveground-belowground plant and animal behaviour in complex communities.

Opportunity 3: Multiple drivers of environmental heterogeneity and environmental change

We have only begun to identify the full array of environmental changes confronting ecosystems today (Bowler et al., 2020). The iDiv Ecotron allows for independent manipulation

of several abiotic factors (e. g. precipitation, light, nutrients, and temperature) in gradient-based or factorial combinations. Non-additive, synergistic or unexpected responses may be detected from heretofore untested combinations of environmental change drivers. There is also much potential to use the iDiv Ecotron to examine the influence of minor or extreme levels of drivers and to detect non-linear relationships between drivers and ecosystem responses (Damgaard et al., 2018, De Boeck et al., 2015). Therefore, the iDiv Ecotron is an ideal tool to complement environmental change experiments where ecological responses are evaluated over longer time periods or greater spatial scales, but at the cost of examining a reduced number of scenarios (e.g. Schädler et al., 2019). Future studies may therefore be considered as a step towards precision and mechanistic understanding supplementing other laboratory or field studies.

In conclusion, the iDiv Ecotron provides a flexible collaborative research platform that operates at an intermediate scale, connecting simplistic microcosm experiments and real-world heterogeneity. Their size allows for evaluation of naturally complex aboveground-belowground interactions, often overlooked mechanisms (e.g. behaviour, plant volatiles), as well as a broad range of environmental drivers. Therefore, this robust experimental facility can help to fill several critical knowledge gaps identified in synthesis studies. The iDiv Ecotron will be used to assemble, disassemble, and reassemble ecological communities in rigorous tests of basic and applied ecological questions. We start with an empty box with strong technical capabilities to control environmental conditions, endless possible combinations of species, and an open call to potential collaborators: What would you do if you could rebuild the world?

Additional Information

App.1

Table App.1

App.2

Figure App.2

App.3 - Aboveground invertebrates

To test for effects of aboveground invertebrates on tritrophic interactions and nematode communities, we implemented different combinations of herbivore and predator species presence and absence. We used the pea aphid *Acyrtosiphon pisum* Harris as aboveground herbivore feeding specifically on the broad bean *Vicia faba* L. Eight mature individuals were added to each replicate of respective treatments between April 27th and May 3rd. We used adult beetles of the seven-spot ladybird (*Coccinella septempunctata* Linnaeus) as specialized aphid predators which were added in two different densities (two or four individuals) on May 10th to respective treatments. In total, we tested four aboveground treatments: Control (no invertebrates), Herbivores only (with aphids, without ladybirds), Coccinella low (with aphids, with two individuals of *C. septempunctata*), and Coccinella high (with aphids, with four individuals of *C. septempunctata*).

498 App.4 - Experimental setup

499 The 24 Lysimeters were filled with steam-sterilized top soil (purchased at Bauzentrum Farys
500 GmbH, Laucha). For sterilization, the soil was subjected to water steam at approx. 100°C for
501 30 minutes. Such sterilization leads to a heavy release of nutrients due to the death of soil
502 organisms (Alphei and Scheu, 1993, Trevors, 1996), which is why the soil was thoroughly
503 rinsed with tap water afterwards (Jager et al., 1970). The soil was inoculated with nematode
504 and microbial communities on February 2nd 2017, marking the start of the experiment. Live
505 soil organisms were extracted from top soil of an experimental grassland site (Jena
506 Experiment, Roscher et al., 2004). We added four independent samples of soil wash solution
507 (extracted from 100 g of soil each, filtered through a 125 µm sieve) to each lysimeter on
508 February 3rd. In addition, we added three independent inoculates of nematode solution
509 between February 2nd and March 10th, which were previously live-extracted from 20 g wet soil
510 each, following the modified Baermann funnel method (Cesarz et al., 2019; for details on
511 nematode communities in the Jena Experiment, see Eisenhauer, Migunova, Ackermann,
512 Ruess, & Scheu, 2011 and Cesarz et al., 2017). To exclude that unintended additions of
513 nematodes might have confounded the controlled inoculation, soil samples from the sterilized
514 soil filled into lysimeters were extracted with the same method and yielded no live nematodes.
515 The following environmental parameters were set in the EcoUnits: light/dark cycle 16/8 h (max
516 illumination at day, gradual change), temperature 21°C at day and 17°C at night (gradual
517 change over the course of 3 h), irrigation of 400 ml on each lysimeter area daily at 4 am, soil
518 temperature set to 17°C in 43.5 cm soil depth.

519

520 App.5 - Soil invertebrates

521 The steam-sterilized soil (Dietrich et al., 2020) got inoculated with microorganisms and
522 nematodes (see App.4). To test for interactions between the aboveground tritrophic system

and belowground invertebrate presence (meso-/macro-fauna), we added the following soil invertebrates to one of the two lysimeters in each replicate: 15 juvenile anecic earthworms (*Lumbricus terrestris* Linnaeus, mean weight 4.4 g) and 20 individuals each of two Collembola species (*Folsomia candida* Willem, *Protaphorura armata* Tullberg). Collembola populations have been shown to develop rapidly in the experimental soil until the carrying capacity of the system is reached (Eisenhauer et al., 2011). Fifty g of commercial grassland litter was provided as substrate to both lysimeters (bunny® Frischgras-Heu). Since hay is a natural product, its grain properties vary according to season. Typically, the in Table App.5 shown groups of plants are included.

Table App.5

App.6 - Plants

Nineteen 23-day old seedlings each of two herbaceous (*Bellis perennis* L., *Centaurea jacea* L.) and two grass species (*Festuca pratensis* Huds., *Holcus lanatus* L.) were transplanted in regular distances of 5 cm and within monoculture quarters (see App.6) into each lysimeter on February 16th to mimic a simplified grassland community. In the center of each monoculture quarter, a single individual of an 8-day old broad bean seedling (*Vicia faba* L., variety “Dreifach Weiße”, Bruno Nebelung GmbH) was transplanted on April 19th, representing the specific host plant of aboveground herbivores. Consequently, there were four host plant individuals per lysimeter and thus eight individuals per replicate.

App.7 - Measurements

Numbers of aphids on each bean were counted every 7 days. For analyses, we used the peak number of all assessments during the experiment (hereafter called ‘maximum number of

aphids'). Furthermore, we recorded the number of days beans were infested with aphids by counting live aphids on each bean individual from first discovery until last discovery; last discovery could either be the end of the experiment or the time a bean got in a bad status and was not a suitable host for aphids anymore.

All beans were harvested 49 days after their transplantation by taking 5-cm-diameter soil cores to a depth of 10 cm with beans in their center. The soil was sieved through a 2-mm sieve, bean roots were extracted and both were stored at 4°C until further processing. After the removal of aphids the bean aboveground parts were dried at 45°C for 3 days and weighed. Nematodes were extracted from the previous stored soil following a modified Baermann technique with an extraction time of 48 h (Cesarz et al., 2019). Extracted nematodes were transferred to formalin (4%) and counted to obtain the total density of nematodes. Subsequently, 100 individuals were randomly selected and identified to genus level following (Bongers and Bongers, 1998) or separated into morphospecies where not possible.

App.8 - GLMM

Table App.8

- ALPHEI, J. & SCHEU, S. 1993. Effects of biocidal treatments on biological and nutritional properties of a mull-structured woodland soil. *In*: BRUSSAARD, L. & KOOISTRA, M. J. (eds.) *Soil Structure/Soil Biota Interrelationships*. Amsterdam: Elsevier.
- BARDGETT, R. D. & WARDLE, D. A. 2010. *Aboveground-belowground linkages: biotic interactions, ecosystem processes, and global change*, Oxford University Press.
- BENTON, T. G., SOLAN, M., TRAVIS, J. M. & SAIT, S. M. 2007. Microcosm experiments can inform global ecological problems. *Trends Ecol Evol*, 22, 516-21.
- BLOWES, S. A., SUPP, S. R., ANTAO, L. H., BATES, A., BRUELHEIDE, H., CHASE, J. M., MOYES, F., MAGURRAN, A., MCGILL, B., MYERS-SMITH, I. H., WINTER, M., BJORKMAN, A. D., BOWLER, D. E., BYRNES, J. E. K., GONZALEZ, A., HINES, J., ISBELL, F., JONES, H. P., NAVARRO, L. M., THOMPSON, P. L., VELLEND, M., WALDOCK, C. & DORNELAS, M. 2019. The geography of biodiversity change in marine and terrestrial assemblages. *Science*, 366, 339-345.
- BONGERS, T. & BONGERS, M. 1998. Functional diversity of nematodes. *Applied Soil Ecology*, 10, 239-251.
- BOWLER, D. E., BJORKMAN, A. D., DORNELAS, M., MYERS-SMITH, I. H., NAVARRO, L. M., NIAMIR, A., SUPP, S. R., WALDOCK, C., WINTER, M., VELLEND, M., BLOWES, S. A., BÖHNING-GAESE, K., BRUELHEIDE, H., ELAHI, R., ANTÃO, L. H., HINES, J., ISBELL, F., JONES, H. P., MAGURRAN, A. E., CABRAL, J. S., BATES, A. E. & FISH, R. 2020. Mapping human pressures on biodiversity across the planet uncovers anthropogenic threat complexes. *People and Nature*, 2, 380-394.
- BROWN, G. G. 1995. How do earthworms affect microfloral and faunal community diversity? *Plant and Soil*, 170, 209-231.
- CALATAYUD, J., ANDIVIA, E., ESCUDERO, A., MELIAN, C. J., BERNARDO-MADRID, R., STOFFEL, M., APONTE, C., MEDINA, N. G., MOLINA-VENEGAS, R., ARNAN, X., ROSVALL, M., NEUMAN, M., NORIEGA, J. A., ALVES-MARTINS, F., DRAPER, I., LUZURIAGA, A., BALLESTEROS-CANOVAS, J. A., MORALES-MOLINO, C., FERRANDIS, P., HERRERO, A., PATARO, L., JUAN, L., CEA, A. & MADRIGAL-GONZALEZ, J. 2020. Positive associations among rare species and their persistence in ecological assemblages. *Nat Ecol Evol*, 4, 40-45.
- CAMERON, E. K., MARTINS, I. S., LAVELLE, P., MATHIEU, J., TEDERSOO, L., BAHRAM, M., GOTTSCHALL, F., GUERRA, C. A., HINES, J., PATOINE, G., SIEBERT, J., WINTER, M., CESARZ, S., FERLIAN, O., KREFT, H., LOVEJOY, T. E., MONTANARELLA, L., ORGIAZZI, A., PEREIRA, H. M., PHILLIPS, H. R. P., SETTELE, J., WALL, D. H. & EISENHAUER, N. 2019. Global mismatches in aboveground and belowground biodiversity. *Conserv Biol*, 33, 1187-1192.
- CARDINALE, B. J., DUFFY, J. E., GONZALEZ, A., HOOPER, D. U., PERRINGS, C., VENAIL, P., NARWANI, A., MACE, G. M., TILMAN, D., WARDLE, D. A., KINZIG, A. P., DAILY, G. C., LOREAU, M., GRACE, J. B., LARIGAUDERIE, A., SRIVASTAVA, D. S. & NAEEM, S. 2012. Biodiversity loss and its impact on humanity. *Nature*, 486, 59-67.
- CARDINALE, B. J., SRIVASTAVA, D. S., DUFFY, J. E., WRIGHT, J. P., DOWNING, A. L., SANKARAN, M. & JOUSEAU, C. 2006. Effects of biodiversity on the functioning of trophic groups and ecosystems. *Nature*, 443, 989-92.
- CARPENTER, S. R. 1996. Microcosm Experiments have Limited Relevance for Community and Ecosystem Ecology. *Ecology*, 77, 677-680.
- CARPENTER, S. R. 1999. Microcosm Experiments Have Limited Relevance for Community and Ecosystem Ecology: Reply. *Ecology*, 80, 1085-1088.
- CESARZ, S., CIOBANU, M., WRIGHT, A. J., EBELING, A., VOGEL, A., WEISSER, W. W. & EISENHAUER, N. 2017. Plant species richness sustains higher trophic levels of soil

- nematode communities after consecutive environmental perturbations. *Oecologia*, 184, 715-728.
- CESARZ, S., SCHULZ, A. E., BEUGNON, R. & EISENHAUER, N. 2019. Testing soil nematode extraction efficiency using different variations of the Baermann-funnel method. *Soil Org*, 91, 61-72.
- CORDERO-RIVERA, A. 2017. Behavioral Diversity (Ethodiversity): A Neglected Level in the Study of Biodiversity. *Frontiers in Ecology and Evolution*, 5.
- DAMGAARD, C., HOLMSTRUP, M., SCHMIDT, I. K., BEIER, C. & LARSEN, K. S. 2018. On the problems of using linear models in ecological manipulation experiments: lessons learned from a climate experiment. *Ecosphere*, 9, e02322.
- DE BOECK, H. J., BLOOR, J. M. G., AERTS, R., BAHN, M., BEIER, C., EMMETT, B. A., ESTIARTE, M., GRUNZWEIG, J. M., HALBRITTER, A. H., HOLUB, P., JENTSCH, A., KLEM, K., KREYLING, J., KROEL-DULAY, G., LARSEN, K. S., MILCU, A., ROY, J., SIGURDSSON, B. D., SMITH, M. D., STERNBERG, M., VANDVIK, V., WOHLGEMUTH, T., NIJS, I. & KNAPP, A. K. 2020. Understanding ecosystems of the future will require more than realistic climate change experiments - A response to Korell et al. *Glob Chang Biol*, 26, e6-e7.
- DE BOECK, H. J., VICCA, S., ROY, J., NIJS, I., MILCU, A., KREYLING, J., JENTSCH, A., CHABBI, A., CAMPIOLI, M., CALLAGHAN, T., BEIERKUHNLEIN, C. & BEIER, C. 2015. Global Change Experiments: Challenges and Opportunities. *BioScience*, 65, 922-931.
- DEL-VAL, E. & CRAWLEY, M. J. 2005. What limits herb biomass in grasslands: competition or herbivory? *Oecologia*, 142, 202-211.
- DÍAZ, S., SETTELE, J., BRONDÍZIO, E. S., NGO, H. T., GUÈZE, M., AGARD, J., ARNETH, A., BALVANERA, P., BRAUMAN, K. & BUTCHART, S. H. 2019. Summary for policymakers of the global assessment report on biodiversity and ecosystem services of the Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services. *Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services*.
- DIETRICH, P., CESARZ, S., EISENHAUER, N. & ROSCHER, C. 2020. Effects of steam sterilization on soil abiotic and biotic properties. *SOIL ORGANISMS*, 92, 99–108.
- DORNELAS, M., GOTELLI, N. J., MCGILL, B., SHIMADZU, H., MOYES, F., SIEVERS, C. & MAGURRAN, A. E. 2014. Assemblage time series reveal biodiversity change but not systematic loss. *Science*, 344, 296-9.
- EISENHAUER, N. 2010. The action of an animal ecosystem engineer: Identification of the main mechanisms of earthworm impacts on soil microarthropods. *Pedobiologia*, 53, 343-352.
- EISENHAUER, N. 2011. Aboveground–belowground interactions as a source of complementarity effects in biodiversity experiments. *Plant and Soil*, 351, 1-22.
- EISENHAUER, N., DOBIES, T., CESARZ, S., HOBBIE, S. E., MEYER, R. J., WORM, K. & REICH, P. B. 2013. Plant diversity effects on soil food webs are stronger than those of elevated CO₂ and N deposition in a long-term grassland experiment. *Proc Natl Acad Sci U S A*, 110, 6889-94.
- EISENHAUER, N., SABAI, A. C. W. & SCHEU, S. 2011. Collembola species composition and diversity effects on ecosystem functioning vary with plant functional group identity. *Soil Biology and Biochemistry*, 43, 1697-1704.
- EISENHAUER, N. & SCHEU, S. 2008. Earthworms as drivers of the competition between grasses and legumes. *Soil Biology and Biochemistry*, 40, 2650-2659.
- EISENHAUER, N. & TÜRKE, M. 2018. From climate chambers to biodiversity chambers. *Frontiers in Ecology and the Environment*, 16, 136-137.
- FERLIAN, O., BIERE, A., BONFANTE, P., BUSCOT, F., EISENHAUER, N., FERNANDEZ, I., HAUSE, B., HERRMANN, S., KRAJINSKI-BARTH, F., MEIER, I. C., POZO, M. J., RASMANN, S., RILLIG, M. C., TARKKA, M. T., VAN DAM, N. M., WAGG, C. & MARTINEZ-MEDINA, A. 2018. Growing Research Networks on Mycorrhizae for Mutual Benefits. *Trends Plant Sci*, 23, 975-984.

- 671 GESSNER, M. O., SWAN, C. M., DANG, C. K., MCKIE, B. G., BARDGETT, R. D., WALL, D.
672 H. & HATTENSCHWILER, S. 2010. Diversity meets decomposition. *Trends Ecol Evol*,
673 25, 372-80.
- 674 GRIFFIN, K. L., ROSS, P. D., SIMS, D. A., LUO, Y., SEEMANN, J. R., FOX, C. A. & BALL, J.
675 T. 1996. EcoCELLs: tools for mesocosm scale measurements of gas exchange. *Plant*,
676 *Cell & Environment*, 19, 1210-1221.
- 677 GUREVITCH, J., MORROW, L. L., WALLACE, A. & WALSH, J. S. 1992. A Meta-Analysis of
678 Competition in Field Experiments. *The American Naturalist*, 140, 539-572.
- 679 HALPERN, B. S., BERLOW, E., WILLIAMS, R., BORER, E. T., DAVIS, F. W., DOBSON, A.,
680 ENQUIST, B. J., FROEHLICH, H. E., GERBER, L. R., LORTIE, C. J., O'CONNOR, M.
681 I., REGAN, H., VÁZQUEZ, D. P. & WILLARD, G. 2020. Ecological Synthesis and Its
682 Role in Advancing Knowledge. *BioScience*.
- 683 HARPOLE, W. S., SULLIVAN, L. L., LIND, E. M., FIRN, J., ADLER, P. B., BORER, E. T.,
684 CHASE, J., FAY, P. A., HAUTIER, Y., HILLEBRAND, H., MACDOUGALL, A. S.,
685 SEABLOOM, E. W., WILLIAMS, R., BAKKER, J. D., CADOTTE, M. W., CHANETON,
686 E. J., CHU, C., CLELAND, E. E., D'ANTONIO, C., DAVIES, K. F., GRUNER, D. S.,
687 HAGENAH, N., KIRKMAN, K., KNOPS, J. M., LA PIERRE, K. J., MCCULLEY, R. L.,
688 MOORE, J. L., MORGAN, J. W., PROBER, S. M., RISCH, A. C., SCHUETZ, M.,
689 STEVENS, C. J. & WRAGG, P. D. 2016. Addition of multiple limiting resources reduces
690 grassland diversity. *Nature*, 537, 93-96.
- 691 HEINEN, R., BIERE, A., HARVEY, J. A. & BEZEMER, T. M. 2018. Effects of Soil Organisms
692 on Aboveground Plant-Insect Interactions in the Field: Patterns, Mechanisms and the
693 Role of Methodology. *Frontiers in Ecology and Evolution*, 6.
- 694 HILLEBRAND, H., BLASIUS, B., BORER, E. T., CHASE, J. M., DOWNING, J. A., ERIKSSON,
695 B. K., FILSTRUP, C. T., HARPOLE, W. S., HODAPP, D., LARSEN, S.,
696 LEWANDOWSKA, A. M., SEABLOOM, E. W., VAN DE WAAL, D. B., RYABOV, A. B.
697 & CADOTTE, M. 2018. Biodiversity change is uncoupled from species richness trends:
698 Consequences for conservation and monitoring. *Journal of Applied Ecology*, 55, 169-
699 184.
- 700 HILLEBRAND, H., DONOHUE, I., HARPOLE, W. S., HODAPP, D., KUCERA, M.,
701 LEWANDOWSKA, A. M., MERDER, J., MONTOYA, J. M. & FREUND, J. A. 2020.
702 Thresholds for ecological responses to global change do not emerge from empirical
703 data. *Nat Ecol Evol*, 4, 1502-1509.
- 704 HINES, J., EBELING, A., BARNES, A. D., BROSE, U., SCHERBER, C., SCHEU, S.,
705 TSCHARNTKE, T., WEISSER, W. W., GILING, D. P., KLEIN, A. M. & EISENHAUER,
706 N. 2019. Mapping change in biodiversity and ecosystem function research: food webs
707 foster integration of experiments and science policy. In: EISENHAUER, N., BOHAN,
708 D. A. & DUMBRELL, A. J. (eds.) *Mechanisms underlying the relationship between*
709 *biodiversity and ecosystem function*. Academic Press.
- 710 HINES, J., EISENHAUER, N. & DRAKE, B. G. 2015a. Inter-annual changes in detritus-based
711 food chains can enhance plant growth response to elevated atmospheric CO₂. *Glob*
712 *Chang Biol*, 21, 4642-50.
- 713 HINES, J., VAN DER PUTTEN, W. H., DE DEYN, G. B., WAGG, C., VOIGT, W., MULDER,
714 C., WEISSER, W. W., ENGEL, J., MELIAN, C., SCHEU, S., BIRKHOFER, K.,
715 EBELING, A., SCHERBER, C. & EISENHAUER, N. 2015b. Towards an Integration of
716 Biodiversity–Ecosystem Functioning and Food Web Theory to Evaluate Relationships
717 between Multiple Ecosystem Services. In: WOODWARD, G. & BOHAN, D. A. (eds.)
718 *Ecosystem Services - From Biodiversity to Society, Part 1*. Academic Press.
- 719 HOOPER, D. U., BIGNELL, D. E., BROWN, V. K., BRUSSARD, L., DANGERFIELD, J. M.,
720 WALL, D. H., WARDLE, D. A., COLEMAN, D. C., GILLER, K. E., LAVELLE, P., VAN
721 DER PUTTEN, W. H., DE RUITER, P. C., RUSEK, J., SILVER, W. L., TIEDJE, J. M.
722 & WOLTERS, V. 2000. Interactions between Aboveground and Belowground
723 Biodiversity in Terrestrial Ecosystems: Patterns, Mechanisms, and Feedbacks: We
724 assess the evidence for correlation between aboveground and belowground diversity
725 and conclude that a variety of mechanisms could lead to positive, negative, or no

- relationship—depending on the strength and type of interactions among species. *BioScience*, 50, 1049-1061.
- JAGER, G., VAN DER BOON, J. & RAUW, G. 1970. The influence of soil steaming on some properties of the soil and on the growth and heading of winter glasshouse lettuce. III. The influence of nitrogen form, manganese level and shading studied in sand culture experiments with trickle irrigation. *NJAS wageningen journal of life sciences*, 18, 158-167.
- JELTSCH, F., BONTE, D., PE'ER, G., REINEKING, B., LEIMGRUBER, P., BALKENHOL, N., SCHRODER, B., BUCHMANN, C. M., MUELLER, T., BLAUM, N., ZURELL, D., BOHNING-GAESE, K., WIEGAND, T., ECCARD, J. A., HOFER, H., REEG, J., EGGERS, U. & BAUER, S. 2013. Integrating movement ecology with biodiversity research - exploring new avenues to address spatiotemporal biodiversity dynamics. *Mov Ecol*, 1, 6.
- KALINKAT, G., BROSE, U. & RALL, B. C. 2013. Habitat structure alters top-down control in litter communities. *Oecologia*, 172, 877-887.
- KOMATSU, K. J., AVOLIO, M. L., LEMOINE, N. P., ISBELL, F., GRMAN, E., HOUSEMAN, G. R., KOERNER, S. E., JOHNSON, D. S., WILCOX, K. R., ALATALO, J. M., ANDERSON, J. P., AERTS, R., BAER, S. G., BALDWIN, A. H., BATES, J., BEIERKUHNLEIN, C., BELOTE, R. T., BLAIR, J., BLOOR, J. M. G., BOHLEN, P. J., BORK, E. W., BOUGHTON, E. H., BOWMAN, W. D., BRITTON, A. J., CAHILL, J. F., JR., CHANETON, E., CHIARIELLO, N. R., CHENG, J., COLLINS, S. L., CORNELISSEN, J. H. C., DU, G., ESKELINEN, A., FIRN, J., FOSTER, B., GOUGH, L., GROSS, K., HALLETT, L. M., HAN, X., HARMENS, H., HOVENDEN, M. J., JAGERBRAND, A., JENTSCH, A., KERN, C., KLANDERUD, K., KNAPP, A. K., KREYLING, J., LI, W., LUO, Y., MCCULLEY, R. L., MCLAREN, J. R., MEGONIGAL, J. P., MORGAN, J. W., ONIPCHENKO, V., PENNINGS, S. C., PREVEY, J. S., PRICE, J. N., REICH, P. B., ROBINSON, C. H., RUSSELL, F. L., SALA, O. E., SEABLOOM, E. W., SMITH, M. D., SOUDZILOVSKAIA, N. A., SOUZA, L., SUDING, K., SUTTLE, K. B., SVEJCAR, T., TILMAN, D., TOGNETTI, P., TURKINGTON, R., WHITE, S., XU, Z., YAHDJIAN, L., YU, Q., ZHANG, P. & ZHANG, Y. 2019. Global change effects on plant communities are magnified by time and the number of global change factors imposed. *Proc Natl Acad Sci U S A*, 116, 17867-17873.
- KORELL, L., AUGÉ, H., CHASE, J. M., HARPOLE, S. & KNIGHT, T. M. 2020. We need more realistic climate change experiments for understanding ecosystems of the future. *Glob Chang Biol*, 26, 325-327.
- LAWTON, J. H. 1996. The Ecotron Facility at Silwood Park: The Value of "Big Bottle" Experiments. *Ecology*, 77, 665-669.
- LAWTON, J. H., NAEEM, S., WOODFIN, R. M., BROWN, V. K., GANGE, A., GODFRAY, H. J. C., HEADS, P. A., LAWLER, S., MAGDA, D., THOMAS, C. D., THOMPSON, L. J. & YOUNG, S. 1993. The Ecotron: A Controlled Environmental Facility for the Investigation of Population and Ecosystem Processes. *Philosophical Transactions: Biological Sciences*, 341, 181-194.
- LEFCHECK, J. S., BYRNES, J. E. K., ISBELL, F., GAMFELDT, L., GRIFFIN, J. N., EISENHAUER, N., HENSEL, M. J. S., HECTOR, A., CARDINALE, B. J. & DUFFY, J. E. 2015. Biodiversity enhances ecosystem multifunctionality across trophic levels and habitats. *Nat Commun*, 6, 6936.
- LIU, X., ZHANG, L. & HONG, S. 2011. Global biodiversity research during 1900–2009: a bibliometric analysis. *Biodiversity and Conservation*, 20, 807-826.
- MAJDI, N., BOICHÉ, A., TRAUNSPURGER, W. & LECERF, A. 2014. Predator effects on a detritus-based food web are primarily mediated by non-trophic interactions. *Journal of Animal Ecology*, 83, 953-962.
- MATTHEWS, B., NARWANI, A., HAUSCH, S., NONAKA, E., PETER, H., YAMAMICHI, M., SULLAM, K. E., BIRD, K. C., THOMAS, M. K., HANLEY, T. C. & TURNER, C. B. 2011. Toward an integration of evolutionary biology and ecosystem science. *Ecol Lett*, 14, 690-701.

- MAXWELL, S. L., FULLER, R. A., BROOKS, T. M. & WATSON, J. E. 2016. Biodiversity: The ravages of guns, nets and bulldozers. *Nature*, 536, 143-5.
- MILCU, A., PUGA-FREITAS, R., ELLISON, A. M., BLOUIN, M., SCHEU, S., FRESCHET, G. T., ROSE, L., BAROT, S., CESARZ, S., EISENHAEUER, N., GIRIN, T., ASSANDRI, D., BONKOWSKI, M., BUCHMANN, N., BUTENSCHOEN, O., DEVIDAL, S., GLEIXNER, G., GESSLER, A., GIGON, A., GREINER, A., GRIGNANI, C., HANSART, A., KAYLER, Z., LANGE, M., LATA, J. C., LE GALLIARD, J. F., LUKAC, M., MANNERHEIM, N., MULLER, M. E. H., PANDO, A., ROTTER, P., SCHERER-LORENZEN, M., SEYHUN, R., URBAN-MEAD, K., WEIGELT, A., ZAVATTARO, L. & ROY, J. 2018. Genotypic variability enhances the reproducibility of an ecological study. *Nat Ecol Evol*, 2, 279-287.
- MILCU, A., ROSCHER, C., GESSLER, A., BACHMANN, D., GOCKELE, A., GUDERLE, M., LANDAIS, D., PIEL, C., ESCAPE, C., DEVIDAL, S., RAVEL, O., BUCHMANN, N., GLEIXNER, G., HILDEBRANDT, A. & ROY, J. 2014. Functional diversity of leaf nitrogen concentrations drives grassland carbon fluxes. *Ecol Lett*, 17, 435-44.
- MÖLLER, B., CHEN, H., SCHMIDT, T., ZIESCHANK, A., PATZAK, R., TÜRKE, M., WEIGELT, A. & POSCH, S. 2019. rhizoTrak: a flexible open source Fiji plugin for user-friendly manual annotation of time-series images from minirhizotrons. *Plant and Soil*, 444, 519-534.
- MURPHY, G. E. & ROMANUK, T. N. 2014. A meta-analysis of declines in local species richness from human disturbances. *Ecol Evol*, 4, 91-103.
- NAEEM, S., THOMPSON, L. J., LAWLER, S. P., LAWTON, J. H. & WOODFIN, R. M. 1994. Declining biodiversity can alter the performance of ecosystems. *Nature*, 368, 734-737.
- NEWBOLD, T., HUDSON, L. N., HILL, S. L., CONTU, S., LYSENKO, I., SENIOR, R. A., BORGER, L., BENNETT, D. J., CHOIMES, A., COLLEN, B., DAY, J., DE PALMA, A., DIAZ, S., ECHEVERRIA-LONDONO, S., EDGAR, M. J., FELDMAN, A., GARON, M., HARRISON, M. L., ALHUSSEINI, T., INGRAM, D. J., ITESCU, Y., KATTGE, J., KEMP, V., KIRKPATRICK, L., KLEYER, M., CORREIA, D. L., MARTIN, C. D., MEIRI, S., NOVOSOLOV, M., PAN, Y., PHILLIPS, H. R., PURVES, D. W., ROBINSON, A., SIMPSON, J., TUCK, S. L., WEIHER, E., WHITE, H. J., EWERS, R. M., MACE, G. M., SCHARLEMANN, J. P. & PURVIS, A. 2015. Global effects of land use on local terrestrial biodiversity. *Nature*, 520, 45-50.
- PEREIRA, H. M., NAVARRO, L. M. & MARTINS, I. S. 2012. Global Biodiversity Change: The Bad, the Good, and the Unknown. *Annual Review of Environment and Resources*, 37, 25-50.
- PIERIK, R., BALLARE, C. L. & DICKE, M. 2014. Ecology of plant volatiles: taking a plant community perspective. *Plant Cell Environ*, 37, 1845-53.
- RILLIG, M. C., RYO, M., LEHMANN, A., AGUILAR-TRIGUEROS, C. A., BUCHERT, S., WULF, A., IWASAKI, A., ROY, J. & YANG, G. 2019. The role of multiple global change factors in driving soil functions and microbial biodiversity. *Science*, 366, 886-890.
- RINEAU, F., MALINA, R., BEENAERTS, N., ARNAUTS, N., BARDGETT, R. D., BERG, M. P., BOEREMA, A., BRUCKERS, L., CLERINX, J., DAVIN, E. L., DE BOECK, H. J., DE DOBBELAER, T., DONDINI, M., DE LAENDER, F., ELLERS, J., FRANKEN, O., GILBERT, L., GUDMUNDSSON, L., JANSSENS, I. A., JOHNSON, D., LIZIN, S., LONGDOZ, B., MEIRE, P., MEREMANS, D., MILBAU, A., MORETTI, M., NIJS, I., NOBEL, A., POP, I. S., PUETZ, T., REYNS, W., ROY, J., SCHUETZ, J., SENEVIRATNE, S. I., SMITH, P., SOLMI, F., STAES, J., THIERY, W., THIJS, S., VANDERKELEN, I., VAN LANDUYT, W., VERBRUGGEN, E., WITTERS, N., ZSCHEISCHLER, J. & VANGRONSVELD, J. 2019. Towards more predictive and interdisciplinary climate change ecosystem experiments. *Nature Climate Change*, 9, 809-816.
- ROMERO, G. Q. & KORICHEVA, J. 2011. Contrasting cascade effects of carnivores on plant fitness: a meta-analysis. *Journal of Animal Ecology*, 80, 696-704.
- ROSCHER, C., SCHUMACHER, J., BAADE, J., WILCKE, W., GLEIXNER, G., WEISSER, W. W., SCHMID, B. & SCHULZE, E.-D. 2004. The role of biodiversity for element cycling

- and trophic interactions: an experimental approach in a grassland community. *Basic and Applied Ecology*, 5, 107-121.
- ROY, J., RINEAU, F., DE BOECK, H. J., NIJS, I., PUTZ, T., ABIVEN, S., ARNONE, J. A., 3RD, BARTON, C. V. M., BEENAERTS, N., BRUGGEMANN, N., DAINESE, M., DOMISCH, T., EISENHAUER, N., GARRE, S., GEBLER, A., GHIRARDO, A., JASONI, R. L., KOWALCHUK, G., LANDAIS, D., LARSEN, S. H., LEEMANS, V., LE GALLIARD, J. F., LONGDOZ, B., MASSOL, F., MIKKELSEN, T. N., NIEDRIST, G., PIEL, C., RAVEL, O., SAUZE, J., SCHMIDT, A., SCHNITZLER, J. P., TEIXEIRA, L. H., TJOELKER, M. G., WEISSER, W. W., WINKLER, B. & MILCU, A. 2020. Ecotrons: Powerful and versatile ecosystem analysers for ecology, agronomy and environmental science. *Glob Chang Biol*, n/a.
- ROY, J., TARDIEU, F., TIXIER-BOICHARD, M. & SCHURR, U. 2017. European infrastructures for sustainable agriculture. *Nat Plants*, 3, 756-758.
- SABAIS, A. C., EISENHAUER, N., KÖNIG, S., RENKER, C., BUSCOT, F. & SCHEU, S. 2012. Soil organisms shape the competition between grassland plant species. *Oecologia*, 170, 1021-1032.
- SCHÄDLER, M., BUSCOT, F., KLOTZ, S., REITZ, T., DURKA, W., BUMBERGER, J., MERBACH, I., MICHALSKI, S. G., KIRSCH, K., REMMLER, P., SCHULZ, E. & AUGÉ, H. 2019. Investigating the consequences of climate change under different land-use regimes: a novel experimental infrastructure. *Ecosphere*, 10, e02635.
- SCHERBER, C., EISENHAUER, N., WEISSER, W. W., SCHMID, B., VOIGT, W., FISCHER, M., SCHULZE, E. D., ROSCHER, C., WEIGELT, A., ALLAN, E., BESSLER, H., BONKOWSKI, M., BUCHMANN, N., BUSCOT, F., CLEMENT, L. W., EBELING, A., ENGELS, C., HALLE, S., KERTSCHER, I., KLEIN, A. M., KOLLER, R., KONIG, S., KOWALSKI, E., KUMMER, V., KUU, A., LANGE, M., LAUTERBACH, D., MIDDELHOFF, C., MIGUNOVA, V. D., MILCU, A., MULLER, R., PARTSCH, S., PETERMANN, J. S., RENKER, C., ROTTSTOCK, T., SABAIS, A., SCHEU, S., SCHUMACHER, J., TEMPERTON, V. M. & TSCHARNTKE, T. 2010. Bottom-up effects of plant diversity on multitrophic interactions in a biodiversity experiment. *Nature*, 468, 553-6.
- SCHEU, S. 2001. Plants and generalist predators as links between the below-ground and above-ground system. *Basic and Applied Ecology*, 2, 3-13.
- SCHEU, S. 2003. Effects of earthworms on plant growth: patterns and perspectives. *Pedobiologia*, 47, 846-856.
- SCHINDLER, D. W. 1998. Whole-Ecosystem Experiments: Replication Versus Realism: The Need for Ecosystem-Scale Experiments. *Ecosystems*, 1, 323-334.
- SCHMIDT, A., AUGÉ, H., BRANDL, R., HEONG, K. L., HOTES, S., SETTELE, J., VILLAREAL, S. & SCHÄDLER, M. 2015a. Small-scale variability in the contribution of invertebrates to litter decomposition in tropical rice fields. *Basic and Applied Ecology*, 16, 674-680.
- SCHMIDT, A., JOHN, K., ARIDA, G., AUGÉ, H., BRANDL, R., HORGAN, F. G., HOTES, S., MARQUEZ, L., RADERMACHER, N., SETTELE, J., WOLTERS, V. & SCHÄDLER, M. 2015b. Effects of Residue Management on Decomposition in Irrigated Rice Fields Are Not Related to Changes in the Decomposer Community. *PLoS ONE*, 10, e0134402.
- SCHWARZMULLER, F., EISENHAUER, N. & BROSE, U. 2015. 'Trophic whales' as biotic buffers: weak interactions stabilize ecosystems against nutrient enrichment. *J Anim Ecol*, 84, 680-691.
- SEABLOOM, E. W., ADLER, P. B., ALBERTI, J., BIEDERMAN, L., BUCKLEY, Y. M., CADOTTE, M. W., COLLINS, S. L., DEE, L., FAY, P. A., FIRN, J., HAGENAH, N., HARPOLE, W. S., HAUTIER, Y., HECTOR, A., HOBBIE, S. E., ISBELL, F., KNOPS, J. M. H., KOMATSU, K. J., LAUNGANI, R., MACDOUGALL, A., MCCULLEY, R. L., MOORE, J. L., MORGAN, J. W., OHLERT, T., PROBER, S. M., RISCH, A. C., SCHUETZ, M., STEVENS, C. J. & BORER, E. T. 2021. Increasing effects of chronic nutrient enrichment on plant diversity loss and ecosystem productivity over time. *Ecology*, 102, e03218.

- SOLIVERES, S., VAN DER PLAS, F., MANNING, P., PRATI, D., GOSSNER, M. M.,
 RENNER, S. C., ALT, F., ARNDT, H., BAUMGARTNER, V., BINKENSTEIN, J.,
 BIRKHOFFER, K., BLASER, S., BLUTHGEN, N., BOCH, S., BOHM, S., BORSCHIG,
 C., BUSCOT, F., DIEKOTTER, T., HEINZE, J., HOLZEL, N., JUNG, K., KLAUS, V. H.,
 KLEINEBECKER, T., KLEMMER, S., KRAUSS, J., LANGE, M., MORRIS, E. K.,
 MULLER, J., OELMANN, Y., OVERMANN, J., PASALIC, E., RILLIG, M. C.,
 SCHAEFER, H. M., SCHLOTER, M., SCHMITT, B., SCHONING, I., SCHRUMPF, M.,
 SIKORSKI, J., SOCHER, S. A., SOLLY, E. F., SONNEMANN, I., SORKAU, E.,
 STECKEL, J., STEFFAN-DEWENTER, I., STEMPFHUBER, B., TSCHAPKA, M.,
 TURKE, M., VENTER, P. C., WEINER, C. N., WEISSER, W. W., WERNER, M.,
 WESTPHAL, C., WILCKE, W., WOLTERS, V., WUBET, T., WURST, S., FISCHER, M.
 & ALLAN, E. 2016. Biodiversity at multiple trophic levels is needed for ecosystem
 multifunctionality. *Nature*, 536, 456-9.
- SPAACK, J. W., BAERT, J. M., BAIRD, D. J., EISENHAEUER, N., MALTBY, L., POMATI, F.,
 RADCHUK, V., ROHR, J. R., VAN DEN BRINK, P. J. & DE LAENDER, F. 2017. Shifts
 of community composition and population density substantially affect ecosystem
 function despite invariant richness. *Ecology Letters*, 20, 1315-1324.
- STEWART, R. I., DOSSENA, M., BOHAN, D. A., JEPPESEN, E., KORDAS, R. L., LEDGER,
 M. E., MEERHOFF, M., MOSS, B., MULDER, C. & SHURIN, J. B. 2013. Mesocosm
 experiments as a tool for ecological climate-change research. *Advances in ecological
 research*. Elsevier.
- STORK, H. & ASTRIN, J. J. 2014. Trends in Biodiversity Research — A Bibliometric
 Assessment. *Open Journal of Ecology*, 04, 354-370.
- THAKUR, M. P. 2020. Climate warming and trophic mismatches in terrestrial ecosystems: the
 green–brown imbalance hypothesis. *Biology Letters*, 16, 20190770.
- TILMAN, D. 1982. *Resource competition and community structure*, Princeton university press.
- TILMAN, D. & DOWNING, J. A. 1994. Biodiversity and stability in grasslands. *Nature*, 367,
 363-365.
- TREVORS, J. T. 1996. Sterilization and inhibition of microbial activity in soil. *Journal of
 Microbiological Methods*, 26, 53-59.
- TURCOTTE, M. M. & LEVINE, J. M. 2016. Phenotypic Plasticity and Species Coexistence.
Trends Ecol Evol, 31, 803-813.
- ULRICH, J., BUCHER, S. F., EISENHAEUER, N., SCHMIDT, A., TURKE, M., GEBLER, A.,
 BARRY, K., LANGE, M. & ROMERMANN, C. 2020. Invertebrate Decline Leads to
 Shifts in Plant Species Abundance and Phenology. *Front Plant Sci*, 11, 542125.
- VAN GROENIGEN, J. W., LUBBERS, I. M., VOS, H. M., BROWN, G. G., DE DEYN, G. B. &
 VAN GROENIGEN, K. J. 2014. Earthworms increase plant production: a meta-
 analysis. *Sci Rep*, 4, 6365.
- VANDERKELEN, I., ZSCHEISCHLER, J., GUDMUNDSSON, L., KEULER, K., RINEAU, F.,
 BEENAERTS, N., VANGRONSVELD, J., VICCA, S. & THIERY, W. 2020. A novel
 method for assessing climate change impacts in ecotron experiments. *Int J
 Biometeorol*, 64, 1709-1727.
- VEEN, G. F., WUBS, E. R. J., BARDGETT, R. D., BARRIOS, E., BRADFORD, M. A.,
 CARVALHO, S., DE DEYN, G. B., DE VRIES, F. T., GILLER, K. E., KLEIJN, D.,
 LANDIS, D. A., ROSSING, W. A. H., SCHRAMA, M., SIX, J., STRUIK, P. C., VAN
 GILS, S., WISKERKE, J. S. C., VAN DER PUTTEN, W. H. & VET, L. E. M. 2019.
 Applying the Aboveground-Belowground Interaction Concept in Agriculture: Spatio-
 Temporal Scales Matter. *Frontiers in Ecology and Evolution*, 7.
- VELLEND, M., BAETEN, L., MYERS-SMITH, I. H., ELMENDORF, S. C., BEAUSEJOUR, R.,
 BROWN, C. D., DE FRENNE, P., VERHEYEN, K. & WIPF, S. 2013. Global meta-
 analysis reveals no net change in local-scale plant biodiversity over time. *Proc Natl
 Acad Sci U S A*, 110, 19456-9.
- VOIGT, W., PERNER, J., DAVIS, A. J., EGGERS, T., SCHUMACHER, J., BÄHRMANN, R.,
 FABIAN, B., HEINRICH, W., KÖHLER, G., LICHTER, D., MARSTALLER, R. &

945 SANDER, F. W. 2003. Trophic Levels Are Differentially Sensitive to Climate. *Ecology*,
 946 84, 2444-2453.
 947 WARDLE, D. A., BARDGETT, R. D., KLIRONOMOS, J. N., SETALA, H., VAN DER PUTTEN,
 948 W. H. & WALL, D. H. 2004. Ecological linkages between aboveground and
 949 belowground biota. *Science*, 304, 1629-33.
 950 WARDLE, D. A., WILLIAMSON, W. M., YEATES, G. W. & BONNER, K. I. 2005. Trickle-down
 951 effects of aboveground trophic cascades on the soil food web. *Oikos*, 111, 348-358.
 952 WILSON, M. W., RIDLON, A. D., GAYNOR, K. M., GAINES, S. D., STIER, A. C. & HALPERN,
 953 B. S. 2020. Ecological impacts of human-induced animal behaviour change. *Ecol Lett*,
 954 23, 1522-1536.

955

956

958 **Table App.1:** Abiotic parameters of an EcoUnit, control and data storage options; combined temperature/humidity sensors: *MELA FE09, Galltec*
 959 *Meß- und Regeltechnik GmbH Bondorf, Germany*; integrated flow meter: *FCH-midi-POM 97478976, B.I.O.-TECH e.K. Vilshofen, Germany*;
 960 temperature/moisture sensors belowground: *SMT100, TRUEBNER GmbH Neustadt, Germany*; observation cameras: *YUC-Hi82M, Yudor*
 961 *Technology Co, Ltd Tao Yuan City 324, Taiwan.*

Parameter	Controlling	User interface	Sensing
air temperature	via adjustable ambient temperature of hall	GUI	4 combined temperature/humidity sensors
air humidity	indirect only by air temperature and air flow rate	-	4 combined temperature/humidity sensors
air flow rate	by blower speed	GUI	manually by air velocimeter
lighting timing	1 h setting resolution with automatically calculated intermediate dim steps for each channel	GUI	logging of execution confirmation only
light intensity	nominal 1% setting resolution with internal mapping to nearest dim step	GUI	logging of execution confirmation only
light color mix	4 dim channels (UV, blue, red, NIR); 1 non dim channel (FIR)	GUI	logging of execution confirmation only
irrigation volume	50 ml setting resolution	GUI	integrated flow meter
irrigation timing	1 h setting resolution	GUI	logging of execution confirmation only

soil temperature	by cooling at the bottom down to ~10 °C with resulting temperature gradient to soil surface	GUI	up to 12 combined temperature/moisture sensors in three levels belowground
soil moisture	indirect only by change of irrigation volume, soil water removal, and the manipulation of evaporation rate by air flow rate	-	up to 12 combined temperature/moisture sensors in three levels belowground
suction low pressure	1 kPa setting resolution with low pressure down to -60 kPa below ambient air pressure	GUI	each suction system includes an integrated pressure sensor
video observation	orientation of vision and operation mode manually only	camera's web GUI	observation camera
still pictures	by external script with access to video stream of running cams	camera's web GUI + Linux shell	observation camera

962

963

964

965 **Table App.5** Species list of the commercial grassland litter from bunny® Frischgras-Heu
 966 provided as substrate to both lysimeters.

English name	Latin name
Timothy (grass)	<i>Phleum pratense</i>
Meadow fescue	<i>Festuca pratensis</i>
Meadow foxtail	<i>Alopecurus pratensis</i>
Ryegrass	<i>Lolium sp.</i>
Red fescue	<i>Festuca rubra</i> agg.
Kentucky bluegrass	<i>Poa pratensis</i>
Bent grass	<i>Agrostis sp.</i>
Cat grass	<i>Dactylis glomerata</i>
Common dandelion	<i>Taraxacum officinale</i>
Common silverweed	<i>Potentilla anserina</i>
Mouse-ear chickweed	<i>Cerastium sp.</i>
Yarrows	<i>Achillea sp.</i>
Ribwort plantain	<i>Plantago lanceolata</i>
White clover	<i>Trifolium repens</i>
Red clover	<i>Trifolium pratense</i>
Common bird's-foot trefoil	<i>Lotus corniculatus</i>

967

968

969 **Table App.8** Effects of aboveground (AG) and belowground (BG) invertebrates and bean plant neighbor species identity as well as their interactions on five response variables using GLMM type III sum of squares
970 analyses; significant effects (P < 0.05) are indicated in bold font; *sp. richness*: species richness, *max. no. of aphids*: maximum number of aphids.

Grouping variable	Response variable														
	Bean dry weight			Nematode density			Nematode sp. richness			Max. no. of aphids			Days aphid infestation		
	df	F	P	df	F	P	df	F	P	df	F	P	df	F	P
AG invertebrates	3, 48	6.48	< 0.001	3, 48	0.16	0.92	3, 48	0.51	0.67	1, 24	8.24	0.01	3, 48	63.19	< 0.001
BG invertebrates	1, 48	0.26	0.61	1, 48	1.01	0.32	1, 48	1.99	0.16	1, 24	0.22	0.64	1, 48	0.20	0.65
AG invertebrates*BG invertebrates	3, 48	1.37	0.26	3, 48	0.84	0.48	3, 48	0.12	0.95	1, 24	0.25	0.62	3, 48	1.25	0.30
Neighbor species	3, 48	5.16	< 0.01	3, 48	2.86	0.05	3, 48	1.62	0.20	3, 24	1.42	0.26	3, 48	0.36	0.78
Neighbor species*AG invertebrates	9, 48	0.22	0.99	9, 48	1.40	0.22	9, 48	2.21	0.04	3, 24	0.44	0.73	9, 48	0.87	0.56
Neighbor species*BG invertebrates	3, 48	0.30	0.82	3, 48	1.25	0.30	3, 48	0.62	0.61	3, 24	5.82	< 0.01	3, 48	1.58	0.21
Neighbor species*AG invertebrates*BG invertebrates	9, 48	0.92	0.52	9, 48	0.76	0.65	9, 48	0.86	0.56	3, 24	0.24	0.87	9, 48	0.30	0.97

971

972

Figure captions

Figure 1 Illustration of an EcoUnit; (A) construction drawing with corner cutout to visualize the technical interior features; (B) EcoUnit with earth-filled lower part, upper part equipped with illustrative vegetation.

Figure 2 Grassland (upper picture) and tree saplings (bottom picture) planted in EcoUnits of the iDiv Ecotron.

Figure 3 Experimental setup of the case study.

Figure 4 Effects of (a) aboveground invertebrate treatment (Control, Herbivores only, Coccinella low, Coccinella high; details in App.3) as well as (b) the interaction of belowground invertebrate presence (with/+ soil fauna, without/- soil fauna) and bean plant neighbor species identity (*Bellis perennis* L., *Centaurea jacea* L., *Festuca pratensis* Huds., *Holcus lanatus* L.) on the maximum number of aphids; (c) aboveground invertebrate treatment and (d) bean plant neighbor species identity on bean dry weight; and (e) the interaction of aboveground invertebrate treatment and bean plant neighbor species identity on the species richness of nematodes. *p = 0.05; **p < 0.01; ***p < 0.001. For detailed results see App.8.

Figure App.2: (A) Total light intensity ($\mu\text{mol}/\text{m}^2\text{s}$), 5 cm above the standard soil surface, at 36 locations within an EcoUnit, averaged over 24 EcoUnits, and (B) normalized deviations of light intensity at the 36 spots within an EcoUnit; normalization based on the highest measured average value of total light intensity (shown in (A)), highest value is set to 1. Five cm above the standard soil surface. Distance luminaire to backlighting layer 160 mm; outside corners are not included in calculations due to high edge effects inevitably created by the construction itself.

1000 Data Availability Statement

1001 The datasets generated for this study will be published through the iDiv Data Repository upon
1002 acceptance of this manuscript. The respective DOI will be provided.

1003

1004 Competing Interest Statement

1005 The authors declare no competing interests.

1006

1007 Acknowledgements

1008 Numerous scientific and technical staff from iDiv and UFZ were involved in the
1009 conceptualization of the iDiv Ecotron. Special thanks to the participants of the first sDiv
1010 Ecotron workshop in 2013, to all iDiv scientists who provided valuable input for the Ecotron
1011 concept by participating in surveys, and to the other members of the Ecotron Board Ulrich
1012 Brose, Stanley Harpole, and Reinart Feldmann. All authors acknowledge funding by the
1013 German Centre for Integrative Biodiversity Research (iDiv) Halle-Jena-Leipzig funded by the
1014 German Research Foundation (DFG– FZT 118, 202548816). We thank Oliver Bednorz,
1015 Konrad Kirsch, Ines Merbach, and all other employees of the Bad Lauchstädt Experimental
1016 Research Station of the Helmholtz Centre for Environmental Research – UFZ for support on-
1017 site. Finally, we acknowledge the members of the Experimental Interaction Ecology group and
1018 all student helpers for their support during field and lab work.

1019